



Markham Sub-Committee Meeting

Agenda

June 18, 2025, 9:00 AM - 11:00 AM

Council Chamber

Sub-Committee Members:

Regional Councillor Alan Ho, Chair

Councillor Karen Rea

Councillor Andrew Keyes

Councillor Amanda Collucci

Mayor Frank Scarpitti (Ex-Officio)

Deputy Mayor Michael Chan (Ex-Officio)

Regional Councillor Jim Jones (Ex-Officio)

Councillor Joe Li (Ex-Officio)

Pages

1. CALL TO ORDER

2. DISCLOSURE OF PECUNIARY INTEREST

3. REPORTS AND PRESENTATIONS

3.1 SWAN LAKE PARK DESIGNATION

3

1. That the presentation titled "Agenda Item Requests from Friends of Swan Lake Park" be received for information purposes only; and further,
2. That Staff be authorized and directed to do all things necessary to give effect to this resolution.

3.2 SWAN LAKE PARK ANNUAL REPORTS

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1. That the report entitled "Swan Lake - 2024 Water Quality Status and Updates" be received; and,
2. That the consolidated report entitled "Swan Lake Flow Diversion Assessment" be received; and,
3. That the presentation titled "Swan Lake - Annual Meeting with Markham Subcommittee" be received; and,

4. That Staff continue to implement the Long-term Management Plan for Swan Lake approved by Council in December 2021, including advancements previously made from Phases 2 and 3 of the Plan; and,
5. That Staff report back annually on water quality results and evaluation of adapted Core and Complementary measures for consideration in Phase 2 of the Plan through the Markham Sub-Committee with the participation of the Friends of Swan Lake Park; and,
6. That Staff consider findings and evaluations of chloride diversion options in Phase 3 of the Plan if required given future chloride levels in the Lake; and,
7. That the Plan review be initiated in 2025 with consideration for a workshop to review external feedback; and further,
8. That Staff be authorized and directed to do all things necessary to give effect to this resolution.

4. ADJOURNMENT



Agenda Item Requests from Friends of Swan Lake Park

Markham Sub-Committee

June 18, 2025

Agenda

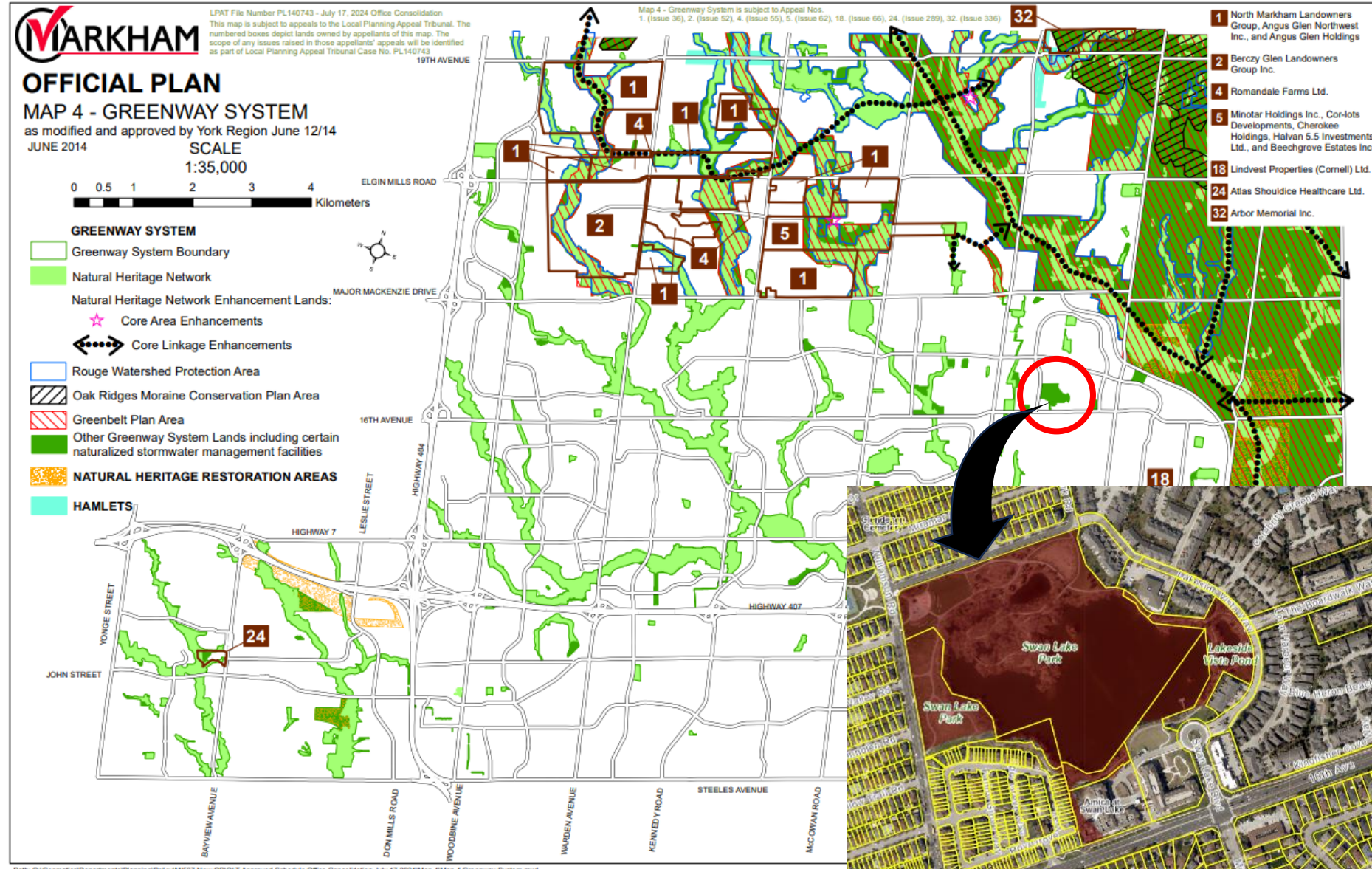
FOSLP has requested that “a review of the classification of Swan Lake Park” be added as a separate item on the agenda for the upcoming Markham Subcommittee

The FOSLP is requesting:

1. staff confirm whether the classification and identification of Swan Lake and Swan Lake Park as “Natural Heritage Network” in the Greenway System is being considered in the Official Plan Review workplan
2. staff be asked to outline the processes involved for reviewing the park classification for Swan Lake and Swan Lake Park (to classify Swan Lake as a “Destination Park”)

Swan Lake Park

- Currently identified as “Other Greenway System Lands”
- Identification based on 2014 Official Plan information
- Official Plan Review will apply criteria to determine what features, including the lake, meet requirements for “Natural Heritage Network” identification



What is “Natural Heritage Network”?

- Natural Heritage Network is a sub-category of the City’s Greenway System which identify all of the City’s natural heritage features such as wetlands, woodlands and valleylands
 - Identification of lands in the Natural Heritage Network is based on policy criteria and technical studies completed by the City and our consultants which assess whether natural heritage features exist on the landscape
 - Majority of Swan Lake Park is a man-made waterbody while some of the shoreline areas to the north of Swan Lake include woodlands and wetland features
- Natural Heritage Network is a land use and planning designation. It allows the City to identify requirements for planning and development including requirements for buffers and environmental impact studies prior to development on adjacent lands.
 - Staff note that Swan Lake Park and surrounding lands are fully built-out.
 - Identifying lands as ‘Natural Heritage Network’ does not influence the City’s level of service for park/asset management

Identifying Swan Lake as “Natural Heritage Network” in the Greenway System

Staff Response:

- a. In 2022, staff advised FOSLP that their request would be considered through the future Official Plan Review. In Jan 2025, Council held a Special Meeting to start the Official Plan Review process. As of June 2025, staff are in the process of preparing an RFP to retain consultants to undertake the Official Plan Review.
- b. Staff will confirm through the OPR whether the required criteria for identifying Swan Lake as “Natural Heritage Network” are met.
- c. FOSLP are also in the process of updating their original 2022 report and staff will review the updated information, once received, as part of the review work for the Official Plan Review.

Swan Lake Park Classification

Staff Response:

- Staff will review park classification criteria outlined in the Official Plan through Official Plan Review (OPR) project. Staff anticipates that current classification of Swan Lake Park would be reviewed through the upcoming OPR. Policy Planning is leading the OPR and as of June 2025, staff are in the process of preparing an RFP to retain consultants to undertake the Review.
- Swan Lake Park was classified as a 'Community Park' in the past, however the Swan Lake Park doesn't meet the criteria of a 'Community Park' according to the park classification criteria outlined in the 2014 Official Plan.

Current Park Classification - 'Community Park'; refer to Policy 4.3.2.2 b) ii) of the 2014 Official Plan:

'Community Parks': generally in excess of 6 hectares, that provide programs and facilities for a number of communities, neighbourhoods and areas and include water play, playgrounds, skateparks, basketball and tennis courts and organized sporting activities for all age groups and supporting infrastructure such as large park pavilions and maintenance facilities. Community Parks are intended to serve park users generally within a 10 minute walking distance (approximately 800 metres)



What is a "Destination Park"?

Policy 4.3.2.2 a) of the 2014 Official Plan:

"Destination Parks, including large and unique parks which attract residents from across Markham and the Region and include conservation areas and lands associated with the Rouge Park that are intended to serve broader regional, provincial, and in some instances, national interests.

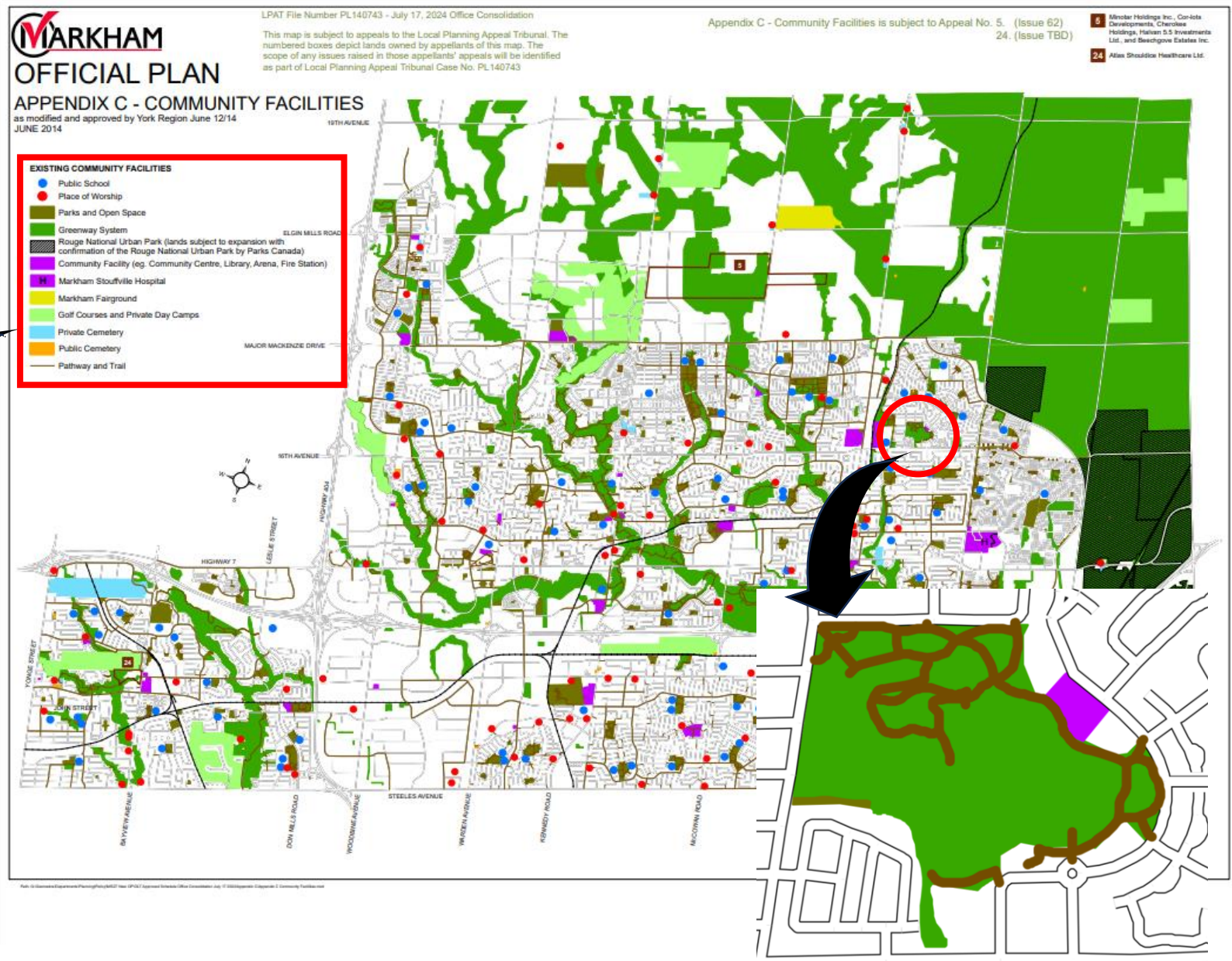
In general, these parks perform an important environmental function. While Destination Parks provide additional uses and opportunities not typically provided by City Parks, they cannot compensate for the parks and open space facilities and services that would otherwise be required under the Planning Act as City Parks."



Planning and Urban Design

Swan Lake Park

- Identified as “Greenway System - Other Greenway System” in the ‘**Appendix C - Community Facilities**’ of 2014 Official Plan. The Swan Lake Property/ Parcel(s) include Pathways and Trail.





Swan Lake Water Quality Monitoring 2024 Annual Report

February 2025

Project Number: 24254

**Swan Lake Monitoring Program
2024 Annual Report**

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List of Acronyms

DO	Dissolved Oxygen
DOC	Dissolved Organic Carbon
GIS	Geographic Information System
LED	Light Emitting Diode
MASL	Meter Above Sea Level
MDL	Method Detection Limit
OGS	Oil and Grit Separator
OMNRF	Ontario Ministry of Natural Resources and Forestry
PAC	Poly Aluminum Chloride
SAV	Submerged Aquatic Vegetation
TKN	Total Kjeldahl Nitrogen
TP	Total Phosphorus
TRCA	Toronto and Region Conservation Authority
VAC	(Swan Lake) Village Amenities Committee
WHO	World Health Organization

Executive Summary

Background

Swan Lake is situated in the City of Markham at the intersection of Sixteenth Avenue and Williamson Road. Swan Lake has an approximate area of 5.5 ha and a maximum water depth of 4.5 m (from the edge of the Lake at 210 meter above sea level). A gravel pit in the 1960s and 1970s, Swan Lake is currently a community feature with multiple trails and urban development surrounding it.

Several issues were discovered with Swan Lake in 2010, including high phosphorus levels and significant algal blooms during the summer months, which led to low oxygen levels and degraded fish habitats. A Phoslock treatment was administered in 2013 to reduce the phosphorus levels and algal blooms in Swan Lake.

In 2019, the City of Markham conducted a study to define a Water Quality Management Strategy for Swan Lake. The Strategy, finalized in July 2020, recommended a chemical treatment in 2021. In August 2021, 13 tonnes of Poly Aluminum Chloride (PAC) were applied to the Lake in a controlled manner over several days.

The Swan Lake Long-Term Management Plan, which was developed based on the 2019 Strategy and extensive consultation with stakeholders, was received by the Markham Sub Committee in November 2021 and approved by the Council in December 2021. It describes a phased adaptive approach, including Core, Complementary and Alternative measures, and periodic reviews to adapt the Plan to the Lake conditions.

In 2024, all Core measures were implemented as planned, including a second application of PAC based on the treatment plan developed by our consultant, AECOM. About nine tonnes of PAC was applied over two application events in late June, with each application event separated by one or two days of downtime to allow for floc formation and environmental testing.

Additional submerged aquatic vegetation was planted in the Lake following PAC application. It is expected that the relative water clarity would help establish the plants, which in turn will improve water clarity further. A geese management program, and a fish inventory and the removal of bottom-dwelling fish were completed in 2024 similar to previous years.

A Flow Diversion feasibility study and a chloride treatment pilot project continued in 2024.

Water quality monitoring of Swan Lake has been conducted almost annually since the first treatment in 2013 to track water quality and the continued effectiveness of the treatment. The collected data presented in this report is part of the ongoing monitoring program that will allow for continuous assessment of the water quality in Swan Lake and will be used to implement and adapt the Long-Term Management Plan for Swan Lake.

This report discusses observations at the monitored stations in the Lake throughout 2024.

Results- Lake Water Quality

Water quality is regularly monitored at two shoreline sites: the Dock and the Bridge, on a bi-weekly basis (from April to November). Samples and measurements are taken at 0.5 m or 1 m increments for the depth of the lake. A level logger is used to record the water level in the Lake. A Dissolved Oxygen (DO) logger was also installed 1 m from the Lake bottom to record the diurnal changes in DO.

Swan Lake Monitoring Program 2024 Annual Report

Trent University collected samples and launched loggers in Swan Lake in the summer of 2024 to support a study on the environmental fates of lanthanum from La-modified bentonite in the ecosystem of Swan Lake. Data provided by Trent researchers have been incorporated in this report.

The following paragraphs provide the monitoring results for the 2024 monitoring period, as well as annual summaries of available data from 2016 to 2024. The figures include plots of measured DO, water clarity, phosphorus concentration, chloride concentration, and geese count.

Targets

Phosphorus concentration and clarity were compared to the eutrophication thresholds and/or the interim targets developed for Swan Lake through the 2019 Water Quality Management Strategy. For DO and chloride, Federal and/or Provincial water quality Guidelines or Objectives are shown for perspective. It should be noted that Swan Lake is not a natural waterbody, and there is no requirement for it to comply with these guidelines and objectives. Where technically and economically feasible, the City will aim to meet these guidelines and objectives to protect and enhance the aquatic environment.

Dissolved Oxygen (DO), Temperature, and pH

The minimum dissolved oxygen concentration required for the protection of warm water fish is 5 mg/L for water temperatures up to 20 °C, and 4 mg/L for temperatures above 20 °C. DO concentrations for the 1m from the surface and 1m from the bottom layers are shown below.

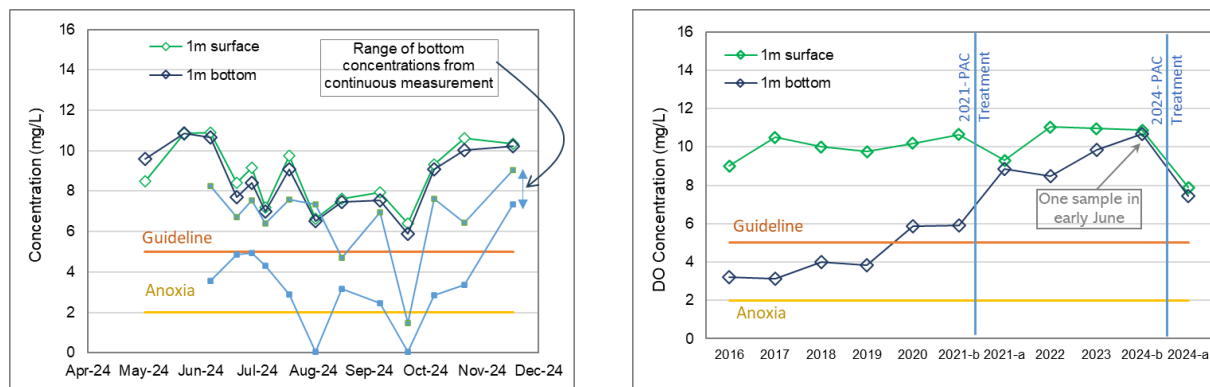
Day-time concentrations measured during biweekly visits were above the DO guideline (above 6.4 mg/L at the surface and above 5.9 mg/L at 1 m from the bottom).

While measured daytime DO levels did not indicate anoxia during the sampling events, continuous measurements at the Dock as well as at the deepest point in the lake (i.e., monitored at a logger commissioned by Trent University) showed a decline in bottom water DO about 50% of the time during August and September, when a dry and warm period followed a wet summer. These declines could have led to periodic anoxic episodes, increasing the potential for nutrient release from the sediments. However, the PAC treatments effectively bind phosphorus in the sediments, preventing its release even under anoxic conditions. Monitoring data support this effect, as surface and bottom phosphorus concentrations do not indicate anoxia-driven phosphorus enrichment, and overall bottom water DO has improved since the PAC treatments.

Lower DO concentrations could have lethal or sub-lethal (physiological and behavioral) effects on fish; however, some fish can acclimate to lower oxygen levels and survive concentrations between 1 and 3 mg/L. During anoxic episodes, which are temporary and restricted to the bottom of the water column, they can also avoid low oxygen conditions.

Swan Lake Monitoring Program 2024 Annual Report

Figure ES-2: 2024 Monitoring Results and 2016-2024 Annual Results- Dissolved Oxygen



Note 1: DO concentrations are shown at 1 m from the surface (average of 0.5 and 1 m) and 1 m from the bottom (average of two bottom depths). The range shown (light blue lines) is minimum and maximum daily concentrations from the two loggers at the Dock and the deepest point for the sample collection days. DO trends logged in days between sampling days are not reflected in the plot.

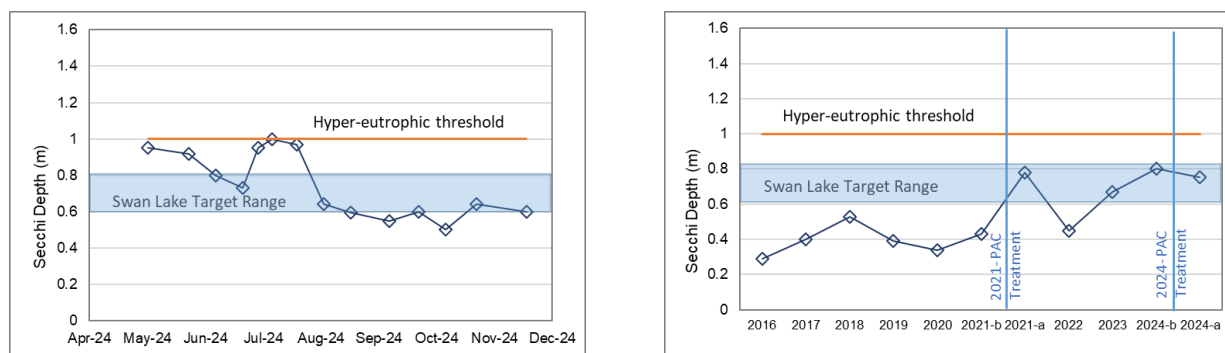
Note 2: Historical data are shown for the average growing period (June-Sep) unless otherwise indicated. Data before 2016 are not shown for legibility.

The pH measured at the lab was about 8 throughout the year. High pH is consistent with high levels of algae. Algae take up carbon dioxide, a weak acid, from the water for photosynthesis, causing the water to become more basic (higher pH).

Water Transparency (Secchi Depth)

Secchi depth represents water transparency, which declines when the algae level increases. In the trophic state classification scheme, growing period average water clarity of under 1 m is the threshold for a hyper-eutrophic condition. The proposed interim target for Swan Lake is 0.6-0.8 m based on correlation with the phosphorus target. In 2024, the average water clarity during the growing season was within the target, with occasional declines to 0.5 m later in the fall.

Figure ES-3: 2024 Monitoring Results and 2016-2024 Annual Results- Secchi Depth



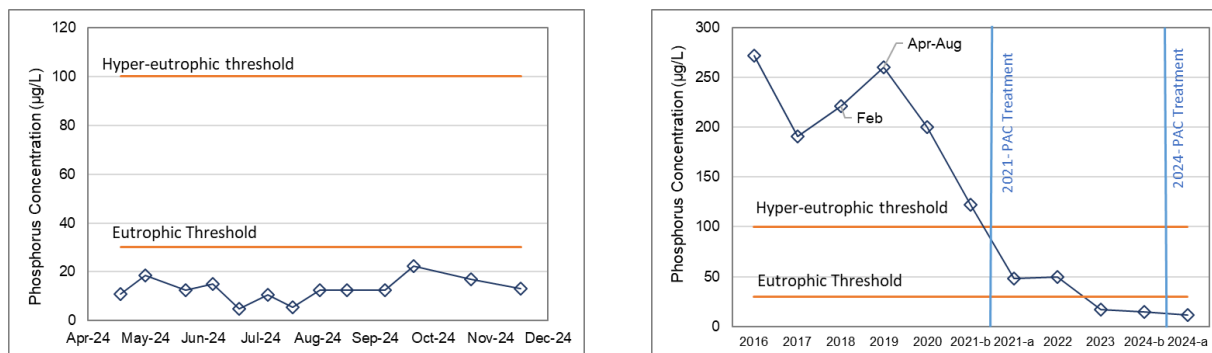
Phosphorus and Nitrogen Concentrations

Phosphorus concentration is the most important indicator of the trophic state in Swan Lake. It is an indication of how prone the Lake is to algae growth.

Phosphorus concentrations above 100 $\mu\text{g/L}$ represent a hyper-eutrophic condition, which lead to high algae concentrations. In 2024, total phosphorus concentration in the top 0.5 and 1.5 m depths averaged under 12 $\mu\text{g/L}$ during the growing season (under the 30 $\mu\text{g/L}$ threshold for eutrophic condition, and well below the interim target of 50-100 $\mu\text{g/L}$). There was significant improvement in phosphorus concentrations after treatment by PAC.

Swan Lake Monitoring Program 2024 Annual Report

Figure ES-1: 2024 Monitoring Results and 2016-2024 Annual Results- Total Phosphorus



Note 1: The 2024 values are averages of samples collected at 0.5 and 1.5 m from the surface.

Note 2: Annual concentrations are summaries of the growing period (June-Sep) unless otherwise indicated.

In 2024, total nitrogen concentrations over the growing season averaged about 0.52 mg/L (below the 0.65 mg/L threshold for a eutrophic condition). In 2024, ammonia and nitrate concentrations (the forms available for uptake by biota) were generally very low (except in September and November), and nitrogen was mainly present in its organic form.

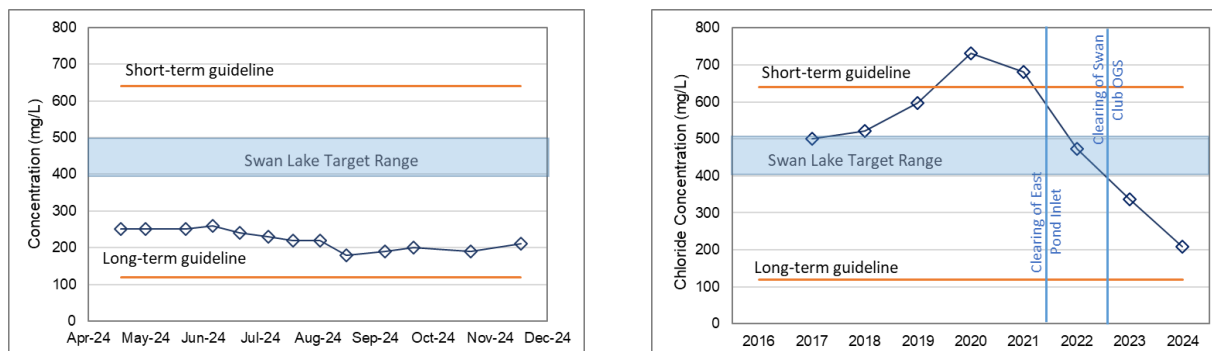
Chloride Concentration

Chloride concentration has been increasing in urban lakes as a result of de-icer application for winter maintenance of roads and walkways. Chloride does not biodegrade, readily precipitate, volatilize, or bioaccumulate. It does not adsorb readily onto mineral surfaces and therefore when introduced, concentrations remain high in surface water.

Chloride guidelines developed for generic environmental data include a long-term guideline (120 mg/L) and a short-term guideline (640 mg/L). The long-term guideline has been developed to protect all organisms (present in Canadian aquatic systems) against negative effects during chronic indefinite exposure. The short-term guideline aims to protect most species against lethality during a sudden hike in chloride concentration for an acute short period (24-96 hrs). These guidelines may be over-protective for areas with an elevated concentration of chloride and associated adapted ecological community. For such circumstances, it has been suggested that site-specific (higher) targets be derived considering local conditions such as water chemistry, background concentrations, and aquatic community structure. The site-specific interim target for chloride for Swan Lake is 400-500 mg/L consistent with 2013-2014 values. In 2024, chloride levels were below the target and declined considerably compared to previous years, continuing previous declines observed since 2020.

Swan Lake Monitoring Program 2024 Annual Report

Figure ES-4: 2024 Monitoring Results and 2016-2024 Annual Results- Chloride

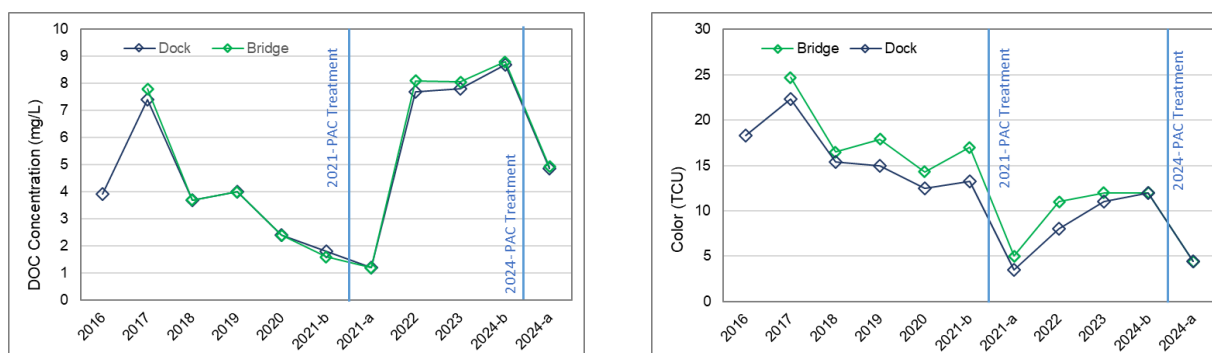


In 2024, water samples were collected from various inlets to the Lake and analyzed for chloride. These data will be used to update the chloride mass balance following the completion of the Flow Diversion Feasibility Study currently underway.

Dissolved Organic Carbon and Color

Dissolved organic carbon (DOC) and colour indicate the organic content of lake water. In 2024, DOC ranged between 4 and 9 mg/L, with color change from 3 to 12 TCU at both stations. DOC in 2022 and 2023 was considerably higher than in previous years, even before treatment. The increase may potentially be associated with the remnants of Phragmites in the Lake, as the roots were not removed. Both color and DOC declined sharply in 2024 following the PAC treatment, which precipitates organic matter.

Figure ES-5: 2016-2024 Annual Results- DOC and Color



Geese Count

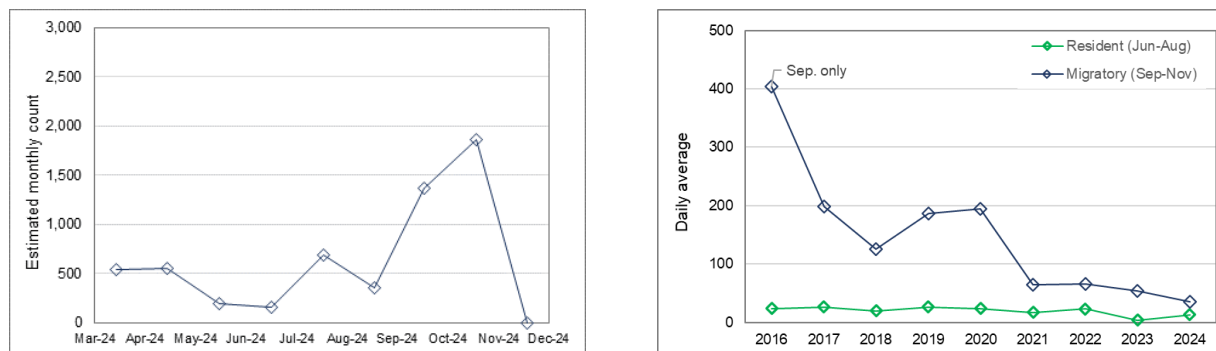
Geese are the primary external source of nutrients in the Lake. Therefore, active geese management is completed annually. The geese control program started in 2014, focusing on resident geese. The program extended to the management of migratory geese in 2016.

The 2024 program focus of control was laser light, avian distress call and limited strategic zinc crackler pyro, as well as geese relocation. A laser emitter was installed on the south island emitting at 1 ft above water surface to prevent overnight goose populations from accumulating in Swan Lake.

The 2024 efforts were very effective in reducing the number of migratory geese visiting the Lake, further lowering the counts. The geese count data helped provide more certainty in the results, and were used to more effectively schedule hazing efforts.

Swan Lake Monitoring Program 2024 Annual Report

Figure ES-6: 2024 Monitoring Results and 2016-2024 Annual Results- Geese Count



Note 1: 2024 data are the sum of counts in each month, compensated for days with no count.

Note 2: Annual trends are shown as daily averages of counts over June-August and September to November, representing resident and migratory geese, respectively.

Fish Inventory

A fish inventory and removal campaign were completed to remove bottom-dwelling fish, which could interfere with the chemical treatment efficacy. A limited number of fish species were caught in the Lake through this intensive effort: Common Carp (non-native), Brown Bullhead, and Fathead Minnow, as well as Goldfish and a hybrid of Carp/Goldfish.

Algal Growth

In 2024, very limited surface scums were observed along the shoreline around the Dock, as well as in the northern bay at the Bridge site. While the Lake was dominated by phytoplankton from late June, surface scums were not widespread.

Samples were collected and sent to the laboratory for phytoplankton analysis. Test results showed higher diversity and 37 to 44 percent lower cyanobacteria count compared to 2023 at the Dock and the Bridge stations, respectively.

Five samples were analyzed for phytoplankton between May and November. The total cyanobacteria cell count was below or close to Health Canada's indicator value for the potential production of cyanotoxins of 50,000 cells/mL, except in August (three and two times higher at the Dock and the Bridge, respectively).

Several algal blooms with potentially toxic cyanobacteria were observed in years before 2011; however, testing completed before 2011 and following treatment (2013-2016) did not detect any Microcystin in the water. In 2016, a bloom was tested and resulted in a Microcystin concentration of 73 µg/L. Extended blooms were observed at several sites in 2018; however, cell density was at half of World Health Organization (WHO)'s threshold for significantly increased human health risk due to toxins. Since the 2021 PAC treatment, very limited surface scum has been observed at Swan Lake.

Summary and Recommendations

Overall, the management activities in recent years that focused on the significant nutrient loadings identified in the water quality improvement study (i.e., chemical treatment and fish management to reduce internal loads and geese management to reduce external loads), were effective at improving water quality in the Lake as shown in reduced phosphorus concentrations and improved dissolved oxygen levels. These improvements represent a positive step towards improving the aquatic habitat in the Lake and meeting the long-term water quality goals.

**Swan Lake Monitoring Program
2024 Annual Report**

In 2024, chloride levels decreased considerably likely due to clearing the blockage at the East Pond inlet and the Swan Lake Club Oil and Grit Separator (OGS), which resulted in reduced untreated flows to the Lake. Dilution by cleaner water could have contributed to lower chloride concentrations in the Lake.

While internal and external source controls successfully reduced nutrient concentrations, the Lake was dominated by phytoplankton, and water clarity did not improve. In addition to a prolonged dry and warm period in late summer and throughout the fall of 2024, this could be partly due to the absence of Submerged Aquatic Vegetation (SAV), which has been replaced by phytoplankton (algae) due to low water clarity. To ameliorate this condition, an SAV planting initiative was implemented in 2023 and 2024 at fenced areas along the north shore of the Lake.

The 2025 monitoring program will follow the recommendation of the Long-Term Management Plan. As per the Long-Term Management Plan, in 2025 at the end of Phase 1, a 5-year review will be completed to evaluate the effectiveness of Core measures and identify the need for additional Complementary measures in Phase 2.

An evaluation of SAVs planting and fish stocking will be pursued, and studies and research on strategies to further reduce chloride concentration in the Lake by diverting runoff will continue. A new pilot project is being considered to apply ultrasound technology for algae control.

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Appendices

Appendix A : Swan Lake Water Quality Inspection Forms

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1. Introduction

Swan Lake is situated in the City of Markham at the intersection of Sixteenth Avenue and Williamson Road, as shown below in Figure 1. Swan Lake has an approximate area of 5.5 ha and a maximum water depth of 4.5 m (from the deepest point to the Lake edges at 210 meter above sea level). Formerly a gravel pit in the 1960s and 1970s, Swan Lake is currently a community feature with multiple trails and urban development.

Several issues were discovered with Swan Lake in 2010, including high phosphorus levels and significant algal blooms during the summer months, which led to low oxygen levels and degraded fish habitats. A Phoslock treatment was administered in 2013 to reduce the phosphorus levels and algal blooms in Swan Lake.

In 2019, the City of Markham conducted a study to define a Water Quality Management Strategy for Swan Lake. The Strategy, which was finalized in July 2020, recommended chemical treatment starting in 2021. In August 2021, 13 tonnes of PAC were applied to the Lake in a controlled manner over several days.

The Swan Lake Long-Term Management Plan, which was developed based on the 2019 Strategy and extensive consultation with stakeholders, was received by Markham Sub Committee in November 2021 and approved by the Council in December 2021. It describes a phased adaptive approach, including Core, Complementary and Alternative measures, and periodic reviews to adapt the Plan to the Lake conditions.

Core and Complementary activities planned in the Long-Term Management Plan and completed in 2024 included enhanced geese management, fish removal, and water quality monitoring, as well as a second application of PAC and additional planting of submerged aquatic vegetation.

Water quality monitoring of Swan Lake has been conducted annually since treatment in 2013 in order to track water quality and the effectiveness of management activities. The 2024 monitoring results presented in this report are part of the ongoing monitoring program that will allow for continuous assessment of the water quality in Swan Lake and help establish a long-term plan for the treatment of Swan Lake.

In 2024, sampling for chloride measurement was also conducted at several locations to determine the relative contribution of each source to the Lake.

Trent University collected samples and launched loggers in Swan Lake in the summer of 2024 to support a study on the environmental fates of lanthanum from La-modified bentonite in the ecosystem of Swan Lake. Data provided by Trent researchers have been incorporated in this report.

2. Monitoring Program

2.1 Annual Water Quality Monitoring

2.1.1 Locations

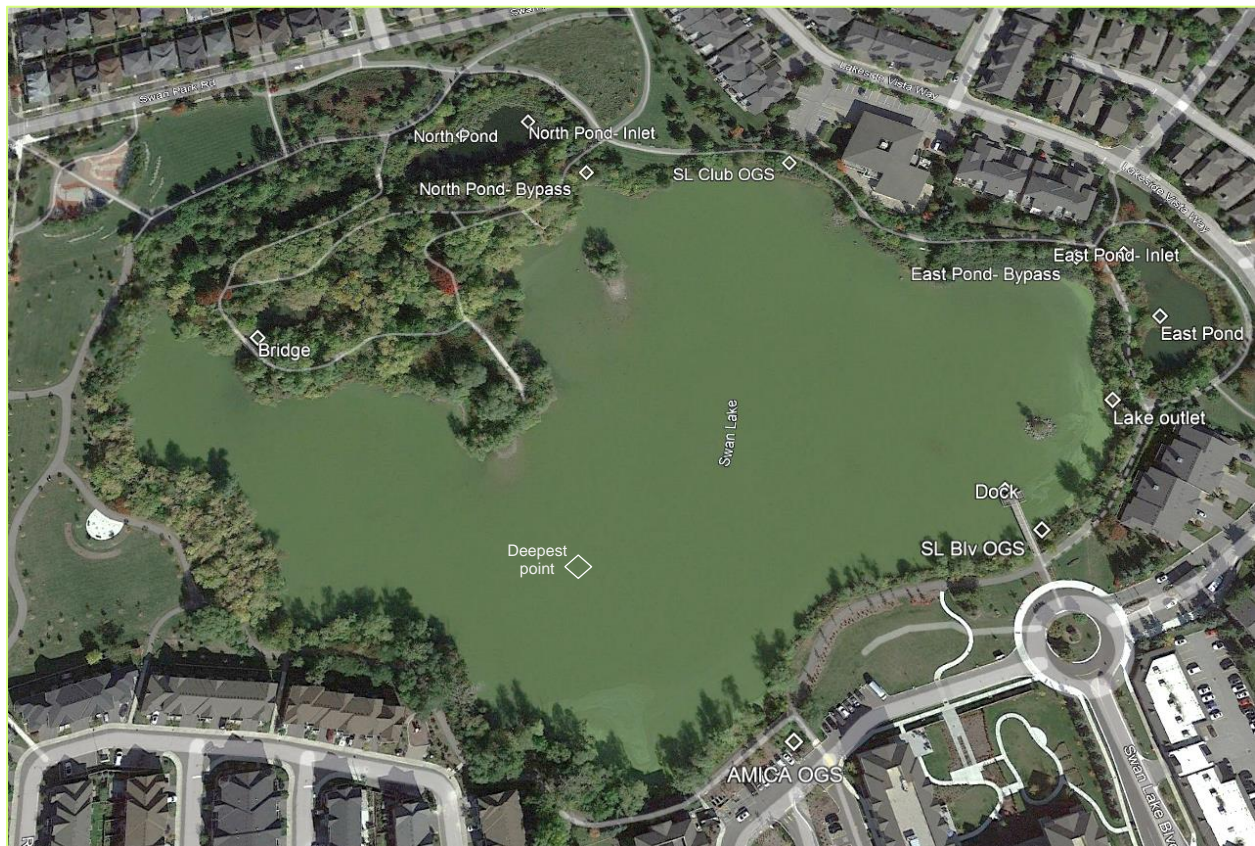
Water quality was monitored at two shoreline sites, the Dock, and the Bridge, as shown in Figure 1. On average, the water depth at the Dock is approximately 2.5 meters, which allows it to represent Swan Lake as a whole. The water depth at the bridge is about 0.5 meters, and it is used to represent the conditions of the shallow bays around Swan Lake. Field testing and sampling for laboratory analysis were completed at both sites to ensure the water conditions at Swan Lake were properly represented.

During the bi-weekly monitoring, samples and measurements were taken at 0.5 m or 1 m increments for the depth of the Lake. The dock site was the deeper of the two sites, allowing for sampling and monitoring from 0.5 – 2.5 m, whereas the bridge site was shallow, and sampling was typically only achievable under the surface, slightly above the bottom of the Lake to avoid sediment contamination.

When the water level dropped to around 2 m, samples were not collected from the 2.5 m depth at the Dock station.

Trent University placed their loggers at the deepest point (shown on the Figure), which was about 3.5 m deep.

Figure 1: Swan Lake and Runoff Monitoring Stations



2.1.2 Duration and Frequency

In 2024, water quality was monitored bi-weekly from April to November.

A total of 13 sampling events were completed.

2.1.3 Parameters and Methodology

Vertical water quality profiling, water transparency readings (Secchi depth), and photographic documentation were performed during each site visit.

Field testing was done utilizing an YSI ProODO meter to determine the temperature and DO at each sampling interval over the vertical profile of the lake. To ensure accurate readings, the meter and probe were stored in a proper carrying bag and regularly calibrated as instructed in the handheld quick-start guide.

A HOBOWare U26 oxygen logger was mounted at the Dock on June 5, 2024, and recorded the DO and temperature of the water every 15 minutes throughout the day. Before the first use, the logger was calibrated for DO at 100% saturation and 0% saturation (using a sodium sulfite solution). An anti-fouling guard was also installed on the sensor cover to protect against fouling. The sensor was placed 1m above the lakebed at the same location as the level logger.

Water transparency was measured as part of the field testing at both the dock and bridge monitoring sites. Transparency was measured using a Secchi disk by lowering it into the water while rotating the handle until the black and white pattern of the Secchi disk was no longer visible. The water depth read from the Secchi disk was then recorded as the transparency (i.e., water clarity).

Water samples for laboratory testing were taken using a horizontal water sampler at different depths. Parameters analyzed at various stations and times included:

- Nutrients including Total Phosphorus (TP), ortho phosphorus, ammonia, nitrate, nitrite, Total Kjeldahl Nitrogen (TKN)
- Chloride, color, Dissolved Organic Carbon (DOC), pH
- Phytoplankton (taxonomic identification and total cell counts)

Observations of Swan Lake were noted, and photographs were taken during each monitoring/inspection site visit. Photographs provide a way to record the condition of vegetation and algae around Swan Lake. Completed inspection forms and photos can be found in Appendix A.

2.1.4 Targets and Thresholds

Generic thresholds for eutrophic and hyper-eutrophic conditions in the lakes are provided in Table 1.

Table 1: Eutrophic State Classification

Parameter	Eutrophic Condition	Hyper-eutrophic Condition
Secchi Depth (m)	1-2.1	<1
Total Phosphorus (µg/L)	31-100	100
Total Nitrogen (mg/L)	0.65-1.20	>1.20
Chlorophyll a (µg/L)	9.1 – 25	> 25

The 2019 Water Quality Management Strategy proposed a set of interim targets for Swan Lake to be used as triggers for management actions if the triggers are tripped in two consecutive years. Numerical values

were defined for total phosphorus (100 µg/L) and Secchi depth (0.6-0.8 m, as updated in 2021 based on correlation with the phosphorus target).

For DO and chloride, Federal and/or Provincial water quality Guidelines¹ or Objectives² were considered for perspective. It should be noted that Swan Lake is not a natural waterbody, and there is no requirement for it to comply with these limits. Where technically and economically feasible, the City will aim to meet these limits to protect and enhance the aquatic environment.

The minimum dissolved oxygen concentration required for the protection of warm water fish is 5 mg/L for water temperatures up to 20 °C, and 4 mg/L for temperatures above 20 °C. Lower concentrations could have lethal or sub-lethal (physiological and behavioral) effects on fish. However, some fish can acclimate to lower oxygen levels and survive concentrations between 1 and 3 mg/L. Furthermore, fish can avoid areas of low oxygen concentrations.

Chloride guidelines developed based on generic environmental data include a long-term guideline (120 mg/L) and a short-term guideline (640 mg/L). The long-term guideline has been developed to protect all organisms (present in Canadian aquatic systems) against negative effects during indefinite exposure. The short-term guideline will protect most species against lethality during a sudden hike in chloride concentration for a short period (24-96 hrs). These guidelines may be over-protective for areas with an elevated concentration of chloride and associated adapted ecological community. For such circumstances, it has been suggested that site-specific (higher) targets be derived considering local conditions such as water chemistry, background concentrations, and aquatic community structure. The site-specific interim target for chloride in Swan Lake is 400-500 mg/L consistent with 2013-2014 values.

For Cyanotoxins, the Health Canada guideline for recreational activities was updated from 20 µg/L to 10 µg/L in 2022³. The 2022 guidelines also provide indicator values for the potential production of cyanotoxins including:

- Total cyanobacteria cells: 50,000 cells/mL
- Total cyanobacterial biovolume: 4.5 mm³/L
- Total chlorophyll a: 33 µg/L

2.2 Runoff Monitoring

In the Swan Lake catchment, salt application for winter maintenance is mainly completed by the City's Road department and the Swan Lake Village Corporation.

Winter maintenance of 1 km of the catchment roads and sidewalks is completed by the City of Markham. The City prescribes and tracks the quantity of salt distributed to the City roadways based on current and future forecast models and temperatures to determine the required action and material usage in compliance with the desired level of service and O.Reg. 239/02 requirements.

The remaining roads and parking areas, as well as private walkways and driveways, are serviced privately. As per the Village Amenities Committee (VAC), the Village Corporation employs "a qualified, reputable cleaning and maintenance service employing Smart About Salt principles to plow/shovel and their insurance recommends the de-icing methods of rock salt, applied as necessary to maintain their insurance and mitigate potential claim".

¹ Canadian Council of Ministers of the Environment (CCME) Water Quality Guidelines for the Protection of Aquatic Life (<http://ceqg-rcqe.ccme.ca/en/index.html>)

² Ontario Provincial Water Quality Objectives (PWQO) (<https://www.ontario.ca/page/water-management-policies-guidelines-provincial-water-quality-objectives#section-13>)

³ Health Canada, 2022. Guidelines for Canadian Recreational Water Quality, Cyanobacteria and their Toxins, Ottawa, Ontario.

Chloride in salting materials is readily dissolved in water and transported overland by runoff or infiltrated into soils, contaminating groundwater and surface water. A fraction of chloride in applied road salt is retained by soil and is not observed in surface runoff. As a result, salt loading to surface water occurs primarily in winter and spring during melt conditions but continues through the summer and fall via the discharge of impacted groundwater, dry deposition of dust to the lake surface, and non-point source runoff washing dry salt from land surfaces. Salt accumulated in the ponds could also be discharged into the Lake through the flushing of stormwater ponds.

In 2024, water samples were collected from various inlets to the Lake to quantify and determine the relative contribution of each source to chloride concentration in Swan Lake. Samples were collected from both ponds' inlets, as well as outfalls from the ponds and OGS's to the Lake.

Conductivity was also measured in a number of samples, as this parameter can be used as a surrogate for chloride. Samples were collected during four snowmelt events from January to March 2024.

2.3 External Data

Trent University collected water, sediment and biological samples and launched loggers in Swan Lake in the summer of 2024 to support a study on the environmental fates of lanthanum from La-modified bentonite in the ecosystem of Swan Lake. Data provided by Trent researchers included DO and temperature profiles and Chl-a concentrations at the deepest point of the Lake.

2.4 Water Level Monitoring

Prior to 2024, water level was monitored using HOBOWare U20 Water logger mounted at the Dock. In 2024, the logger was replaced with a Dipperlog vented data logger from Heron Instrument, which eliminated the need for barometric compensation. The data logger records the pressure and temperature of the water every 15 minutes.

3. Results

3.1 2024 Water Quality

The following sections discuss water quality results in 2024.

3.1.1 Dissolved Oxygen and Temperature

Table 2 provides the measured DO profile over the 2024 monitoring period. At the Dock station, all day-time surface and deep-water concentrations measured during sampling events were above 6 mg/L. All measurements at the Bridge indicated a DO concentration of above 4.6 mg/L with a summer average concentration of 5.6 mg/L

Table 2 also provides measured temperature profiles in 2024, indicating warm water throughout the depth in the summer months.

Profiles of temperature and dissolved oxygen (see Figure 3) indicate that Swan Lake was transiently stratified in May and June (when temperature decline is greater than 1 °C per m of depth). Transient stratification can cause reduced mixing/aeration and lead to anoxia with the release of nutrients from the sediments.

Data collected by Trent University on July 15 shows stratification at the deepest point, where DO drops from 7.5 mg/L at 2 m to below 1 mg/L at depths greater than 3 m (Figure 2). The decline in DO occurs below the top of the thermocline where temperature decreases with depth by 1°C per m or more, reducing mixing and reoxygenation of the bottom water below that depth.

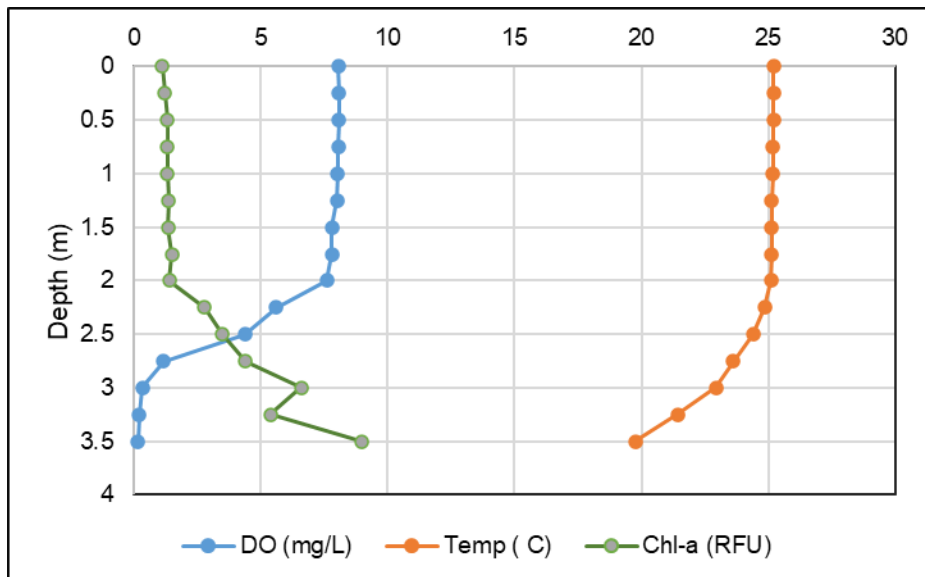
In 2024, in addition to using the YSI ProODO meter for bi-weekly measurements of temperature and DO over the vertical profile, a continuous logger was also placed 1 m above the lake surface to measure the diurnal changes in DO. DO increases during daylight hours when photosynthesis occurs and decreases at night when respiration continues but photosynthesis does not.

Continuous measurement of DO from July to November indicated that DO concentrations have a diurnal pattern, typical of freshwater lakes. These data are mostly consistent with the data logger used by Trent University researchers at the deepest point. While surface concentrations were always above 4 mg/L regardless of the time of day, deep water concentrations dropped to anoxic levels for several days in the second half of August and several days mid September, when there was minimal precipitation and warmer than usual weather.

Table 2: Measured Daytime DO and Temperature

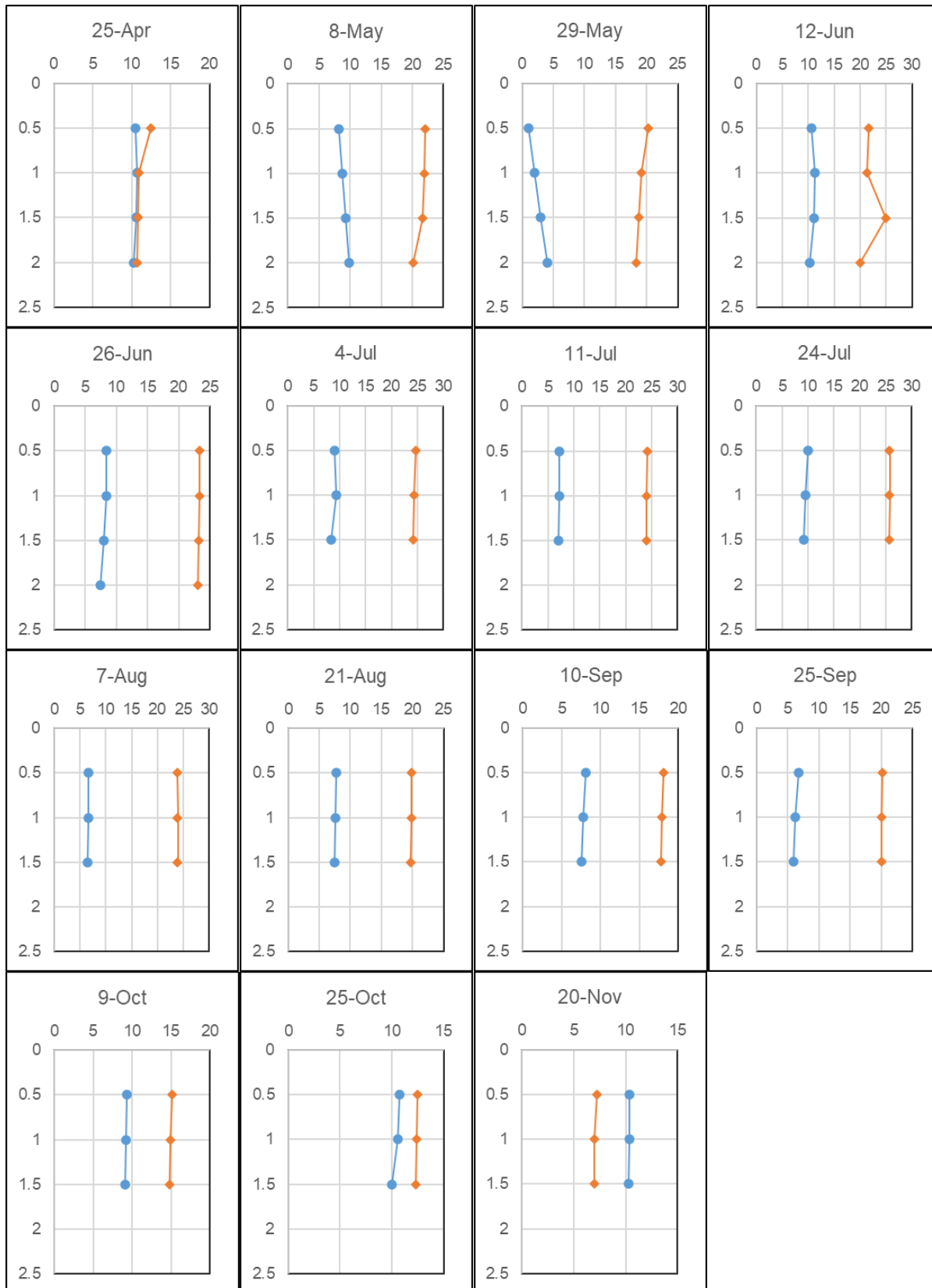
Date	DO Concentration (mg/L)					Temperature (°C)				
	Bridge	Dock				Bridge	Dock			
	Depth (m)	Depth (m)				Depth (m)	Depth (m)			
	0.5	0.5	1	1.5	2	0.5	0.5	1	1.5	2
4/25/2024	10.1	10.5	10.7	10.6	10.3	11.0	12.5	10.9	10.8	10.7
5/8/2024	-	8.3	8.7	9.3	9.9	-	21.6	21.3	25.0	20.0
5/29/2024	13.9	10.8	10.9	11.4	10.3	22.1	22.0	21.9	21.6	20.1
6/12/2024	11.1	10.6	11.2	11.1	10.3	18.0	20.2	19.2	18.7	18.3
6/26/2024	4.9	8.4	8.4	8.0	7.4	23.7	23.4	23.4	23.2	23.1
7/4/2024	6.8	9.1	9.3	8.4	-	23.5	24.6	24.4	24.1	-
7/11/2024	5.9	7.2	7.2	7.0	-	23.2	24.1	24.0	24.0	-
7/24/2024	5.2	9.9	9.6	9.1	-	25.2	25.7	25.7	25.6	-
8/7/2024	5.4	6.7	6.6	6.5	-	23.0	23.8	23.9	23.9	-
8/21/2024	4.6	7.7	7.5	7.5	-	18.2	19.8	19.8	19.7	-
9/10/2024	5.7	8.1	7.8	7.6	-	16.9	18.1	17.9	17.8	-
9/25/2024	6.0	6.7	6.1	5.9	-	19.8	20.2	20.0	20.0	-
10/9/2024	8.5	9.4	9.2	9.1	-	15.1	15.1	14.9	14.8	-
10/25/2024	9.7	10.7	10.6	10.0	-	10.2	12.5	12.4	12.3	-
11/20/2024	9.0	10.4	10.3	10.2	-	7.3	7.2	7.0	7.0	-

Figure 2: Vertical Profiles on July 15, 2024, at the Deepest Point



Data courtesy of Trent University

Figure 3: Temperature (orange) and DO (blue) Profile at the Dock Station in 2024



Note: The vertical axis shows depth (m), while the horizontal axis represents both Temperature (°C) in orange, and DO (mg/L) in blue.

3.1.2 Water Transparency

A robust measure of algal biomass is the measurement of the Secchi disk depth or transparency.

Table 3 summarizes the results of the water transparency readings. Transparency at the Dock station averaged 0.76 m over the growth season, and was within the interim target for Swan Lake of 0.6-0.8 m. Water transparency at the Bridge site was generally equal to the water depth.

Table 3: 2024 Secchi Depth Results (m)

Date	Dock	Bridge
25-Apr	0.95	0.63
8-May	0.92	0.565
29-May	0.8	0.5
12-Jun	0.73	0.34
26-Jun	0.95	0.46
11-Jul	1	0.5
24-Jul	0.97	0.5
7-Aug	0.64	0.5
21-Aug	0.59	0.46
10-Sep	0.55	0.43
25-Sep	0.6	0.53
9-Oct	0.5	0.41
25-Oct	0.64	0.4
20-Nov	0.6	0.4

3.1.3 Nutrients Concentrations

Samples collected during each visit were tested for TP, Orthophosphate, TKN, Nitrate, Nitrite, and Ammonia. The results can be found in Figure 4 for the Dock site and Figure 5 for the Bridge site. The Certificates of Analysis from Bureau Veritas Laboratories are in Appendix B. Nutrient concentrations are shown for the depths sampled.

Total phosphorus concentration at 0.5 and 1.5 m depths averaged under 12 µg/L during the growing season and was below 30 µg/L throughout the year (threshold for a eutrophic condition).

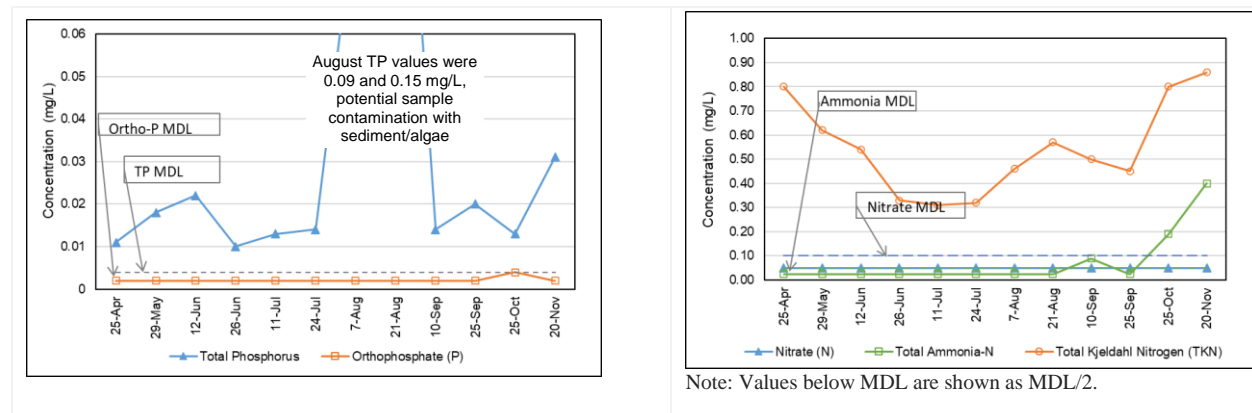
Following the PAC treatment, total phosphorus declined to very low levels near the laboratory detection limit, then increased over the summer reaching a peak in September, consistent with phosphorus loading to the lake from external sources. Orthophosphate and total ammonia concentrations were low throughout this period in the surface and bottom water with few exceptions indicating that internal loading due to anoxia was not significant. In the late fall, total ammonia concentrations increased, likely due to reduced uptake by algae.

Total nitrogen concentrations over the growing season averaged about 0.52 mg/L (below the 0.65 mg/L threshold for a eutrophic condition). Total concentrations at the Bridge site averaged 0.5 mg/L. Ammonia and nitrate are the directly bioavailable forms of nitrogen, with Ammonia being the most usable form for algae. In 2024, Ammonia and Nitrate concentrations were generally close to or below Method Detection Limit (MDL), and nitrogen was mainly present as organic compounds (i.e., TKN less Ammonia) with the exception of fall samples. Bioavailable nutrient pulses (orthophosphate and ammonia) in late summer and fall are consistent with the release of these nutrients due to episodic anoxia and decomposition of organics, including algae.

Figure 4: 2024 Measured Nutrients Concentrations - Dock Site



Figure 5: 2024 Measured Nutrients Concentrations - Bridge Site



3.1.4 pH

pH measured at the lab ranged from 7.4 to 8.38 throughout the year, which is within the PWQO range (6.5-8.5).

3.1.5 Chloride in Lake and Runoff

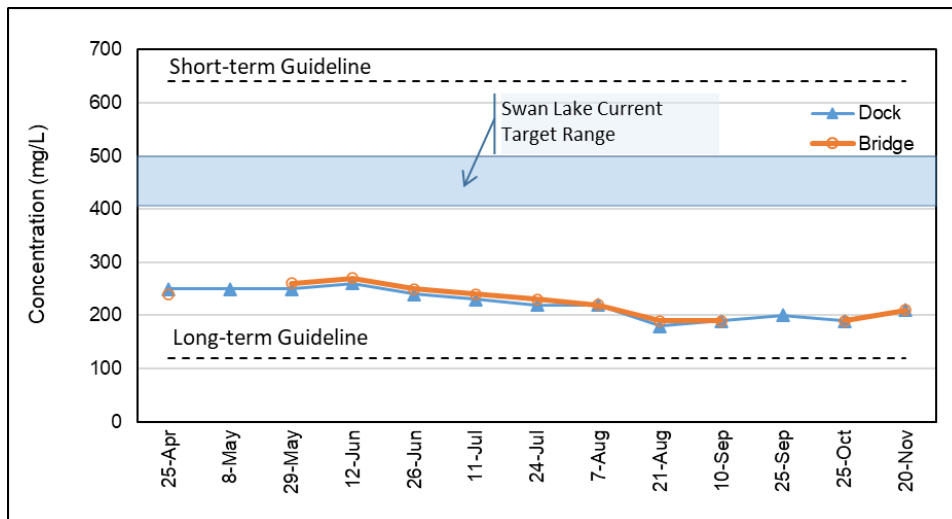
Surface samples collected during each visit were also analyzed for Chloride, as summarized in Figure 6.

Water quality testing results indicated that the samples contained between 180 and 270 mg/L of Chloride.

In 2024, chloride levels further decreased considerably compared to 2021 and 2022, likely due to clearing the blockage at the East Pond inlet and Swan Club OGS, which resulted in lower untreated flows to the Lake.

Chloride guidelines developed based on generic environmental data include a long-term guideline (120 mg/L) and a short-term guideline (640 mg/L). The interim target for chloride is 400-500 mg/L consistent with 2013-2014 values. In 2024, all samples met these targets.

Figure 6: Chloride Concentrations in Swan Lake in 2024



In 2024, water samples were collected from various inlets to the Lake and analyzed for chloride.

These data, along with scattered data from previous years, are shown in Table 4. Based on this limited dataset, chloride concentration in the spring runoff from the pond catchments is about 1700 mg/L for the East Pond and 880 for the North Pond (average of pond inlet measurements, except for January 13th, 2022, and Jan 24, 2024, where samples were collected from standing water). This concentration would not usually end up in the Lake, except through the East Pond bypass when the pond inlet was blocked. At other times, the bypass would carry ‘cleaner’ water (after the first flush), with concentrations around 200 mg/L. Flows from the ponds to the Lake have an average concentration of about 380 mg/L (average of pond and outlet concentrations).

The runoff collected from the Swan Lake Blvd. OGS contained an average of about 2000 mg/L of chloride, while from the AMICA OGS had a concentration of about 620 mg/L. Samples collected from the Swan Club OGS has an average concentration of about 2400 mg/L. Samples were also collected from the shoreline runoff, which resulted in very low chloride concentrations (about 25 mg/L).

These data will be used to update the chloride mass balance following the completion of the Flow Diversion Feasibility Study, which is currently underway.

Table 4: Chloride Concentrations in Runoff

Date	Inflow to Ponds		Bypass from Pond to Lake		Inflow to Lake from Ponds				Inflows to Lake from OGS		
	East Pond	North Pond	East Pond	North Pond	East Pond-in pond	From south	North Pond-in pond	Road	Swan Lake Blvd.	AMICA	Swan Club
3/20/2012 *	577	673			572		56				
3/26/2021	957	98.5			343		199				
4/11/2021		79	131			673					
1/13/2022 **	13200								3160		
2/15/2022	2340	2120						326	836	360	
3/6/2022	380	410			410		180		1200	610	
3/16/2022	3700	3100							4800	470	
3/24/2022	1200	1100	150						1900	240	
4/6/2022	2800		350							1100	
1/18/2023	2000	1200	240							120	
2/8/2023	3900	650								450	
2/9/2023	360	340				960		120	420	300	
3/24/2023	1300	630	180								
1/15/2024	1900	210	270	120					3300	2100	4200
1/24/2024 **	8400										8200
1/26/2024	560		150						320		680
2/29/2024	2100		220	280							
Average	1720	881	210	200	384				1943	620	2440

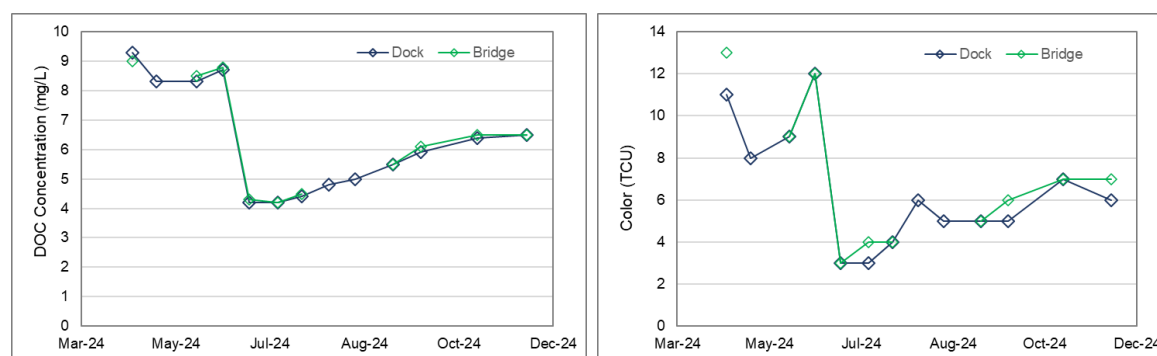
* Data were used cautiously since the exact location of samples and sampling conditions are not known.

** Standing water, not used in calculations.

3.1.6 DOC Concentrations and Color

Surface samples collected during each visit were also analyzed for DOC and Color. The results are summarized in Figure 7. Both DOC and colour dropped after the treatment and remained lower than pre-treatment levels throughout the season.

Figure 7: Measured DOC and Color in 2024



3.1.7 Algae Growth

In 2021, samples were collected before and after chemical treatment and sent to the laboratory for phytoplankton and cyanobacteria identification. Test results are summarized in Figure 8 below and show a significant reduction in concentrations following the treatment, consistent with the particle scavenging characteristics of the treatment chemicals. Phytoplankton density increased almost five weeks post-treatment to values comparable to pre-treatment levels.

In 2022, limited algae scum was observed in early June, and while the Lake was dominated by phytoplankton for the remainder of the monitoring period, surface scums were not widespread. Four sets of samples were collected from the Lake between August and December for phytoplankton identification, as shown in Figure 9. These results should be considered with caution due to lab errors in the identification of *Microcystis*. In general, the 2022 results showed lower diversity and higher total counts compared to 2021.

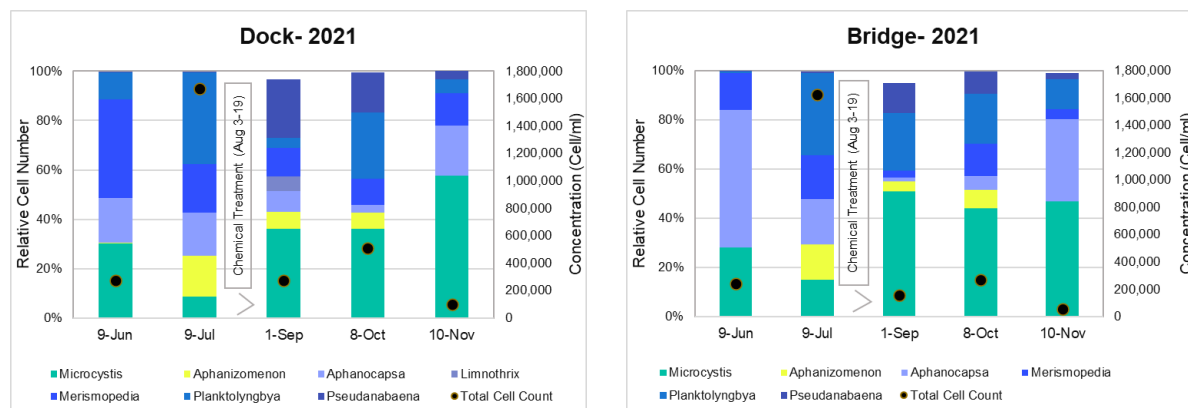
Abraxis tests were performed on June 29, July 14, and August 11, 2022, and resulted in Microcystin levels below the recreational limit (recently updated to 10 µg/L). Nonetheless, the presence of known toxin producers at high cell densities suggests that cyanotoxins can potentially occur at elevated concentrations that exceed recreational guidelines. Toxin concentrations can vary tremendously over small spatial and temporal scales, and it is, therefore, possible that higher concentrations occurred elsewhere in the Lake or at different times.

In 2023, seven samples were analyzed for cyanobacteria identification. The results are shown in Figure 10 and indicate significantly lower cell counts compared to 2022. Cell counts in August and September decreased from about 2,000,000 cells/mL in 2022 to 50,000 cells/mL in 2023. The dominant genera of *Microcystis* (Chroococcales order) and *Cylindrospermopsis* (Nostocales order) stayed as such in both 2022 and 2023, with several genera of the Synechococcales order also present in relatively high percentages in 2023.

In 2024, five samples were collected for phytoplankton identification as shown in Figure 11. Total cyanobacteria cell counts were about 40% lower in 2024 compared to 2023, with average growing season concentrations at 63,000 and 44,000 cells/mL at the Dock and Bridge stations, respectively, compared to the guideline of 50,000 cells/mL. While total cell counts were the highest in June and July in 2023, in 2024, the highest counts were in August.

Besides the actual identification and counting of cells, analysis of chlorophyll-a (chl-a), which is the green algal pigment used in photosynthesis, provides a measure of algae biomass. Concentrations of chl-a were recorded by Trent University using an Enigma probe at a depth of 1 m, indicating an average of 16 µg/L over the growing season, which is within the eutrophic state.

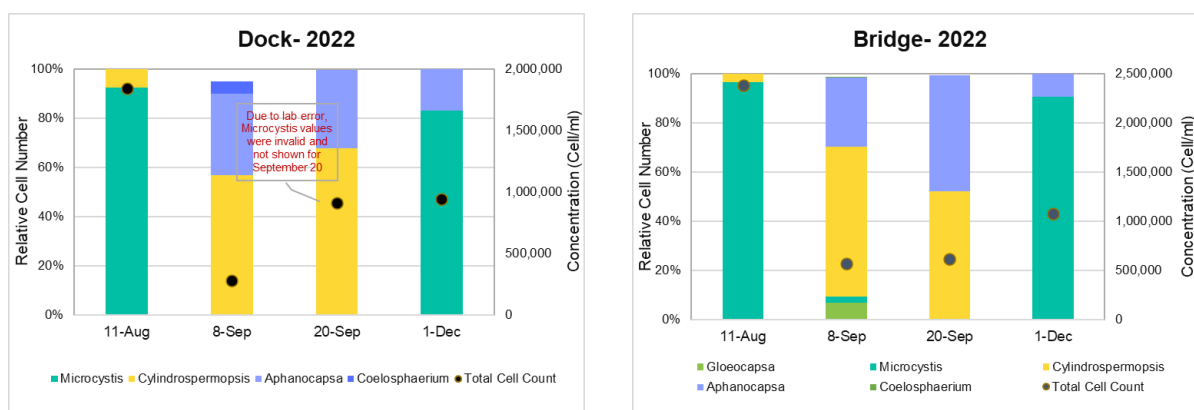
Figure 8: Planktonic Cyanobacteria Population in Swan Lake in 2021



Other genera present at less than 3% include: Planktothrix and Coelosphaerium

Other genera present at less than 3% include: Aphanothece, Gomphosphaeria, Phormidium, Planktothrix and Limnithrix

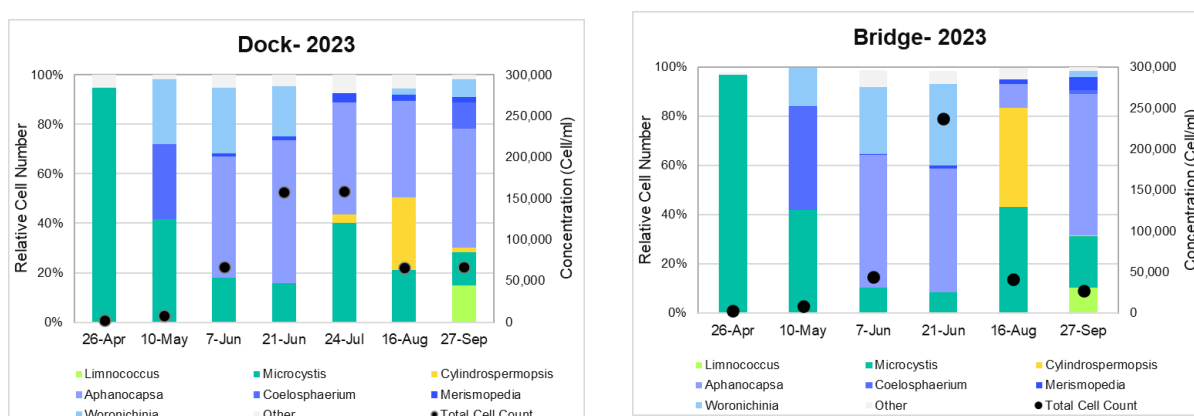
Figure 9: Planktonic Cyanobacteria Population in Swan Lake in 2022



Other genera present at less than 3% include: Gloeocapsa, Anabaena, Oscillatoria and Merismopedia

Other genera present at less than 3% include: Anabaena, Oscillatoria, Coelosphaerium and Merismopedia

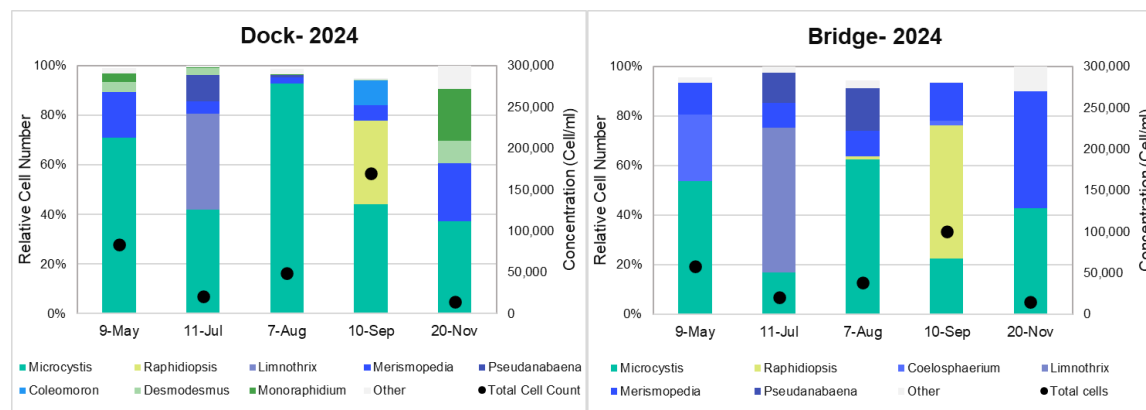
Figure 10: Planktonic Cyanobacteria Population in Swan Lake in 2023



Other genera present at less than 3% include: Gloeocapsa, Gomphosphaeria, Anabaena, Aphanizomenon, Woronichinia, Chroococcus, Glaucoispira, Leptolynbya, Snowella

Other genera present at less than 3% include: Gloeocapsa, Gomphosphaeria, Pseudanabaena, Chroococcus, Glaucoispira, Microchaete, Snowella

Figure 11: Planktonic Cyanobacteria Population in Swan Lake in 2024



Other genera present at less than 3% include: Coelosphaerium, Woronichinia
Total cell count referred to total cyabobacteria until 2023, and total phytoplankton in 2024.

Other genera present at less than 3% include: Coeleomorion, Woronichinia

3.2 2024 Water Level

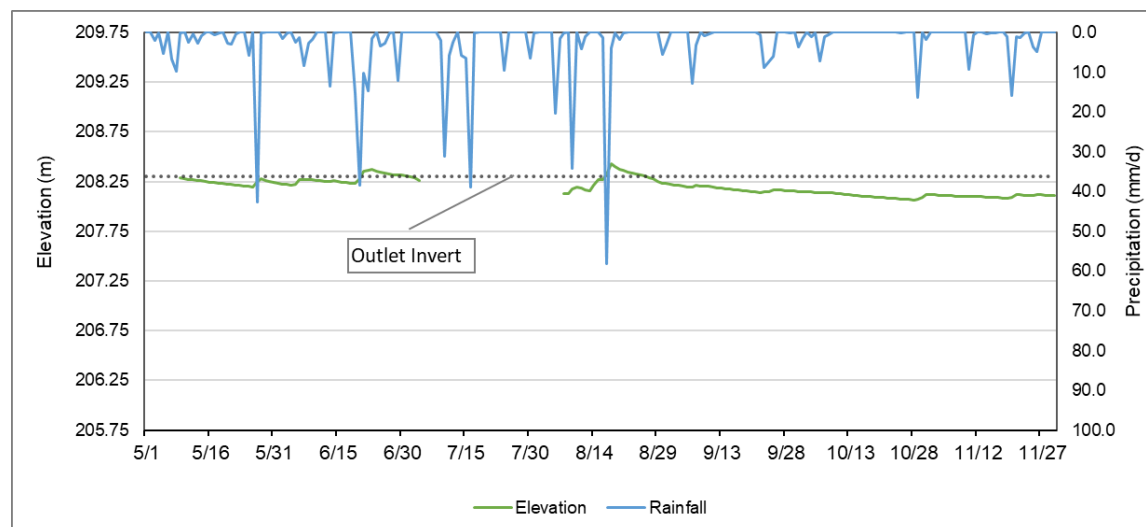
In 2024, the measured water level changed from a maximum of 208.45 m in August to a low of 207.26 m in November. Total precipitation in 2024 was 1055 mm, as recorded at the Markham Museum station.

The maximum water level recorded or estimated between 2017 and 2024 ranged from 208.25 m to 208.48, when total precipitation ranged from 670 to 934 mm.

While 2024 was a wet year during the summer, there was an extended dry and warm period from August to the end of the Fall, which resulted in low water level and lower oxygen levels at depth .

Measured water level and daily precipitation data from the nearby rain gauge are shown in Figure 12.

Figure 12: Lake Elevation Records and Precipitation in 2024



3.3 Water Quality Trends

Water quality monitoring of Swan Lake has been conducted annually since treatment in 2013 to track water quality and the effectiveness of implemented mitigation measures. The following paragraphs and Figure 13 provide a summary of water quality trends for the period of monitoring.

Dissolved Oxygen (DO)

Historical records of DO and temperature profile show that Swan Lake thermally stratifies during the summer despite its shallow depth. Anoxic conditions were observed at depths below 2 m, to as high as 1 to 1.5 m (in 2016). The majority of surface concentrations have been above 5 mg/L since 2014. In 2024, all day-time surface and bottom concentrations measured during sampling events at the Dock station were above 6 mg/L. DO concentrations, however, have a diurnal pattern, often decreasing at night. Data collected by two continuous loggers in 2024 indicated that deep water concentrations dropped to anoxic levels for several days in the second half of August and several days mid September, when there was minimal precipitation and warmer than usual weather potentially resulting in transient thermal stratification and reduced mixing of the bottom water.

Water Clarity (Secchi Depth)

In Swan Lake, Secchi depth has typically been quite low throughout the summer, but it increases in November, reflecting the end of the growing period for phytoplankton. The average annual values shown in Figure 13 are all below 1 m, except in 2014 and 2021, following chemical treatment. In 2024, water clarity, which increased to 1 m following the treatment, was above 0.5 m for the remainder of the monitoring period until late November.

Total Phosphorus (TP)

Average growing period (May - September) TP concentrations indicated hyper-eutrophic conditions in earlier monitored years except for the post-treatment years, 2013 and 2014. TP, has been consistently low since the 2021 treatment and dropped to below 30 µg/L in 2023 and 2024.

Nitrogen Compounds

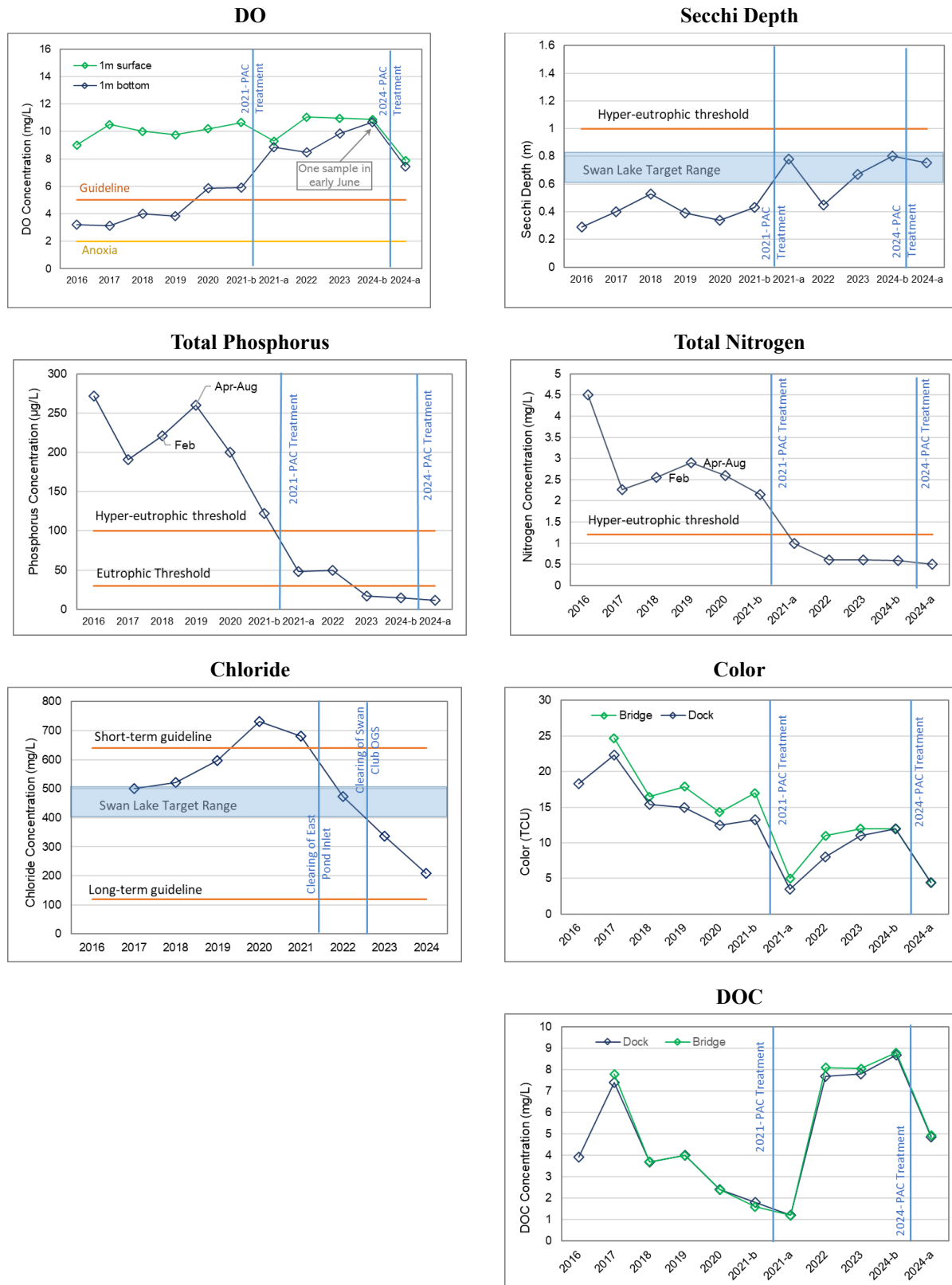
Total nitrogen (TN) concentration over the growing period was above the 1.2 mg/L threshold for a hyper-eutrophic condition in earlier years, except in the post-treatment year, 2014. TN dropped to below the eutrophic threshold after the 2021 and 2024 treatment. Nitrogen is, however, not believed to be the limiting nutrient for eutrophication in Swan Lake (i.e., the nutrient that elicits the largest response in algae growth).

Inorganic nitrogen compounds (NO₂, NO₃, and NH₃) have often been below detection limits, indicating relatively low levels of bioavailable nitrogen concentrations. In 2024, ammonia and nitrate concentrations were generally very low (except higher ammonia in the fall), and nitrogen was mainly present as organic matter.

Chloride

Chloride concentrations were increasing in Swan Lake over the past few years with a slight drop in 2021. Removing the blockage at the East Pond inlet and cleaning of the Swan Club OGS resulted in lower untreated flows to the Lake, lowering chloride concentration in Swan Lake post 2022. The Long-Term Management Plan for the Lake suggests that the main mechanism for lowering chloride levels would be source control. Emerging technologies (chloride removal using biochar) and the feasibility of flow redirection are currently being studied.

Figure 13: Historical Water Quality Results (Growing-Season Averages)



Algae Blooms and Cyanobacteria

Table 5 provides a summary of the observed algae blooms in the Lake over the years. It also shows any tests conducted to measure toxins (mainly in terms of Microcystin concentration) in the Lake water.

Table 5: Records of Algae Blooms and Toxicity

Year/Period	Algae Blooms Observation	Toxicity Test Result
Before 2011	Several blooms of cyanobacteria were observed	Microcystin concentration under detection limit
2013-2016	No apparent cyanobacteria proliferation and blooms; no resident concern related to the Lake's water quality	Microcystin concentration under detection limit
2016	A bloom was detected at one location	Microcystin concentration of 73 µg/L in one sample tested (recreational guideline is 20 µg/L)
2017	No bloom was observed	-
2018	Extended blooms were observed at several sites	Not tested for toxicity; cell density was at half of WHO's threshold for significantly increased risk for human health
2019	Extended blooms were observed at several sites	Microcystin toxicity was measured with test strips; all samples were below 10 µg/L
2020	Blooms were observed at several sites	Microcystin toxicity was measured with test strips; all samples were below 10 µg/L
2021	Blooms were observed at several sites before treatment; the high biomass was inhibited by the August PAC treatment; however, by October, cyanobacteria were as high as in previous summers and falls.	Not tested for toxicity
2022	Surface scums were not widespread; Lab results showed lower diversity and higher total counts compared to 2021.	Microcystin toxicity was measured with test strips; all samples were below 10 µg/L
2023	Surface scums were not widespread; Lab results showed higher diversity and significantly lower total counts compared to 2022.	Not tested for toxicity
2024	Surface scums were not widespread; Lab results showed higher diversity and about 40% decline in total counts compared to 2023.	Not tested for toxicity

While internal and external source controls in recent years have successfully reduced nutrient concentrations to below the specified targets, the Lake has been dominated by phytoplankton, and water clarity improvements were modest. This could be due to warmer weather and partly due to the absence of SAV, which has been replaced by phytoplankton (algae) due to historically high concentrations of total phosphorus. SAV compete with algae for nutrients and light, and the establishment of SAV growth may help to reduce phytoplankton blooms over the growing season.

SAV would prevent sediment resuspension, take up nutrients, and act as habitat for zooplankton, which in sufficient densities would help control algal blooms. The return of SAV could be key to shifting the Lake to a clear state and, this shift seems unlikely without active bio-manipulation to break the cycle of high turbidity- phytoplankton dominance – high turbidity⁴. Therefore, in 2023 and 2024, the Toronto Region Conservation Authority (TRCA) was contracted to implement a SAV planting pilot project in fenced areas

⁴ Scheffer, M. Alternative Attractors of Shallow Lakes. The Scientific World (2001) 1, 254-263.

along the north shore of the Lake. An evaluation of the SAV planting success and habitat conditions is being considered in 2025.

4. Geese Management

4.1 Geese Management Approach

Geese reduction at Swan Lake is necessary due to the nutrient load they contribute to the Lake.

In 2024, the geese management program was completed by two external contractors.

The Wildlife Management Group Inc., an external consultant, was hired to use science- based and Environment Canada-approved techniques for managing Canada goose. Activities included nest depredation (nest monitoring, and subsequent removal between April and May), and adult goose mitigation through laser light, avian distress call and limited strategic zinc crackler pyro.

The TRCA was hired to relocate resident geese from Swan Lake (and Toogood Pond) and to remove the nests and eggs from the area.

The strobe lights purchased in 2020 at the request of Friends of Swan Lake Park were also installed on the Lake and the two adjacent stormwater management ponds. Strobe lights work by using a solar-powered LED light that flashes every two seconds and is intended to disrupt the geese's sleep patterns and discourage them from staying on the Lake.

4.2 Geese Count

In 2024, the geese count was completed by the consultant, City staff, and volunteers from the community.

WMG recorded the number of geese observed during each visit. Staff counted the number of geese every two weeks, coinciding with the water quality sampling site visits. All counts and other wildlife observations were recorded in a geese count App developed using ArcGIS Survey123 software.

4.3 Results

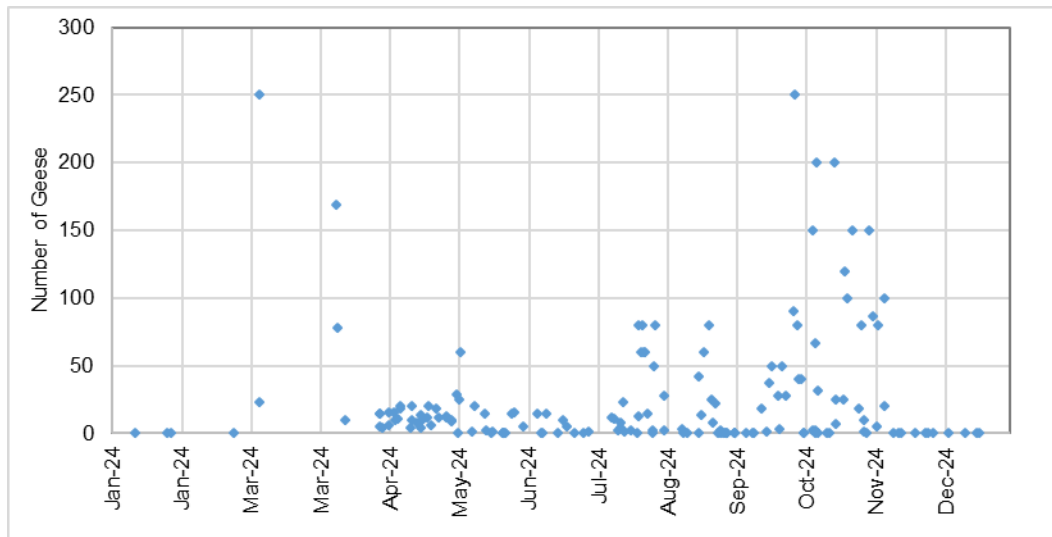
Figure 14 illustrates the number of geese counted at Swan Lake throughout the 2024 monitoring period.

In this figure, a significant increase in geese numbers is evident when they migrate south; however, the mitigation measures employed effectively reduced the number of geese present at different times of the day. Fewer geese were counted in August and September of 2024 compared to previous years, likely due to the prolonged warm weather conditions and delayed migration. Any impact that strobe lights might have had on the geese count is not readily evident from the data.

In total, 14 Canada Geese were rounded up from Swan Lake in 2024. Twenty-six adults and six goslings were rounded up and relocated from Toogood Pond.

In addition, eight nests and 68 eggs were managed at Swan Lake in April and May.

Figure 14: 2024 Geese Count Results at Swan Lake

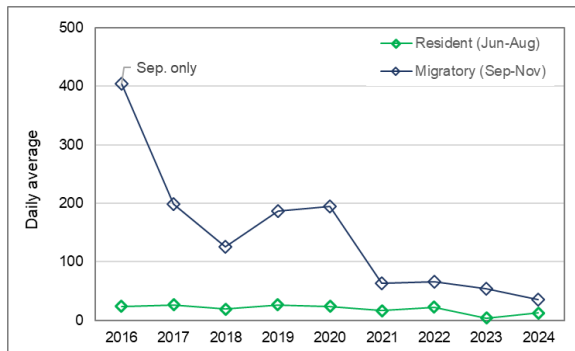


4.4 Historical Trends

Active geese management has been completed annually since 2014. The geese management program focused on resident geese at the beginning and extended to the management of migratory geese in 2016.

Daily Averages of counts are shown for each year in Figure 15. Data are summarized for June to August and September to November, representing resident and migratory geese, respectively. Despite a general increase in geese population in Southern Ontario, the numbers at Swan Lake have been controlled over the past years.

Figure 15: Historical Geese Counts



5. Management Activities

5.1 Chemical Treatment

Between June 17 to June 25, 2024, about nine tonnes of PAC was applied to reduce phosphorus concentrations and ultimately algal production in the Lake. Each application event was separated by one or two days of downtime to allow for floc formation and environmental testing.

5.2 Fish Inventory and Removal

The Long-Term Management Plan for Swan Lake (2021) has a provision for managing bottom-dwelling fish to reduce sediment disturbance.

Since 2021, the City has hired the TRCA annually to complete a fish inventory and removal operation.

In 2021, three fish species were captured across five different sampling events. The three species were Brown Bullhead (*Ameiurus nebulosus*), which were relocated to Milne Dam, Common Carp (*Cyprinus carpio*), which were euthanatized, and Fathead Minnow (*Pimephales promelas*), which were returned to the Lake.

The same three fish species as in 2021 were captured during one electrofishing sampling event and one netting sampling event on August 23 and 24, 2022. In 2022, the TRCA was informed by the Ontario Ministry of Natural Resources and Forestry (OMNRF) that a Fish Stocking license would not be granted due to the possibility of disease transfer. Instead, OMNRF requested that both Common Carp and Brown Bullhead be euthanized.

An electrofishing day on August 21, 2023, resulted in the capture of the same three fish species, with the addition of the non-native goldfish. Fish were captured using an electrofishing boat and a fyke on April 24, 2024, included the four species caught in 2023, as well as two new species: a hybrid species of Goldfish and Common Carp, and a single Emerald Shiner, which is presumed to be the result of a bait bucket release. A total of 266 Brown Bullhead, Common Carp, Goldfish and hybrids were euthanized in 2024.

A summary of the results so far is shown in Table 6. Differences in years could be explained by the timing of sampling (April vs. August) and the use of nets (e.g., fewer fathead minnows in 2023 when nets were not used).

Table 6: Fish Species Collected from Swan Lake

Date	Fish Species	Number of Fish
April 2021 (3 days electrofishing + 2 days nets)	Brown Bullhead	210
	Common Carp	7
	Fathead Minnow	>10,000
August 2022 (1 day electrofishing, 1 day nets)	Brown Bullhead	80
	Common Carp	20
	Fathead Minnow	875
August 2023 (1 day electrofishing)	Brown Bullhead	84
	Common Carp	103
	Fathead Minnow	14
	Goldfish	2
April 2024 (1 day electrofishing, 1 day net)	Brown Bullhead	193
	Common Carp	1
	Fathead Minnow	1521
	Goldfish	13
	Common Carp x Goldfish	59
	Emerald Shiner	1

The TRCA has recommended that in 2025, consideration should be given to returning some/all the Brown Bullhead to the lake (rather than removal) since they are a native fish, and it is important to have some benthic feeders in a lake ecosystem.

The Long-Term Management Plan has provisions for restocking of the Lake with a variety of fish species when the water quality is amenable. It was envisaged that this would be through the OMNRF stocking program; however, as per recent communications, their inventory was low, and therefore, the City is pursuing other avenues for sourcing largemouth bass and bluegill as the species to introduce first.

5.3 Submerged Aquatic Vegetation Planting

Phase 2 of the Long-Term Plan included provisions for introducing native submerged plants in Swan Lake to help solidify the sediment and provide fish habitat.

After a review of 2022 water quality results by the City's limnologist consultant, it was determined that the introduction of submerged aquatic plants (macrophytes) should be advanced to Phase 1 of the plan so that beneficial plant communities can compete with and help mitigate algae (phytoplankton) growth. Macrophytes will increase water clarity, which, in turn, enhances their own growing conditions. Aquatic plantings will complement existing management activities.

The planting of SAVs was implemented in June 2023 by planting 1500 wild celery (*Vallisneria spiralis*) stems in four fenced areas along the north shore of the Lake as a pilot project. In 2023 plantings were targeted around 30cm -1m deep as optimal growing depths for Wild Celery. Observations in 2024 indicated better establishment at the 20 cm depth areas, with an overall low rate of establishment at about 30%, potentially due to the turbidity and fluctuating water levels. In 2024, another 1500 stems were planted in a depth of about 20 - 40 cm. Naturally growing aquatic plants were also abundant in 2024.

In 2025, existing plants will be monitored for survival and natural propagation, and a decision about further SAV planting will be made through the five-year review process.

6. Summary and Conclusions

6.1 Summary of Monitoring Results

Through the Swan Lake monitoring program, data were collected in 2024. The collected data provide insight into long-term trends in water quality and will also help determine the need for and impact of management activities on Swan Lake.

Dissolved oxygen, temperature, and water transparency were measured at two stations through bi-weekly site visits. Profiles of temperature and dissolved oxygen indicated that Swan Lake was thermally stratified in May and June. The minimum dissolved oxygen concentration required for the protection of warm water fish is 5 mg/L, which was met in the surface water, however, continuous measurements at depth indicated that DO dropped to anoxic levels for most days in the second half of August and several days mid September at depths below about 2.5 m, when there was minimal precipitation and warmer than usual weather.

pH measured at the lab was about 8, indicative of alkaline conditions due in part to high levels of photosynthesis by algae.

Transparency at the Dock station averaged at 0.76 m over the growing season and was within the interim target for Swan Lake of 0.6-0.8 m based on correlation with the phosphorus target.

Water samples were analyzed for nutrients (phosphorus and nitrogen compounds). Total phosphorus concentration in the 0.5 and 1.5m depth averaged at 12 µg/L during the growing season (June-September) and throughout the year (below the 30 µg/L threshold for eutrophic conditions). Concentrations also met the interim PWQO for total phosphorus of 20 µg/L for lakes to avoid nuisance concentrations of algae.

Total nitrogen concentrations over the growing season averaged about 0.52 mg/L (below the 0.65 mg/L threshold for a eutrophic condition).

Chloride concentrations in the Lake were within the interim target range specified for the Lake (between 190 and 250 mg/L compared to 400-500 mg/L), and were considerably lower than 2021 values, continuing the prior declining trend.

Chloride concentrations were also measured in stormwater runoff to the ponds and the Lake (from ponds, OGS's, and overland flow) during snow melt and spring freshet. The data will be used to update the chloride balance and determine the relative contribution of each source to chloride concentration in Swan Lake.

In 2024, limited surface scum was found at both the Dock and Bridge sampling sites; however, the Lake was dominated by phytoplankton. Samples analyzed for cyanobacteria indicated lower total counts than 2023. Chlorophyll-a concentrations were below the hyper-eutrophic concentration.

The water level at the logger location changed from a maximum of 208.45 m in August to 207.26 in November.

6.2 Summary of Management Activities

In 2024, geese management was completed through nest depredation and adult goose mitigation using laser light, avian distress call and limited strategic zinc crackler pyro, as well as geese relocation in the spring. These mitigation measures effectively reduced the number of geese present at different times of

the day further down from 2023 counts. Any impact that strobe lights might have had on the geese count is not readily evident.

A second application of PAC was completed in June to reduce phosphorus concentrations and ultimately algal production in the Lake.

Fish management and the removal of bottom-dwelling fish was completed by the TRCA, and 193 Brown Bullhead and 73 Common Carp/Goldfish were captured and euthanized. Any Fathead Minnow captured were released back to the Lake.

Further planting of SAV in fenced areas along the north shore of the Lake was completed to improve water clarity.

6.3 Conclusions

Based on the measured nutrient concentrations in 2024, Swan Lake is classified as mesotrophic for total phosphorus (as well as nitrogen, but not for transparency; see Table 1 for definitions). Figure 16 provides a summary of phosphorus concentrations since 2010.

Overall, the management activities in 2021-2024 that focused on the significant nutrient loadings identified in the water quality management plan (i.e., chemical treatment and fish management to reduce internal loads and geese management to reduce external loads), were effective at improving water quality in the Lake as shown by reduced phosphorus concentrations, improved dissolved oxygen levels, and lower densities of cyanobacteria. These improvements represent a positive step towards improving the aquatic habitat in the lake and meeting the long-term water quality goals.

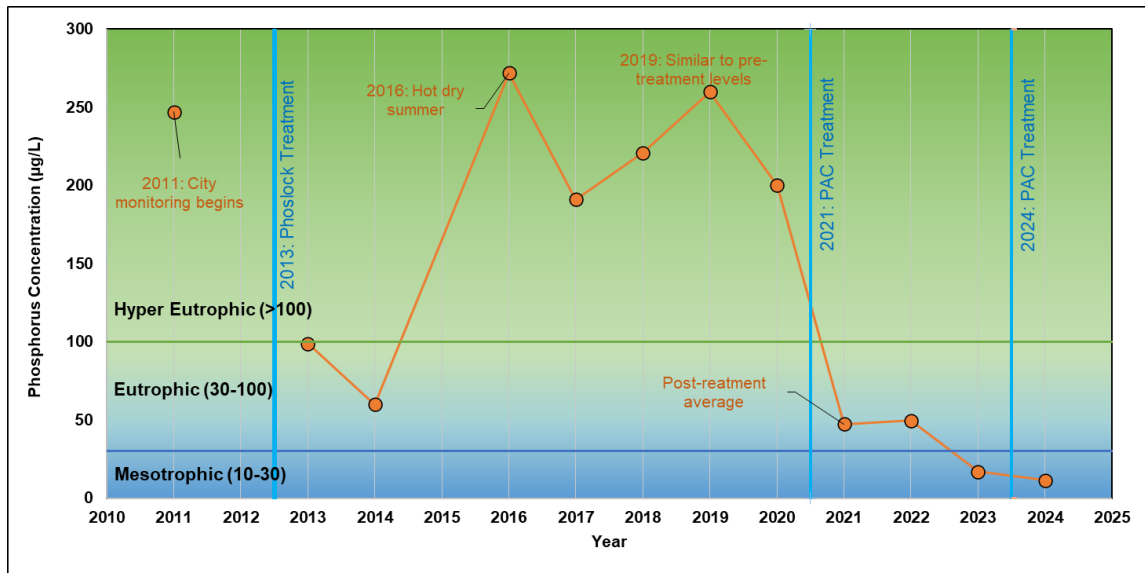
In recent years, chloride levels decreased considerably, likely due to clearing the blockage at the East Pond inlet and the Swan Club OGS, which resulted in lower untreated catchment flows to the Lake.

While internal and external source controls successfully reduced nutrient concentrations, the Lake was dominated by phytoplankton, and water clarity did not improve. This could be partly due to the absence of SAV, which has been replaced by phytoplankton (algae) due to low water clarity. The planting of SAVs stated in June 2023 and continued in 2024 to help improve water clarity.

The 2025 monitoring program will follow the recommendation of the Long-Term Management Report.

Additional measures that will be implemented in 2025 include a review of the Phase 1 of the Long-Term Management Plan, evaluation of SAV planting outcome, fish stocking, and evaluation of cost and feasibility of treatment options to reduce chloride concentration, and research into using biochar for chloride removal. A new pilot project is being considered to apply ultrasound technology for algae control.

Figure 16: Trophic State Classification for Swan Lake based on Phosphorus Concentration



Appendix A : Swan Lake Water Quality Inspection Forms

Appendix B : Certificates of Analysis

Appendix C : Canada Goose Management Program 2024- Summary Report



Consolidated Report Swan Lake Flow Diversion Assessment

City of Markham

Project Number: 60721132

May 2025

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2	2025-05-02	Menghui Xu	Addressed City of Markham's comments

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- Appendix A.** Hydrograph Comparison
- Appendix B.** Mintleaf Gate Downspout Inspection Results
- Appendix C.** Cost Estimate Sheet

1. Introduction

Swan Lake, located at the intersection of 16th Avenue and Williamson Road in the City of Markham (the City), has approximately 5.5 hectares of open water surface area with a maximum depth of 4.5 metres. Over the past several years, numerous studies have identified water quality issues in Swan Lake, including elevated chloride and phosphorus levels, the occurrence of algal blooms, and reduced dissolved oxygen levels. In response, the City launched a Long-Term Management Plan in 2021 (Markham, 2021) which outlines a phased adaptive strategy to improve the lake's ecological sustainability. For chlorides specifically, the Plan prioritizes source control to reduce chloride levels within Swan Lake, which have already contributed to significant improvements in the water quality of Swan Lake. As a part of the Long-Term Management plan, the City is exploring potential mitigation options which may include redirection of some urban stormwater runoff from the Lake to the local storm sewer system.

To support the City's stormwater diversion evaluation, AECOM has been retained to develop a one-dimensional dual drainage hydraulic model using InfoWorks ICM to assess the feasibility of diverting stormwater runoff from Swan Lake without increasing the flood risk within the study area or locations downstream. The City has proposed that the following flow diversion alternatives be assessed and modelled as part of this assignment:

1. Existing conditions (combine the existing Swan Lake catchment Infoworks model with the downstream area Markham Village and Unionville models).
2. Redirecting minor system flow from the AMICA oil grit separator (OGS) and Swan Lake Blvd. OGS units to the 16th Ave. sewers.
3. Redirecting minor system flow from AMICA OGS and Swan Lake Blvd. OGS units to the Lake outlet.
4. Redirecting the "first flush" portion of the minor system flow from the AMICA OGS and Swan Lake Blvd. OGS units to the 16th Ave. sewer (i.e., redirect the most pollutant-laden runoff in a small diversion sewer).
5. Redirecting minor system flow from Swan Club OGS to the North Pond.
6. Adjusting the flow splitter weir for the East Pond and North Pond to reduce flow bypass to the Lake.
7. Expanding the storage capacity in the East Pond and North Pond to reduce flow bypass to the Lake (to consider if the flow redirection scenarios increase flood risk).
8. Creating underground storage capacity to attenuate the flows from AMICA OGS and Swan Lake Blvd. OGS before they enter the local sewer system (consider if there is a feasible candidate site and if the redirecting scenarios increase flood risk).
9. Redirecting/pumping flows from some foundation drain collectors (FDCs) toward Swan Lake (i.e., supply potentially cleaner, cool groundwater to the Lake).

The primary goal of the project is to develop a dual drainage model to estimate annual flow volume that can be diverted under each scenario, and to identify a preferred scenario for diverting runoff away from Swan Lake without increasing flood risk in the study area and locations downstream.

The main objectives of this project are:

- Develop a dual drainage hydrologic and hydraulic model to represent the integrated storm and overland drainage systems, and to calibrate and validate the model using available flow monitoring data;
- Assess the performance of storm and overland systems under existing conditions;
- Develop models for eight City-proposed scenarios which divert flow away from Swan Lake;

- Assess the annual volume reduction and downstream impact for these scenarios; and
- Evaluate these scenarios based on the cost of implementation, anticipated downstream impacts, annualized reduced flow volume, and presumed reduction in chloride loading to Swan Lake, estimated using winter runoff volumes as a proxy

This report provides a summary of the methods used and assumptions applied in the development of the Swan Lake dual drainage hydraulic model. The model was built using InfoWorks ICM Version 2021.1, following the procedures specified in the City of Markham Stormwater Modelling Guidelines Version 1 (Cole Engineering, 2020).

2. Background Information

2.1 Study Area

As shown in **Figure 1**, the study area covers approximately 148 hectares, and includes residential, commercial, and parkland land uses. The study area consists of the catchment areas of Swan Lake, City Pond 103, and the area south of 16th Avenue that drains directly to Mount Joy Creek. Most streets in the study area are serviced by conventional storm sewers.

The Swan Lake catchment area is 42.9 ha (excluding the lake itself). In the Swan Lake catchment, runoff is collected by local storm sewers and conveyed to the North and East Ponds. Low intensity rainfall events drain directly into these ponds, while high intensity events are diverted to Swan Lake when flow levels are high enough to spill over weirs located at the inlet of each pond. The outflow from both Swan Lake and the East Pond is then collected by downstream storm sewers on Lakeside Vista Way, Lehman Crescent, Larkin Street, and Fincham Avenue, ultimately discharging to the creek. The outflow from the North Pond is received by the 375 mm sewers on Williamson Rd, which eventually discharge to City Pond 102.

The major industrial, commercial, and institutional (ICI) zone in the study area is along the south shoreline of Swan Lake. Runoff in this area is pre-treated by three OGS units before draining to Swan Lake.

The area to the south of 16th Avenue was included in the City's existing hydraulic model developed for the Markham Village and Unionville Flood Control Study (RVA, 2021). The Swan Lake hydraulic model focuses on the catchment areas of Swan Lake and City Pond 103. The completed model was then integrated with the City's existing downstream model to assess the overall downstream impact of flow diversion options.



Figure 1: Study Area

2.2 Data Collection and Review

2.2.1 Storm Network

The storm network shapefiles were provided by the City in geodatabase format. The provided dataset consists of all stormwater assets within the City. AECOM filtered these assets to focus on the study area for a more detailed review and integration into the model.

For modeling purposes, the necessary asset features include asset ID, pipe upstream and downstream invert elevation, pipe upstream and downstream asset ID, pipe material, maintenance hole lid elevation, maintenance hole diameter, maintenance hole depth, catch basin grate type, and roof downspout connection. The provided shapefiles were generally reliable and contain corresponding attributes for most of the required features.

A comprehensive review of the GIS data was undertaken to gain an understanding of the storm system and to identify any data gaps. **Table 1** outlines a summary of the storm features in the study area, and **Section 3** highlights the identified data gaps.

Table 1: Storm Network

Data	Received Data	Data Source	Format	Quantity in Study Area (units)	Notes
Storm Maintenance Hole	12-11-2023	City of Markham	GIS shapefile	444 (-)	<ul style="list-style-type: none"> ■ Depth available ("Depth" attribute) ■ Diameter available ("Width" attribute) ■ Lid elevation NOT available ■ FDC maintenance holes are included in the shapefile of Storm maintenance hole.
Storm Sewer	12-11-2023	City of Markham	GIS shapefile	20.2 (km)	<ul style="list-style-type: none"> ■ Inverts level available ("UPELEV" and "DOWNELEV" attributes) ■ Corresponding maintenance hole data available ("UPSTREAM" and "DOWNSTREAM" attributes) ■ Material available ("Material" attributes)
Storm OGS	12-11-2023	City of Markham	GIS shapefile	6 (-)	<ul style="list-style-type: none"> ■ Model number available ("Model Number" attribute)
Storm Catch Basin	12-11-2023	City of Markham	GIS shapefile	543 (-)	<ul style="list-style-type: none"> ■ Catch basin type available - included ditch inlets, rear yard and private catch basins. ■ Grate type NOT available
Roof Downspout Survey	12-11-2023	City of Markham	GIS shapefile	0 (-)	<ul style="list-style-type: none"> ■ Downspout survey did not include the area north of 16th Avenue.
Flow Monitoring	12-11-2023	City of Markham	Spreadsheet, Shapefiles	6 locations	<ul style="list-style-type: none"> ■ Time Interval – 11/1/2022 to 10/31/2023 ■ 3 Storm Locations – MH # M718W (2.5 ha), S304 (Pond 104 inflow, 11.6 ha), Y030 (Pond 104 outflow) ■ 1 FDC locations - MH # J689 (Tributary includes 69 lots, along Swan Park Rd) ■ 2 Mixed locations MH # M718N, F973

2.2.2 Stormwater Management Facilities

There were three wet ponds in the study area, and all of them discharged to downstream storm sewers. For modeling purposes, these wet ponds that discharged into storm sewers required the incorporation of a stage-

storage relation into the model. Stormwater management reports and available drawings were reviewed to identify these features of ponds and any data gaps, as summarized in **Table 2**. There is no missing data required for these three ponds and Swan Lake.

Table 2: Storm Ponds

City Asset ID	Designed Drainage Area (ha)	Volume (m ³)		Normal Water Level (m)	Stage-Storage Available in Report	Outlet Structures
		Permanent	Active			
Pond 103	48.7	12635	45334	206	Yes	<ul style="list-style-type: none"> ■ 280 mm orifice at 206.47 m ■ 2 DICBs at 207.1 m
Pond 104	12.6	1558	810	208.3	Yes	<ul style="list-style-type: none"> ■ 100 mm orifice at 208.3 m
Pond 105	19.3	2051	1096	208.3	Yes	<ul style="list-style-type: none"> ■ 66 mm Orifice at 206.8 m
Swan Lake	42.9	62,640 (at normal depth)	99,380 (at maximum depth)	208.3	Yes	<ul style="list-style-type: none"> ■ 1.3 m (crest length) weir at 208.3 m discharge to 165 mm orifice at 207 m

2.3 Background Reports

2.3.1 City of Markham Stormwater Modelling Guideline (Cole Engineering Group Limited, 2020)

This report outlined the best practices for storm system hydraulic modeling, covering asset naming conventions, catchment discretization, runoff routing, high point and sag point identification, data requirements, hydraulic model parameterization, and model validation procedures. Guideline values and procedures for the following items, as outlined in the report, were reviewed and applied in order to develop the InfoWorks hydraulic model for the current study:

- Manning's roughness coefficients for pipes and roads.
- Standard conduit shapes for streets.
- Catchment geometries.
- Catchment hydrologic properties for each type of land use.
- Catch basin rating curve for different types of grates.
- Flag and naming format.
- Model reporting format.

2.3.2 Markham Village & Unionville Flood Remediation Plan, and the Correlated InfoWorks Model (RVA, 2021)

Three significant storm events occurred in the City of Markham between June and July 2017, leading to 350 flood reports. In response, the city recognized the need to assess and mitigate flood risks in the Markham Village and Unionville areas. The purpose of this study was to evaluate the existing stormwater drainage system in these areas and develop a comprehensive plan for implementation.

An InfoWorks model was developed for Markham Village area, and this model will be used as the base model for the current study; the current study will extend this model to include the study area north of 16th Avenue. Several system deficiencies were identified in this study, including surcharging and overflowing to ground level during smaller, 2-year storm events; failure to meet current level of service criteria; and an elevated risk of street ponding.

The proposed final flood remediation plan encompasses system upgrade recommendations, risk priorities, financial planning, and regulatory approvals for implementation.

AECOM reviewed the existing InfoWorks model following standard model review procedures outlined in the City of Markham Stormwater Modelling Guidelines (Cole Engineering Group Limited, 2020). The model parameters were determined to be accurate and aligned with Guideline values. Sub-catchment areas are derived from property parcel fabrics and are verified to be appropriately characterized to represent each type of runoff surface in the study area. Although the model was not calibrated due to a lack of site-specific data, the model results were compared with the storm event on July 16th, 2019. The predicted problematic areas were generally consistent with the recorded flooding locations along Church Street, but the predicted flooding locations were fewer than recorded in other areas.

2.3.3 Swan Lake Long-Term Management Plan (City of Markham, 2021)

This study outlined issues, opportunities, and a strategy for improving the water quality of Swan Lake in Markham. The report analyzed the current state of the lake, identifying issues such as high phosphorus levels, geese-related nutrient inputs, and elevated chloride concentrations. The report presented a phased approach with core measures for the first five years, including continued water quality monitoring, enhanced geese management, and the use of chemical treatments. Complementary measures, such as fish management plans and the installation of shoreline plantings, were introduced in the second phase, while the third phase considered adapted core measures and potential alternative strategies, such as investigating groundwater contributions and stormwater redirection. The 25-year plan aimed to achieve a low eutrophic condition in the lake, improve water clarity, and reduce algal bloom frequency.

The water balance study outlined in the report and the correlated PCSWMM model provide insights into the flow contribution from ponds and oil grit separators (OGSSs) to Swan Lake, as well as the hydrologic characteristics of Swan Lake catchments critical to the current study.

2.4 Evaluation Criteria

Criteria outlined in the City of Markham Stormwater Modelling Guidelines (Cole Engineering Group Limited, 2020) are summarized as follows. These criteria were followed to evaluate existing system and feasibility of diversion scenarios:

Storm Sewers

- Surcharge state 1 – No surcharge will be considered as low risk.
- Surcharge state 2- The pipe is surcharged, but the slope of the HGL is flatter than the pipe slope (i.e. it is surcharged due to downstream conditions), which will be considered as moderate risk.
- Surcharge state 3- The pipe is surcharged, and the slope of the HGL is steeper than the pipe slope (i.e. the surcharge is at least in part caused by the pipe capacity), which will be considered as high risk.

Storm Maintenance Hole

- Maximum HGL is greater than 2.0 m below ground elevation will be considered as low risk.
- Maximum HGL is within 2.0 m of ground elevation will be considered as moderate risk.
- Maximum HGL exceeds ground elevation will be considered as high risk.

Overland

- Overland flow depth lower than 150 mm will be considered as low risk.
- Overland flow depth between 150 mm and 300 mm will be considered as moderate risk.
- Overland flow depth exceeds 300 mm will be considered as high risk.

2.5 Data Gaps

A background review of City-provided data was completed prior to undertaking model updates, and several information gaps were identified. Several data gaps were identified:

a) Storm Sewer Invert Elevations (Including FDC pipes)

Catch basins are marked On Roy Grove Way and Town Villa Way without corresponding pipes and maintenance holes. As shown in the **Figure 2**.

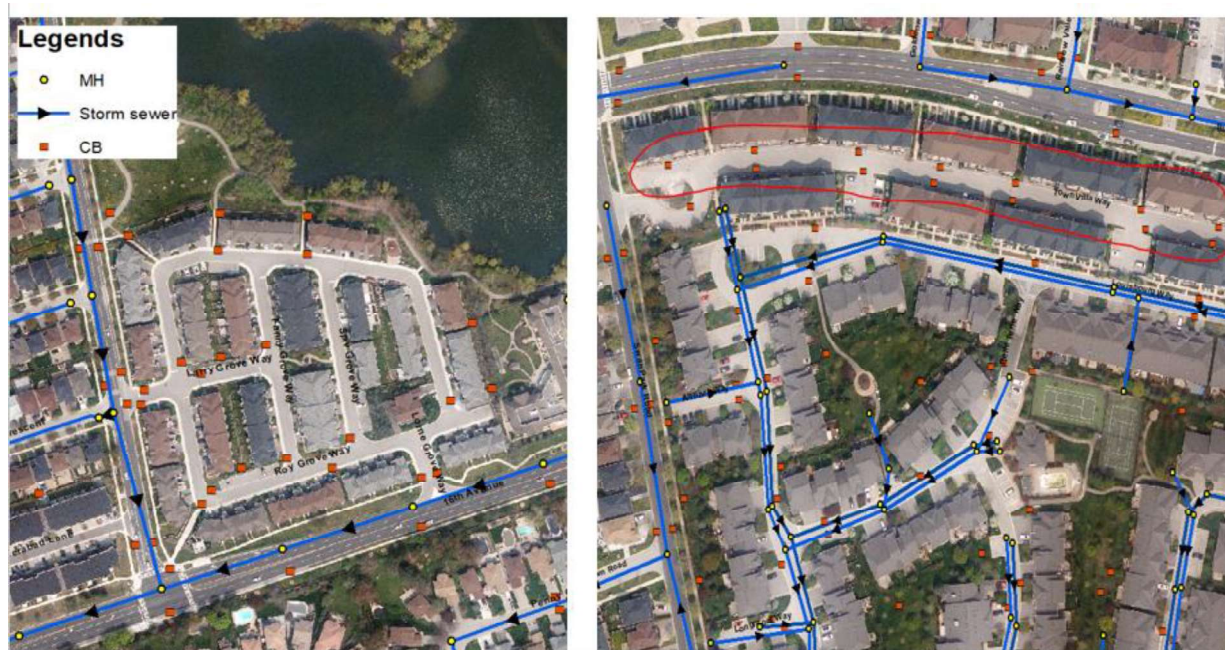


Figure 2: Catch Basin with out Corresponding Pipes

Data gaps in storm sewer invert elevation are listed as follows and shown in **Figure 3**.

- 6 of 467 (1.2%) pipe sections did not have upstream elevations but have downstream elevations.
- 17 of 467 (3.6%) pipe sections did not have downstream elevation but have upstream elevations.
- 129 of 467 (27.6%) pipe sections did not have elevation at both ends.

Note that the pipes with missing elevations with both ends are primarily in the private development area (see sketch below).



Figure 3: Pipe Missing Elevation with Both Ends

While most of the missing inverts were covered in the provided scanned as-built drawings, some of the data in the drawing were not readable due to scanning and resolution issues. As advised by the City, invert interpolations were applied to infer the missing elevation from the available upstream or downstream pipe invert elevations, or slope data available from either the drawing or GIS records. After the initial screening, eight pipe sections listed in **Table 3** were identified as not being available in any of the provided drawings, and located at the most upstream point of a sewer branch. Diameter data are also missing for these pipes.

Table 3: Pipes that Require Invert Survey

Asset ID	Upstream MH ID	Downstream MH ID
Q693Q692	Q693	Q692
Q688Q687	Q688	Q687
Q691Q690	Q691	Q690
Q692Q691	Q692	Q691
Q689Q688	Q689	Q688
Q694Q693	Q694	Q693
Q695Q693	Q695	Q693
Q690Q687	Q690	Q687

Field work was completed to measure the invert levels for these pipes and the collected data was incorporated when developing the Infoworks hydraulic model.

b) Storm Sewer Diameter

15 of 476 (3.2%) pipe sections did not have diameter information, diameters for 7 pipe sections were filled were identified from the provided site servicing plan drawings. Assumptions would not be accurate for the eight pipes listed in **Table 3**, as they are located at the most upstream of a sewer branch, and there is no available drawing for them. Field surveys were completed to measure the diameter for these pipes and the collected data was incorporated when developing the Infoworks hydraulic model.

c) Catch Basin Leads

The catch basin lead layer was not available in the provided geodatabase. AECOM has reviewed the provided drawings to identify catch basin downstream connections. Catch basins which are not included in any of the drawings were assumed to connect to their closest stormwater maintenance holes.

d) Catch Basin Grate Type

Only ditch inlets in catch basin layer were differentiated in the available attribute tables; grate opening numbers and types were not identified for right-of-way and rear yard catch basins. AECOM reviewed street view on Google to identify the type of grate for these catch basins. For areas where street view was not available, AECOM assumed all right-of-way and rear yard catch basins have single grid grates, and catch basins located in sag (depression) areas will be assigned twin herringbone grates, as per instruction provided by the City. This assumption were verified by the field visit. A field survey was conducted to check the grate types of catch basins visible from the right-of-way. A total of 143 catch basins were inspected. Except for 16 rear yard catch basins that have beehive grates for ditch inlets, the remaining 127 catch basins have herringbone grates, with double inlets located at the identified sag locations, which were consistent with the initial assumption.

e) Maintenance Hole Lid (Rim) Elevation

Lid elevations were not available for most the maintenance hole in the study area attributes; the DEM were used to collect lid elevations for maintenance holes to ensure consistency.

f) Maintenance Hole Chamber Diameter

379 of 470 (81%) maintenance holes did not have diameters. An assumption was made that the maintenance hole diameter is 600 mm larger than the largest pipe diameter, with a minimum diameter of 1200 mm.

g) Roof Connection

A downspout survey was provided but did not include the area north of 16th Avenue. Therefore, a visual survey of downspout connectivity was conducted from the right-of-way (ROW) for properties within the Swan Lake catchment area. The results are discussed in Section 3.3.2 of this report.

3. Model Development

The dual drainage model was created with InfoWorks ICM software. Dual drainage represents the surface (major) and underground (minor) flow systems as an interconnected network. Subcatchments were discretized from maintenance hole to maintenance hole. Major overland flow is conveyed to the minor system through “gullies” in the model, representing catchbasins. The details of the dual drainage model are explained in the following sections.

3.1 Minor System

Minor system assets included in the Swan Lake InfoWorks model are shown in **Figure 4** and listed in **Table 4**. The outfalls at manholes F973 and G401 were converted to storm nodes when the Swan Lake hydraulic model was integrated with the existing downstream model.

Table 4: Minor System Assets in the Model

Item	Quantity
Storm Nodes	444
Storm Conduits	446 sections, total length approximately 20.2 km (Including lengths of FDC pipes)
Flow Control Structures	<p>Three flow splitter weirs:</p> <ul style="list-style-type: none"> ■ North Pond flow splitter weir ■ East Pond flow splitter weirs at both two sewer inlets ■ Swan Lake outlet control weir <p>Five orifices:</p> <ul style="list-style-type: none"> ■ 100 mm orifice plate at North Pond outlet ■ 66 mm orifice plate at East Pond outlet ■ 165 mm orifice at Swan Lake outlet ■ Two 100 mm orifice in the pipes on Swan Lake Blvd. to control the outflow from ICI area
Outfalls	<p>Three outfalls:</p> <ul style="list-style-type: none"> ■ Outlet to Markham Village Area at Manhole F973 ■ Outlet to Markham Village Area at Manhole G401 ■ North Pond outlet to downstream system at manhole J681
Storage Nodes	<p>Four major storage nodes:</p> <ul style="list-style-type: none"> ■ Swan Lake ■ East Pond ■ North Pond ■ Pond 103



Figure 4: Swan Lake Model Extent

3.1.1 Data Source

The primary source of information for model development is the geodatabase provided by the City, which contains GIS data for existing storm sewers, property parcels, storm manholes, catch basins, ponds and the LiDAR derived digital terrain model (DTM). Data gaps, including incomplete and inconsistent information, were identified in the previous technical memo (TM#1, Background Review). A comprehensive data validation was then performed in InfoWorks using the built-in engineering validation and tracing tools to identify connectivity errors. These data gaps and inconsistencies were resolved primarily using the as-built documents provided by the City. As discussed in the previous section, eight (8) pipe sections in the residential area between Chancery Road and Augusta Drive, as shown in **Figure 5**, are not included in any provided drawings. Since these pipes are the most upstream sections of a sewer branch, their invert levels can not be reasonably assumed. Therefore, fieldwork was conducted to collect dimensional and invert data for these assets.

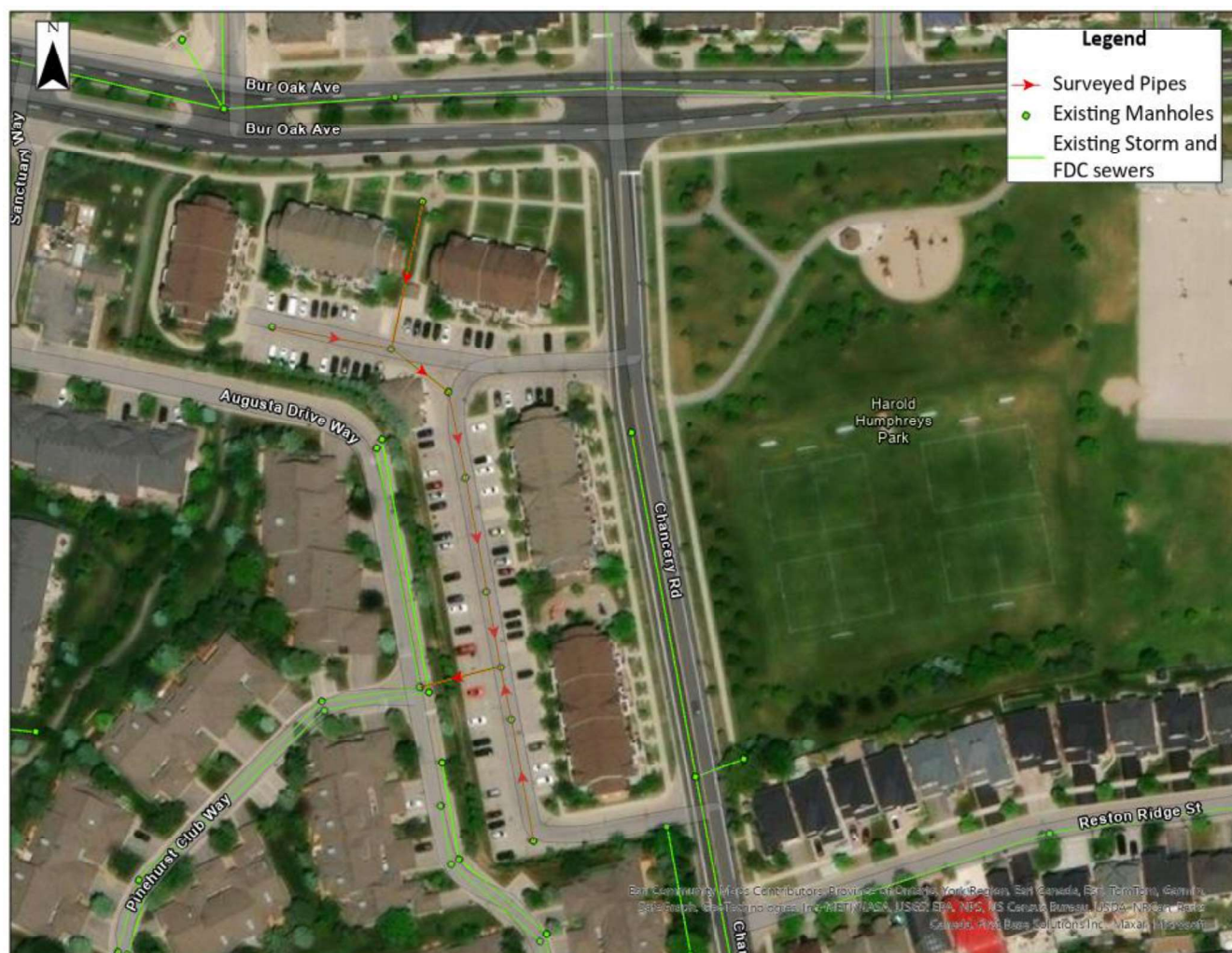


Figure 5: Surveyed Pipes

3.1.2 Manholes

There were 198 manholes in the study area that did not have elevation data available in the provided geodatabase. For the manholes with available elevation in the provided GIS data, a comparison was made between the available data and the LiDAR obtained manhole lid elevation. This comparison shows that the difference between the GIS data and the LiDAR data was generally within 0.2 m. However, in some areas, elevation discrepancies ranged from ± 0.5 m to ± 2 m, affecting 11% of the total number of manholes. Since the GIS data originated from multiple sources, including as-built drawings, design drawings, inspection reports, historical surveys, and Google Earth, the lid elevations for all manholes were extracted from the LiDAR data to ensure consistency. The available elevations in the GIS data were not used for model development. For pipes that had their invert levels collected from field survey, the surveyor measured the depth from the top of the manholes to the pipe invert. The invert elevation was then calculated by subtracting the measured depth from the LiDAR elevation at those manholes.

3.1.3 Storm Sewers

The properties of storm sewers (conduits), including upstream and downstream invert levels, diameters, and pipe material, were obtained from the data sources outlined in the previous sections. A Manning's n value of 0.013 was applied for all types of concrete and PVC sewers.

3.1.4 Control Structure

A detailed review of all provided as-built drawings and site control plans was conducted to identify all flow control structures in the study area. Overflow weirs and orifice plates listed in **Table 4** were identified from the available drawings and incorporated into the InfoWorks model accordingly.

3.1.5 Catch Basins

In InfoWorks, defining a node as a “gully” enables it to perform like a catch basin, connecting the storm and overland systems. Each “gully” node in the model was assigned the specific number of catch basin within the subcatchment, along with a flow rating curve that allows the gully to represent the actual flow rate that can enter the minor system. The attribute “number of inlets” represents the corresponding number of catch basins connected to each manhole, and “gully type” represents the depth-discharge relation specific to each catch basin grate type. This set-up enables flow accumulated on the overland surface to enter the storm sewer system at a specific flow rate that correlated to the depth defined by a depth-discharge curve (rating curve). Conversely, stormwater in the collection system can surcharge to the overland network through gully nodes

The quantity and location of catch basins were obtained from the City-provided GIS data, and the catch basin inlet grate type for each inlet was gathered from field observations and Google Streetview. Due to the absence of service line connection data, catch basins were assigned to their closest manhole. The survey results show that the study area is primarily serviced by herringbone catch basin inlet grate types. The specified head-discharge curve developed by City of Ottawa, which accounts for on-sag and on-grade inlets, were incorporated into the hydraulic model.

3.1.6 Rear Yard Catch Basins

The number and location of rear yard catch basins were obtained from the City-provided GIS geodatabase. The field survey checked the grate types of rear yard catch basins that are visible from the right-of-way and identified them as either Ditch Inlets (DICB) or herringbone catch basins. Since rear yards in the gated community are primarily located in sag areas with no overland outlet, and the rating curve of a herringbone catch basin in sag areas is identical to that of a DICB, each rear yard catch basin was modeled as a gully node and assigned a storage-discharge curve specific to herringbone catch basins situated in sag areas. The provided subdivision plan drawings were used to identify the connection for rear yard catch basins. For rear yard catch basins that tee into a conduit without a manhole, a dummy node was implemented to represent the connection junctions. In cases where rear yard catch basins were not available in the provided drawings, lead pipes were assumed to be circular, with a diameter of 250 mm and a 1% slope, and were connected to the nearest appropriate stormwater manhole as determined by engineering judgment.

3.1.7 Storage Nodes

In Infoworks models, storage units, including stormwater management ponds, natural water bodies and major sag location, are represented by storage nodes. At each storage node, a stage-area relation is required to represent the storage volume at various elevations. Storage nodes were placed at Swan Lake, the North pond, East pond and City pond # 103. The stage-area relation for the North Pond, East Pond and Swan Lake were obtained from the City provided PCSWMM water balance model. The stage-area relationship for Pond 103 was gathered from Appendix C of the Stormwater Management Design Brief: Pond A, Pond E, and Aviva Pond (Revised by Stantec, 2004). The stage-area relationships applied in the Swan Lake InfoWorks model are summarized in **Table 5**.

Table 5: Stage-Storage Relation in the model

Stage (m)	Area (m ²)	Average Area (m ²)	Volume (m ³)	Storage (m ³)	Active Storage (m ³)
East Pond					
205.8	100	0	0	0	0
206.8	521	311	311	311	0
207.8	1339	930	930	1241	0
208	1675	1507	301	1542	0
208.3	1949	1812	544	2086	0
208.5	2150	2050	410	2495	410
208.8	2446	2298	689	3185	1099
North Pond					
205.8	20	0	0	0	0
206.8	304	162	162	162	0
207.8	1090	697	697	859	0
208.3	1500	1295	648	1507	0
208.8	1740	1620	810	2317	810
209	2200	1970	394	2711	1204
Swan Lake					
204.5	0	0	0	0	0
205	320	160	80	80	0
205.5	1880	1100	550	630	0
206	5800	3840	1920	2550	0
206.5	12000	8900	4450	7000	0
207	20000	16000	8000	15000	0
207.5	34000	27000	13500	28500	0
208	46000	40000	20000	48500	0
208.3	48267	47134	14140	62640	0
208.5	52600	50433	10087	72727	10087
209	54000	53300	26650	99377	26650
Pond 103					
204.5	6532	0	0	0	0
205	7717	7125	3562	3562	0
205.5	8856	8287	4143	7706	0
206	10902	9879	4940	12645	0
206.5	12425	11664	5832	18477	5832
207	13719	13072	6536	25013	12368
207.5	15025	14372	7186	32199	19554
208	16479	15752	7876	40075	27430
208.5	17911	17195	8598	48672	36027
209	19317	18614	9307	57979	45334

3.1.8 Naming Conventions

Most minor system assets in the Infoworks model were named using the City's Asset ID. For objects without an Asset ID, as well as dummy objects and duplicate objects, asset names were assigned in compliance with the naming conventions listed in Table 4.1 of the City of Markham Stormwater Modeling Guidelines Version 1 (Cole Engineering, 2020).

3.2 Major System

The major (overland) system in the model consists of streets with flow constrained by the curb and gutter along both sides, and rear yard channels. The following data sources were used when creating major system.

- The City-provided GIS geodatabase
- LiDAR DEM
- Aerial imagery

3.2.1 Overland Conduit and Channel Geometry

The streets were modelled as wide shallow open channel conduits with irregular cross-sectional shape to reflect the appropriate geometry, flow area and channel roughness. The overland conduit invert levels were set at the maintenance hole lid elevation such that flows can transfer between the minor and major systems if there is flooding out of the maintenance holes from the minor drainage system or when the flow is restricted into the minor system at individual catch basin inlets based on the catch basin inlet capture capacity.

An initial overland network was created by duplicating the minor system network, converting the system type to “overland” adjusting conduit invert levels to the ground elevations of their connected manholes, and reversing direction of conduits when the road slope is opposite of the pipe slope. Pipes with slope reversed to street slope are shown in **Figure 6**. Then, the overland flow path on streets was generated using the Esri ArcHydro tool based on the LiDAR data and compared against the original network. Additional overland conduits were added to the network where the minor system was not continuous. Flow splits at intersections were determined by the model based on the physical network layout topography. Local high points were identified using LiDAR data and added as flow split points, which may not follow the minor system direction.

The major system in the model was primarily defined by three types of roads in the study area: arterial road, collector and local road. Further, flow paths on rural and rear yard lands were also included, as shown in **Table 6**. The typical cross-sections for these three road types were obtained from the provided Markham Village Infoworks Model. Each cross-section was defined by unit width at unit height, then multiplied by the actual road width and height. Roadside ditches do not exist in the study area.



Figure 6: Pipes with Slope Reversed to Street Slope

Table 6: Cross-section Geometry – Arterial Road

Road Type	Unit Height	Unit Width	InfoWorks Conduit Cross-section
Arterial Road	0	0	
	0.5	0.17	
	0.75	0.17	
	1	1	
Collector Road	0	0	
	0.32	0.29	
	0.57	0.29	
	1	1	
Rear Yards	0	0	
	1	1	
	3	3	

3.2.2 Naming Conventions

Major system objects were named following the naming convention outlined in Table 4.1 and 4.2 of City of Markham Stormwater Modelling Guideline Version 1.

3.3 Subcatchments and Hydrology

The following data sources were used when creating the subcatchment areas in the model.

- LiDAR DEM (GIS)
- AECOM visual downspout inspection
- Property parcel (GIS)
- Building roof footprint

3.3.1 Catchment Delineation

Subcatchments were delineated on a manhole-to-manhole basis referring to the topography and property boundaries. Reverse driveway surveys were not conducted for this project, instead, all property front lots were assumed to be graded towards the street and confirmed with the flow path generated using ArchHydro tool. Rear yards were incorporated into the main catchment in cases where specific rear yard flow paths were absent; where flow paths were present, rear yards and the roof area draining towards rear yards were separated from the main catchment and routed to rear yard catch basins if available, or to dummy overland nodes located on flow paths.

The imperviousness of each subcatchment was initially determined by processing aerial images using the ESRI Raster Classification tool. This process involved analyzing the color spectrum of the aerial photos in GIS. The initial estimates were further refined by incorporating known impervious surfaces, such as roads and roofs. **Figure 7** shows an example of this process. Since runoff generated from impervious surfaces (roof leader, driveway, sidewalk, etc.) that drains to pervious surfaces may not be captured by stormwater catch basins, the imperviousness in the catchment areas with flow monitors was further calibrated using flow monitoring data. Roof areas were assumed to be equal to the area of the building footprint. The directly connected roof area to the minor system was calculated based on the downspout status information obtained during the field surveys. Roof downspouts directed into the ground were connected to the minor system. Subcatchments were then further adjusted into three categories based on runoff surfaces as follows:

- Main Storm Subcatchments (Named as: Asset ID_S): were established for the three runoff surfaces: impervious surface (street pavement, sidewalk, driveways, and parking lots), disconnected sloped roofs, and pervious surface (bare soil and green areas). Main subcatchments were assigned to the street gully nodes based on the topography, road grade and overall lot drainage direction.
- Connected Sloped Roofs (Named as: Asset ID_RC): were established for sloped roof areas connected directly to the storm sewer. A separate subcatchment was created from each storm subcatchment containing only directly connected sloped roofs. Total and contributing area of the dummy subcatchments were assumed to equal to the total connected sloped roof areas within the same main storm subcatchment. Connected roofs were discharged to the storm sewer system directly through a lateral. The flow discharged to the storm sewer should be limited to the capacity of the roof drainage system.
- Flat Roofs (Named as: Asset ID_FRC): were established for flat roof areas connected directly to the storm sewer. Flat roofs are normally associated with Industrial, Commercial or Institutional Land Use (ICI) or high-rise residential areas, and typically have large areas that drain to internal plumbing. To

account for the ancillary structures, the connected flat roofs were modelled as separate subcatchments and drained to a dummy node with a storage area equal to the roof area and a head-discharge curve for a flat roof downspout to control flow was used to drain rooftop flows to the storm sewer. Unless specific downspout numbers were available from the building drawings, a general assumption of 1 downspout per 160 m² of flat roof area was applied to estimate the discharge limit of flat roof drainage system (in accordance with CoT Infoworks modelling guidelines).



Figure 7: Example – Imperviousness Determination

3.3.2 Roof Connectivity

A visual survey for downspout connectivity was conducted from the right of way (ROW) for properties within the Swan Lake catchment area. Rear yard downspout connections could not be confirmed.

The results of the downspout survey are presented in **Figure 8**. From the 534 properties surveyed:

- 14 properties (2%) have downspouts connected into the ground (storm, sanitary or FDC pipes); and
- 503 properties (95%) have downspout draining to the surface, connected to the overland system; and
- 18 (3%) properties did not have downspouts visible from the ROW.



Figure 8: Roof Downspout Connectivity Survey Results

3.3.3 Roof Area Separation

As shown in **Figure 9**, flat and sloped roofs are identified using aerial imagery. The City's modeling guidelines (Cole Engineering, 2020) recommend using a "split-rainfall method," which assumes that a typical home's downspouts can capture flows from a rainfall event with up to a 5-year peak intensity, while any excess would overflow to the ground. This method involves creating three hyetographs and duplicating the connected roof catchment as overflow catchments. The hyetograph for the full storm rainfall is assigned to the general storm catchments, the hyetograph for up-to 5-year design storm is assigned to the roof catchment, and the overflow catchment is assigned the difference between the rainfall intensities for intervals where the storm rainfall exceeds the peak 5-year design storm rainfall. This method was not applied in the development of the Swan Lake hydraulic model for the following reasons:

- More than 95% of residential roofs in the study area were confirmed to be disconnected from the storm sewers
- The peak intensity of a 5-year design storm varies by event duration, making it difficult to select an appropriate duration for the 5-year design rainfall to match historical events when calibrating the model.

Large flat roofs of these commercial buildings within the study area were incorporated as separate subcatchments and routed to a dummy node with a storage area equal to the roof area and a head-discharge curve for a typical flat roof inlet. It is assumed that each 160 m² of flat roof area will be served by one inlet. The detailed runoff surface configuration for flat roofs follows Table 4.9 in the *City of Markham Stormwater Modelling Guidelines*.

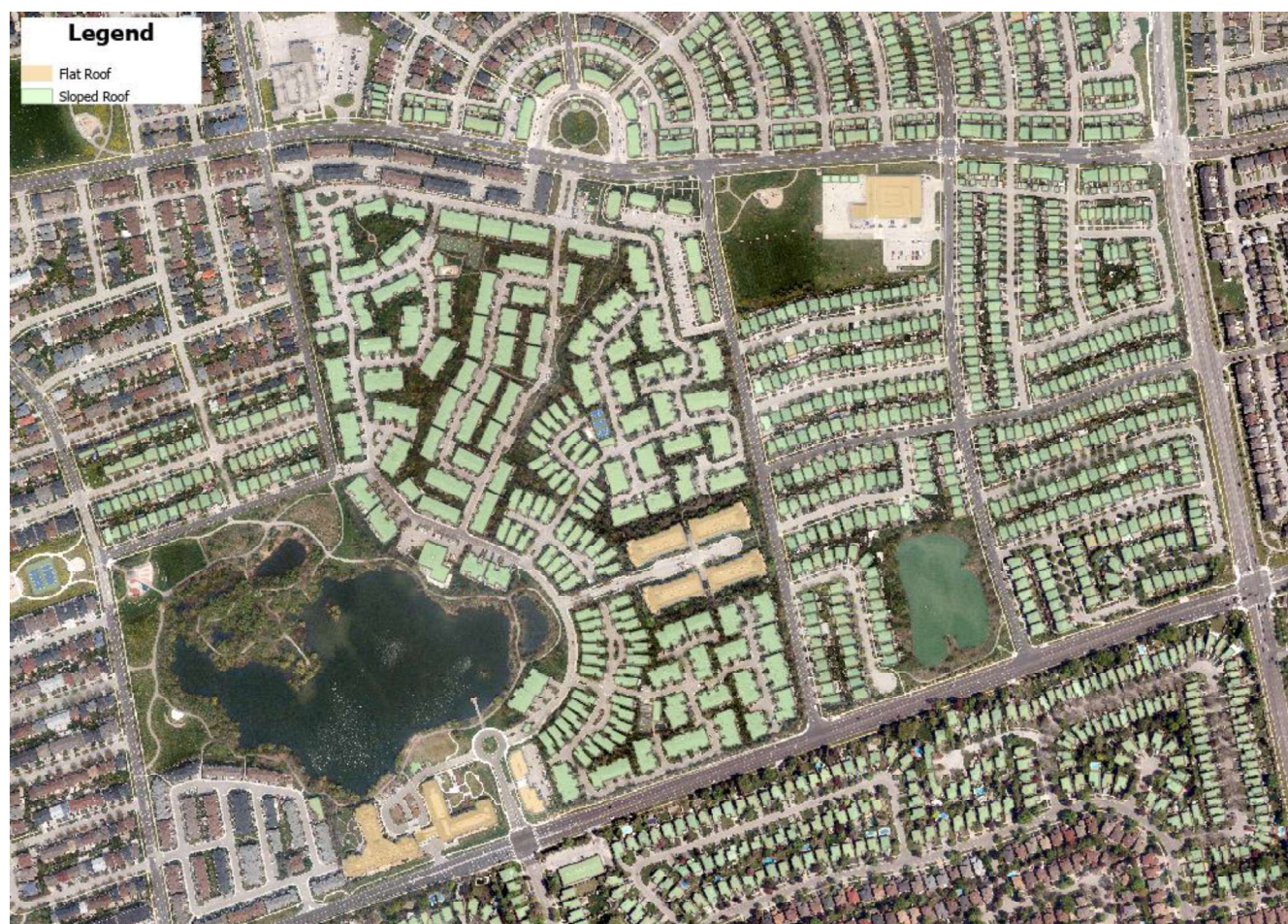


Figure 9: Flat Roofs in the Study Area

3.3.4 Hydrologic Conditions

In InfoWorks, the hydrologic characteristics of a catchment are defined by the land use, which encompasses a list of runoff surfaces. The total runoff generated from each subcatchment is quantified by summing the runoff generated by each surface during a storm event. **Table 7** presents the runoff surface IDs, descriptions, and initial guideline values for the associated hydrological parameters for these runoff surfaces. These parameters were further adjusted based on flow monitoring data during the model calibration process.

Table 7: Runoff Surfaces Hydrologic Parameters

Runoff Surface ID	10	20	30	40	50
Runoff Surface Description	General pervious area	General impervious area	Connected sloped roof area	Disconnected roof area	Flat roof area
Runoff Routing Value	0.025	0.013	0.033	0.013	0.013
Runoff Volume Type	Horton	Fixed	Fixed	Fixed	Fixed
Surface Type	Perv.	Imp.	Imp.	Imp.	Imp.
Ground Slope (m/m)	0.01	0.01	0.33	0.33	0.005
Initial Loss Value	0.005	0.01	0.001	0.001	0.001
Fixed Runoff Coefficient	-	1	1	1	1
Horton Initial Infiltration	125	-	-	-	-
Horton Limiting Infiltration	5	-	-	-	-
Horton Decay	2	-	-	-	-
Horton Recovery	1	-	-	-	-

3.3.5 Naming Conventions

Subcatchments were named following the naming conventions outlined in Table 4.3 of City of Markham Stormwater modelling Guidelines, Version 1. (Cole Engineering, 2020)

3.4 Boundary Conditions

- ◆ Outlets to the downstream system are shown in **Figure 10**. Downstream boundary conditions can significantly impact the operation of the dual drainage system, as backwater effects may constrain flow in the storm sewers when water levels are high. Water levels in storm manhole M724 and G401 were not applied in the Swan Lake hydraulic model because the model was integrated with the City's existing downstream model after completion, and these two outfalls will be converted to storm nodes in the combined model.
- ◆ Outflow from the north pond exits the study area at Manhole J689, and the flow is further conveyed to City Pond 102 via the existing 250- 350 mm storm sewer pipes, with an invert level of 206.9 m. The LiDAR DTM indicates that the current water level in City Pond 102 is at 194 m, and the spill level is at 197 m. It has been determined that the pond does not restrict the outflow from the study area. However, the relatively small pipe sizes and runoff from the surrounding residential development may cause surcharge in the manhole, potentially limiting outflow from the study area. Further investigation is required to assess whether this could constrain outflow from the study area. Currently, manhole MH689 is modeled as a free flow outfall.
- ◆ Study area outflow through manhole M724 and G401 drains to the creek outlet by 1050-1800 mm sewers, spanning approximately 1.7 km along Lehman Crescent, Larkin Avenue, and Heisey Drive, as shown in **Figure 10**. This reach was identified as having insufficient capacity in the previous Markham Village and Unionville Flood Control Study (RVA, 2021). To address capacity deficiencies, a 520 m relief sewer, ranging in diameter from 1200 mm to 1800 mm, was proposed, extending from Manhole A095 to the outfall through the parkland, as shown in **Figure 11 (RCA, 2021)**. However, the previous study did not account for external flow from the Swan Lake area.

3.5 Updated Base Case

- ◆ After combining the two models, the updated model results indicate that the original proposed solution is insufficient to address all capacity constraints. The surcharge level would rise to above 1.8 m below the ground (assumed basement level) during a 100-year design storm event when flow from the Swan Lake area is included, as shown in **Figure 12**. To maintain flow within pipe capacity, and to address the capacity limit which would affect Swan Lake diversion options, additional upgrades were identified that consisted of upsizing a 790 m section of 1350 mm sewers on Larkin Avenue to 1800 mm, as shown in **Figure 13**. Please note that upsizing the entire 790 m length to 1800 mm is a conceptual scenario that removes downstream restriction at the critical locations; spacing constraints and constructability have not been assessed at this stage. During design stages, the pipe sizes may be gradually increased from upstream to downstream. The upsized pipes were incorporated into the base model and scenario models as a baseline condition, and costs of implementation for this external upgrade will not be considered as part of the implementation costs for the Swan Lake flow diversion scenarios.



Figure 10: Swan Lake Connections locations to downstream sewers

Existing Conditions Level of Service - Fincham

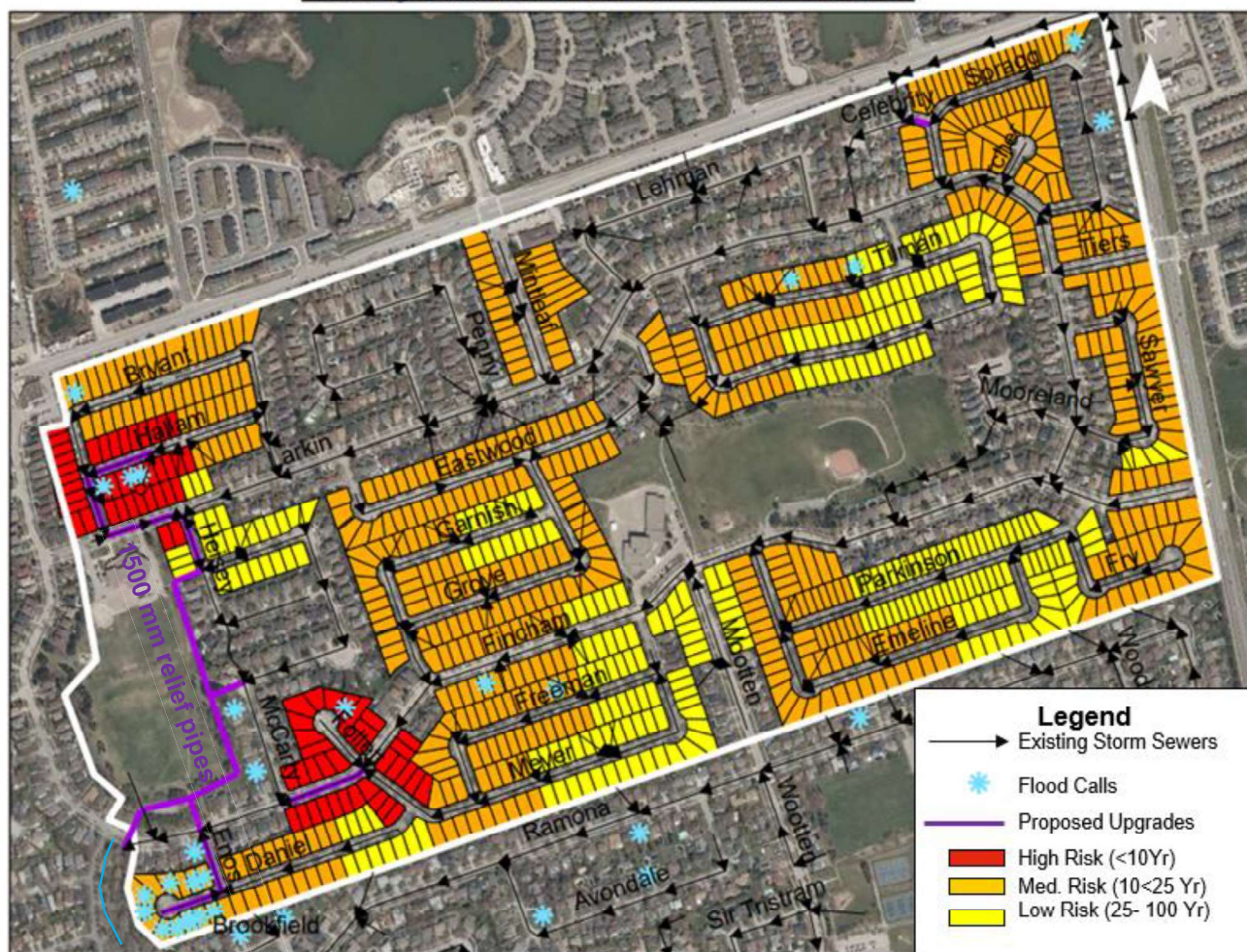


Figure 11: Proposed Solution from the Previous Study (RVA, 2021)

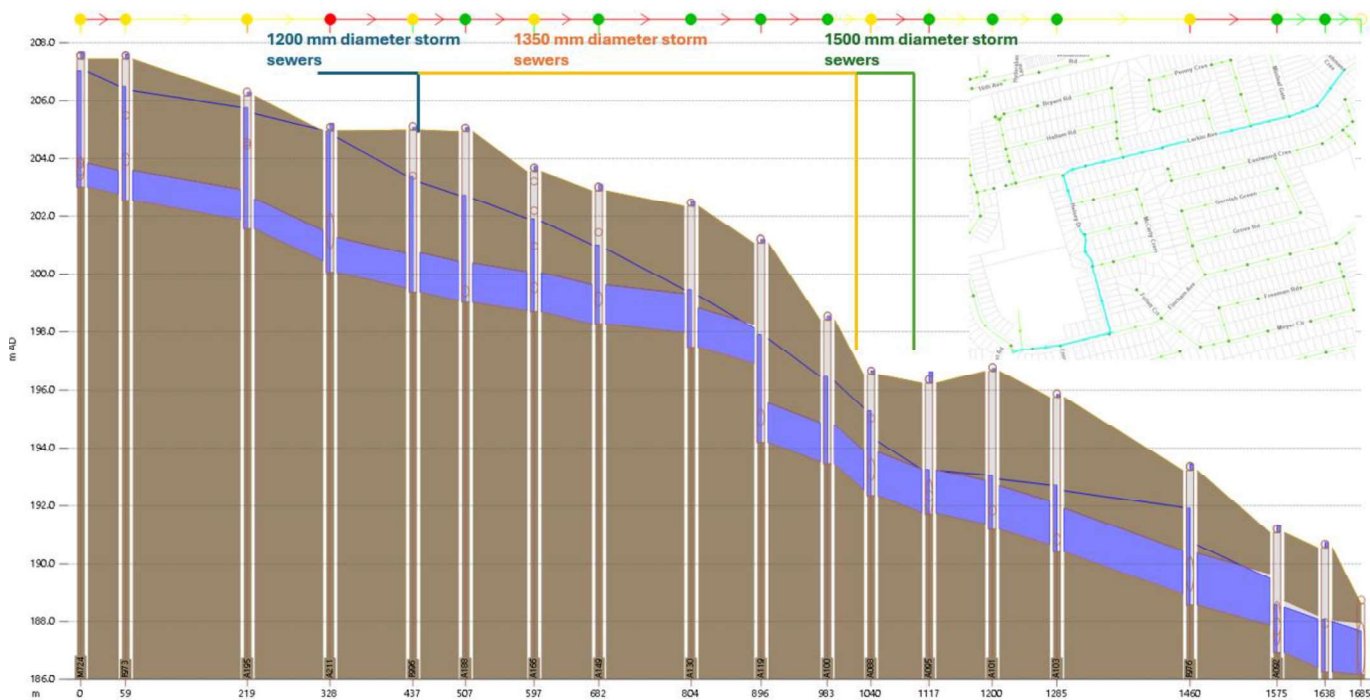


Figure 12: HGL Profile - Original Solution with Additional Flow from Swan Lake

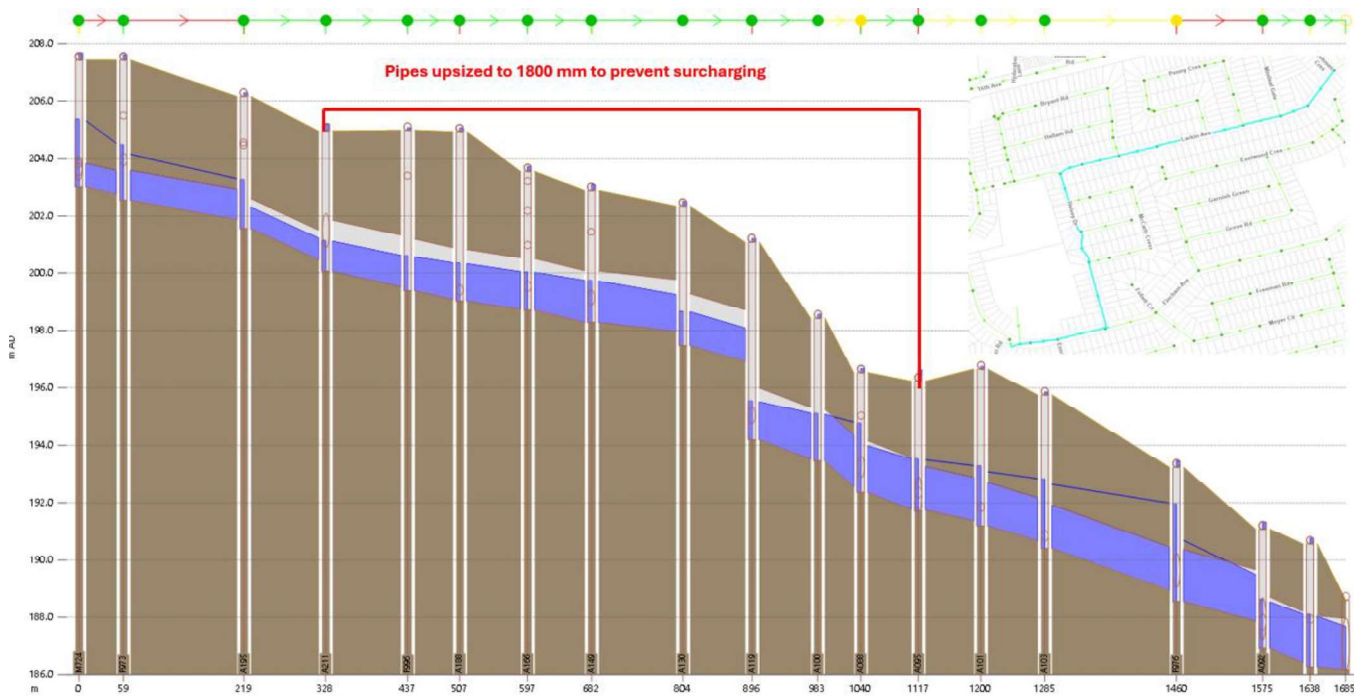


Figure 13: HGL Profile with Additional System Upgrades on Larkin Avenue

3.6 Rainfall and Flow Monitoring Data

Rainfall data from November 1, 2022, to November 4, 2023, was provided by the City for two rain gauges:

- MA-12 located in the open space at the southeast end of Monkhouse Rd, approximately 1.2 km west of Swan Lake.
- MA-20 located at Black Walnut Public School, approximately 2.5 km southeast Swan Lake.

A review of the rain gauge data was conducted to assess the quality and suitability of the recorded significant storm events for flow data analysis and model calibration. As shown in **Table 8**, twenty events with accumulated rainfall depths exceeding 10 mm were identified during the morning period.

Table 8: Summary of Rainfall Data

Event	Total Volume (mm)		Peak 5-min Intensity (mm/hr)	
	MA12	MA20	MA12	MA20
2022-11-30	24.0	26.6	9.6	14.4
2023-03-25	20.8	21.6	7.2	7.2
2023-04-01	19.6	19.4	12.0	12.0
2023-04-05	12.7	12.5	33.6	26.4
2023-04-17	12.8	12.0	9.6	9.6
2023-04-22	17.4	15.6	12.0	12.0
2023-04-29	20.8	17.6	7.2	9.6
2023-05-02	12.8	8.6	12.0	4.8
2023-05-20	33.8	34.6	26.4	24.0
2023-06-12	64.0	0.0	19.0	0.0
2023-06-24	12.6	15.0	7.2	9.6
2323-06-26	35.6	53.6	67.2	127.2
2023-07-01	12.8	13.0	55.2	86.4
2023-07-13	28.8	24.0	31.2	28.8
2023-07-16	9.6	9.6	9.6	12.0
2023-07-24	49.6	72.0	74.4	84.0
2023-08-25	11.0	19.4	31.2	50.4
2023-09-06	30.6	22.4	136.8	55.2
2023-09-12	11.2	12.6	12.0	14.0
2023-10-06	27.8	28.2	50.4	43.2

Flow monitoring in the study area was conducted by the City from November 2022 to November 2023. The locations of flow monitors in the study area are shown in Figure 14, and additional details are provided in Table 9.

Of the six flow gauges provided in the study area, the gauges at Manhole S304 and Manhole M718 (west leg) were selected for storm flow analysis, while Gauge J689 was chosen for analyzing FDC flow. Gauges in the mixed (storm and FDC) sewers were not used for analysis due to the potential uncertainty in FDC flow, which could compromise the accuracy of the storm flow analysis and vice versa.

Table 9: Flow Monitor Summary

MH ID	Street Location	System Type	Comment	Sewershed Area (ha)
J689	Swan Lake Road & Williamson Road	FDC	FDC flow from 63 buildings	12
M718 (North Leg)	39 Kingfisher Cover	Mixed	FDC flow from all 212 buildings in the gated community + Swan Lake outflow	26.4
M718 (West Leg)	39 Kingfisher Cover	Storm	Storm flow from ICI area along 16th Avenue	12
F973	38 Lehman Crescent	Mixed	Mixed	24
S304	18th Swan Park Road	Storm	North pond inlet	84
Y30	East to the water park	Storm	North pond outflow	43.2



Figure 14: Locations of Flow and Rainfall Monitors

3.7 Model Calibration

The following criteria have been used to select the events utilized for model calibration:

- No evidence of large spatial variability of rainfall when comparing the intensities at two gauges;
- Flow monitor data shows a clear response to the rainfall events;
- No impact of snow melt; and
- Largest event in the year (July 24th, 2023) to be analyzed as requested, although there is considerable difference between records of gauge MA12 and MA20 in this event. MA20 is more reasonable for this event, as discussed in the progress meeting with the City on April 4th, 2024.

By applying these criteria, six events were selected for model calibration, as shown in **Table 10**. The recent storm occurred on June 20th, 2024, will be used to validate the results, as suggested by the City.

Table 10: Selected Events for Calibration

Event	Total Volume (mm)		Peak 5-min Rainfall Intensity (mm/hr)	
	MA12	MA20	MA12	MA20
2023-04-01	19.6	19.4	12.0	12.0
2023-04-05	12.7	12.5	33.6	26.4
2023-04-22	17.4	15.6	12.0	12.0
2023-05-20	33.8	34.6	26.4	24.0
2023-07-24	49.6	72.0	74.4	84.0
2023-10-06	27.8	28.2	50.4	43.2

Model calibration is achieved by changing model parameters to produce results matching the flow monitoring data within a reasonable accuracy in the selected events. Model validation tests the calibrated model performance using measurements different than the calibration period to ensure the repeatability of the model results.

Model calibration procedure involves the following actions:

- Adjusting the percentage of connection (fixed runoff coefficient) of other impervious surfaces to match the total monitored flow volume.
- Reclassifying a portion of roof areas as pervious surfaces to account for volume losses due to roof leaders draining onto pervious areas. Adjust this portion until the modeled runoff volume matches the flow monitoring data. Modifying the “initial abstraction” value to align with the shape of flow monitoring hydrographs.

Following the initial model set up, simulated runoff volumes were generally too large compared to the monitored flows. The contribution of impervious areas was large relative to the volumetric runoff coefficient. Since impervious surface areas were calculated using aerial imagery within ArcGIS, there may be cases wherein a pervious surface was mistakenly considered as impervious, or the runoff generated from some impervious areas are not able to be conveyed to the storm management system. The adjusted impervious runoff surface parameters are listed in **Table 11**.

Table 11: Calibrated Imperviousness Rate

Gauge ID	Residential Imperviousness Rate	ICI Area Imperviousness Rate	Calibrated Impervious Rate
S304	74%	N/A	56%
M718 West Leg	72%	96%	83%

Table 12 compares the calibrated peak flow and flow volume with the monitored records. Detailed hydrograph comparisons are provided in **Appendix A**. In most events, errors between observed and simulated flow volume are within 20%. Discrepancies and relatively high percentage differences in peak flow and flow volume in some events may be attributed to the following:

- Due to the variability of rainfall within the study area the rainfall volumes and pattern could be different for some areas as compared to the distribution used in the model, which could have an impact on the simulation.
- The size of the rainfall events which have been used for model calibration raises some doubt about the accuracy of the level measurements. Accuracy of the level meter usually decreases appreciably at flow depths below 25 mm.
- Unexpected field conditions, such as blocked catch basins, leaking pipes, and broken manholes, result in reductions in the observed flow.

Table 12: Comparison between Observation and Simulation

Event	Gauge S304						Gauge M724 (West Leg)					
	Event Flow Volume Accumulation (m³)			Event Peak Flow (L/s)			Event Flow Volume Accumulation (m³)			Event Peak Flow (L/s)		
	Simulated	Observed	Difference	Simulated	Observed	Difference	Simulated	Observed	Difference	Simulated	Observed	Difference
2023-04-01	1171	1198	-2%	125	157	-26%	533	718	-35%	49	64	-31%
2023-04-05	783	730	7%	191	191	0%	350	360	-3%	65	61	6%
2023-04-22	935	770	18%	77	79	-3%	424	426	-0.1%	31	32	-3%
2023-05-20	1380	1200	13%	224	154	31%	630	630	0%	80	57	29%
2023-07-24	1880	1502	20%	595	626	-5%	851	795	7%	202	238	-18%
2023-10-06	1129	1045	7%	292	412	-41%	518	430	17%	106	93	12%

Hydrograph comparisons for validation event are shown in **Figure 15** and **Figure 16**. Note that in 2024 flow monitors S304 and M718 (west leg) were relocated to MH 50606 and the east leg, respectively, as shown in **Figure 17**. MH50606 is upstream of the original S304 monitor location, and M718's east leg receives additional uncalibrated flows from lake areas and surrounding properties, creating some observation-simulation discrepancies. The validation results show that at manhole MH50606, simulated flow matches well with the observed data.

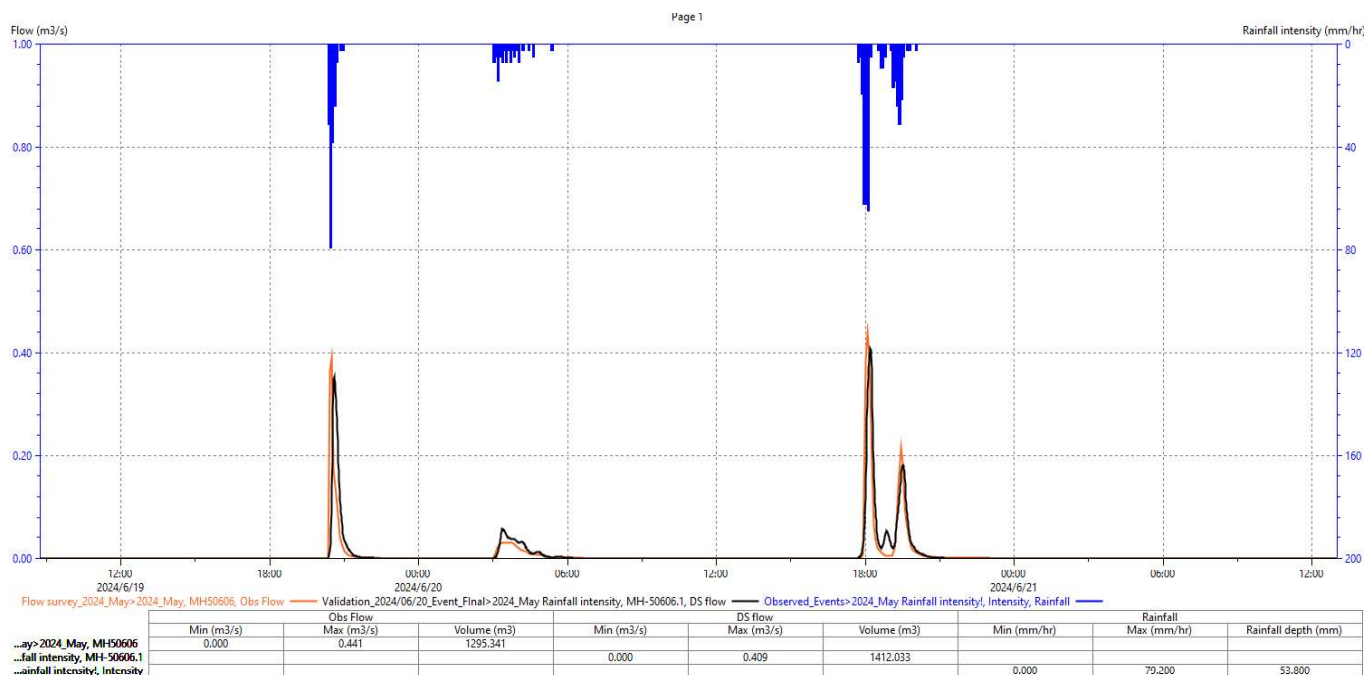


Figure 15: Model Validation - Gauge MH50606

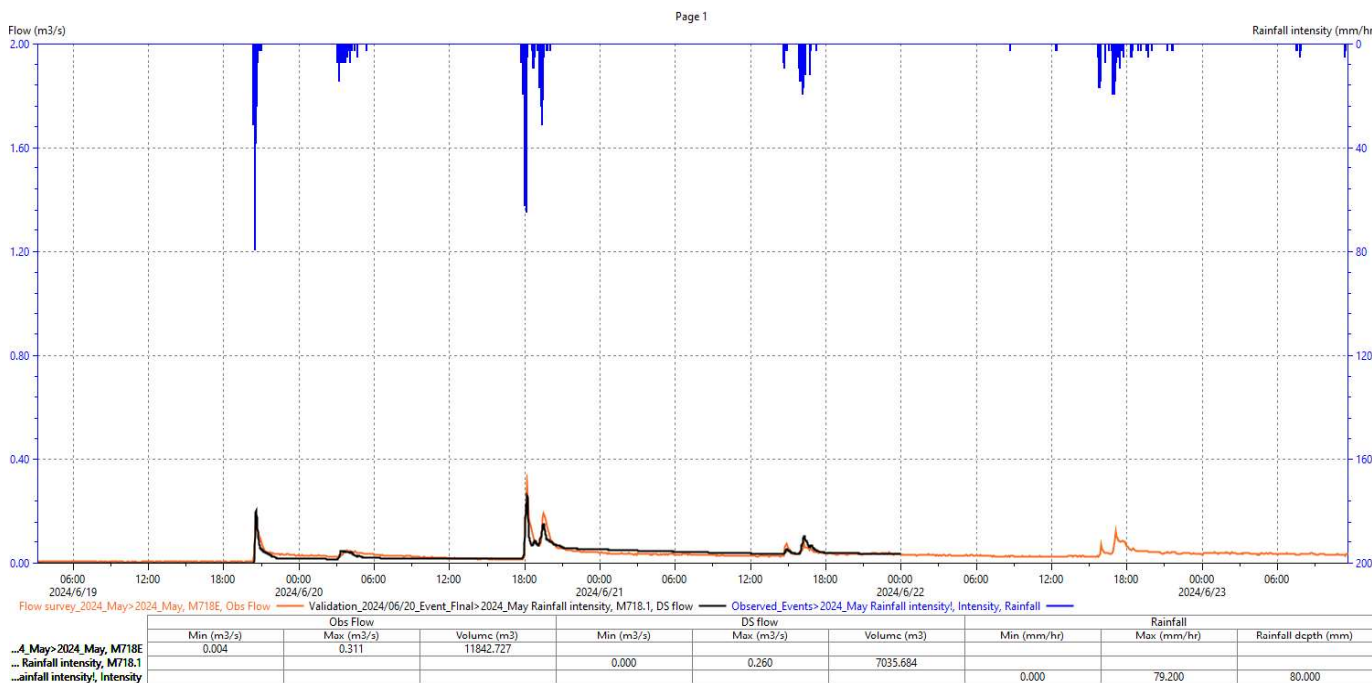


Figure 16: Model Validation - Gauge M718 East Leg



Figure 17: Location of 2024 Flow Gauges

3.8 System Evaluation

The 3-hour Chicago Design storm, obtained from City of Markham Engineering Design Criteria (Markham, 2014), were used for simulations conducted for 2-, 5-, 25-, and 100-year design storms to evaluate the storm drainage system's performance.

An existing conditions stormwater management and drainage assessment for all stormwater infrastructure within the study area was completed utilizing the calibrated hydrologic modelling to assess existing storm drainage system capacity deficiencies. The calibrated model, including minor system (catch basin inlet, manholes, storm sewers, ditch/ swales) and major system (roadways, overland flow), were simulated for the 2- to 100-year design storm events.

The results of this assessment provided an indication of the stormwater management infrastructure with capacity and flooding issues (including surcharging and flooding/overland flow conditions), identified existing levels of service for the storm drainage system and provided an indication of existing spare capacity for future development and whether the existing storm system HGL is below basement levels. Areas with storm system capacity deficiencies such as surcharged nodes, and pipes under capacity were identified from simulation results. System capacity deficiency locations for the 5- and 100-year design storm events are shown in **Figure 18** and **Figure 19**.

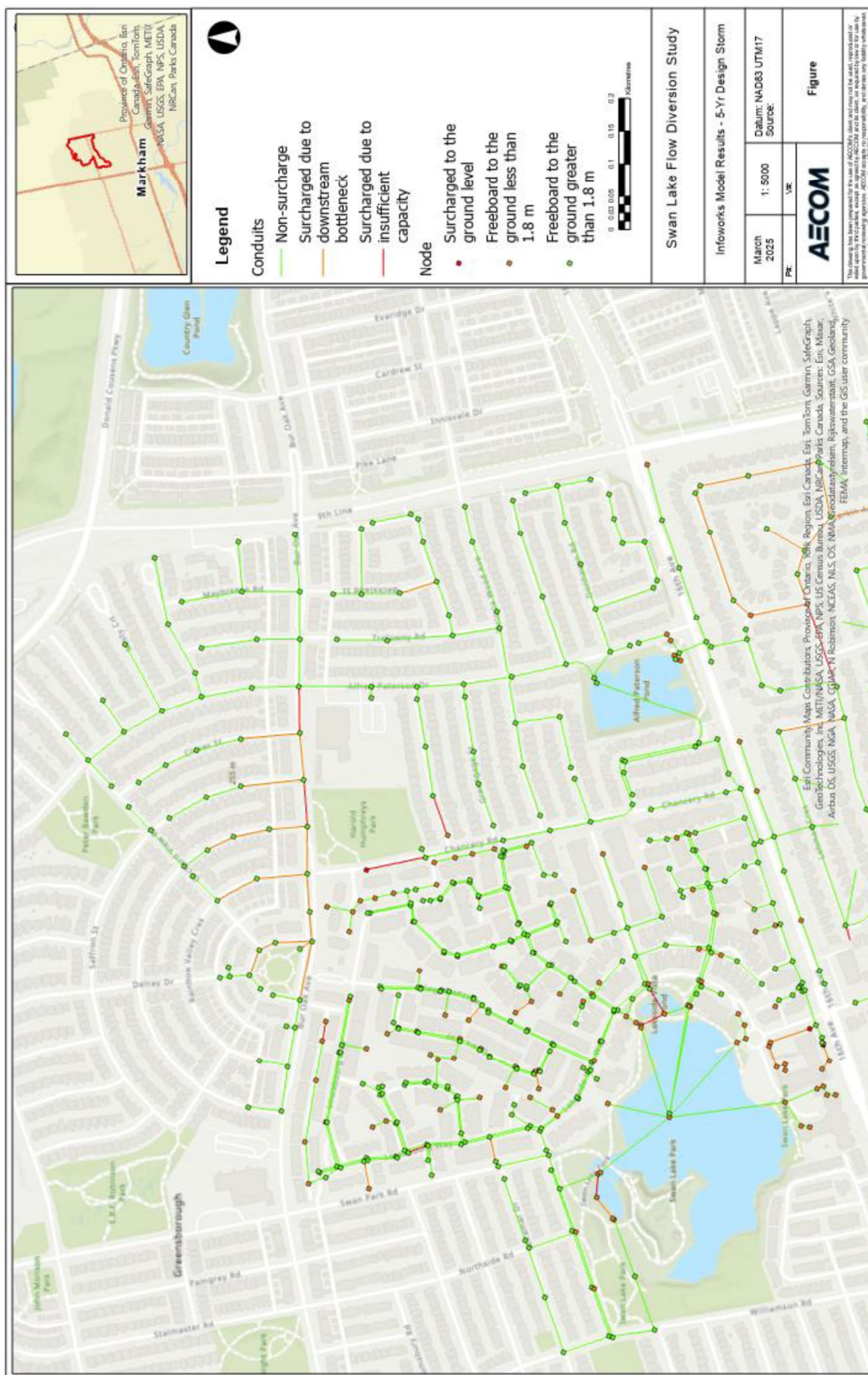


Figure 18: System Performance - 5-Year Design Storm

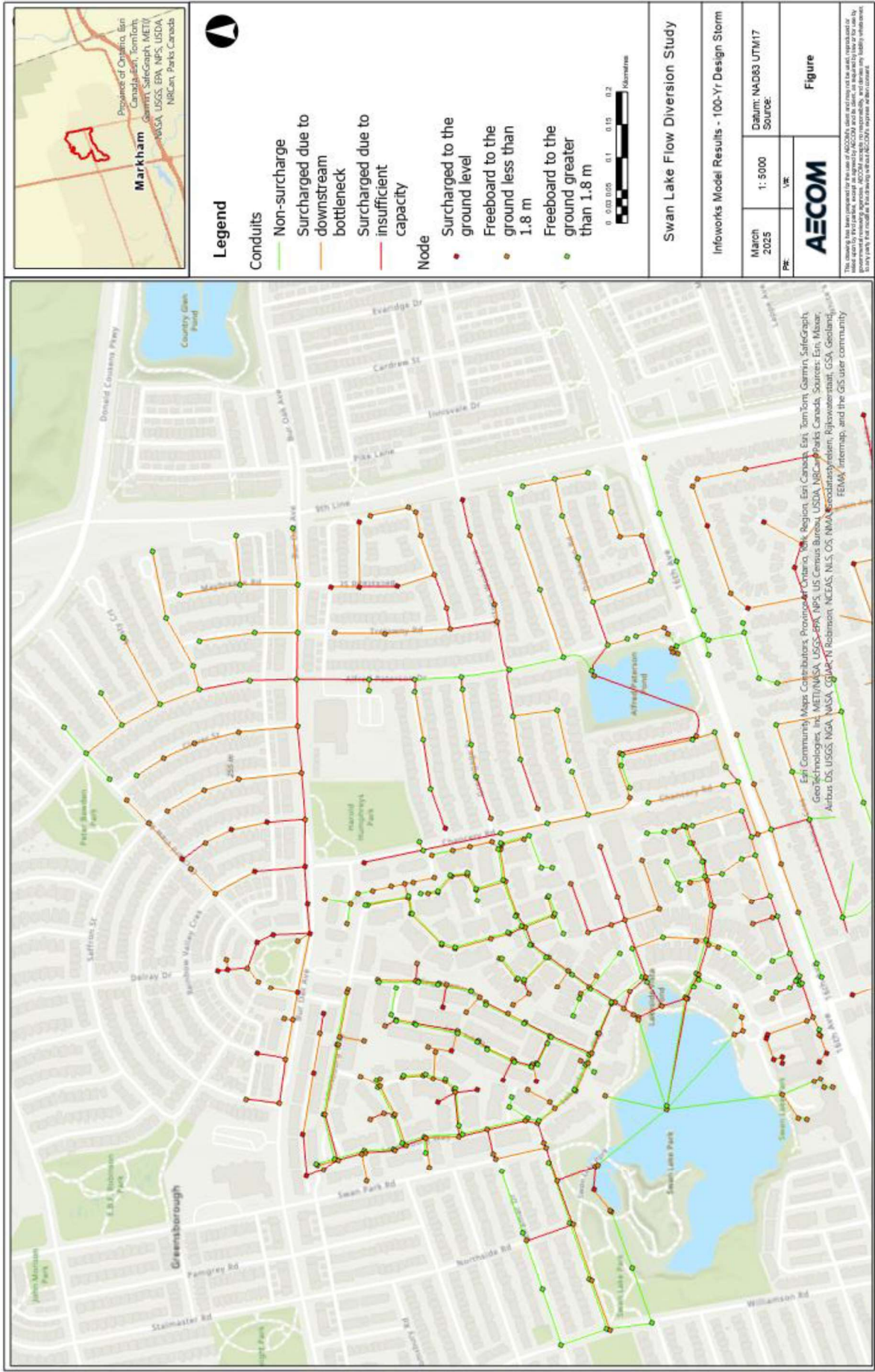


Figure 19: System Performance – 100-Year Design Storm

4. Flow Diversion Scenarios

4.1 General

This section presents the hydraulic analysis results for each diversion scenario, summarizing the reduction efficiency, cost, downstream impacts, and required infrastructure to implement the diversions as illustrated in **Figure 20**. The required new infrastructure and upgrades to the existing system were sized using the validated InfoWorks ICM model results.

City proposed flow diversion scenarios are listed as follows:

- Existing conditions (combine the existing Swan Lake catchment Infoworks model with the downstream area Markham Village and Unionville models).
- Redirecting minor system flow from the AMICA oil grit separator (OGS) and Swan Lake Blvd. OGS units to the 16th Ave. sewers.
- Redirecting minor system flow from AMICA OGS and Swan Lake Blvd. OGS units to the Lake outlet.
- Redirecting the “first flush” portion of the minor system flow from the AMICA OGS and Swan Lake Blvd. OGS units to the 16th Ave. sewer (i.e., redirect the most pollutant-laden runoff in a small diversion sewer).
- Redirecting minor system flow from Swan Club OGS to the North Pond.
- Adjusting the flow splitter weir for the East Pond and North Pond to reduce flow bypass to the Lake.
- Expanding the storage capacity in the East Pond and North Pond to reduce flow bypass to the Lake (to consider if the flow redirection scenarios increase flood risk).
- Creating underground storage capacity to attenuate the flows from AMICA OGS and Swan Lake Blvd. OGS before they enter the local sewer system (to consider if there is a feasible candidate site and if the redirecting scenarios increase flood risk).
- Redirecting/pumping flows from some foundation drain collectors (FDCs) toward Swan Lake (i.e., supply potentially cleaner, cool groundwater to the Lake).

Hydrologic information for the catchment areas for diversion is detailed as follows:

1. **Amica OGS catchment:** Amica OGS catchment is approximately 0.82 hectares in size with an imperviousness rate exceeding 90%. Total flat roof area in the catchment is approximately 0.32 ha and drains directly to the 300 mm local storm sewers. There are no sloped roofs in the Amica OGS catchment. It is assumed that runoff generated from the flat roof area will be attenuated by roof drain inlets, each with a capacity of 3 L/s at a depth of 5 cm, with one inlet per 160 m² of roof area. Overland flow is directed north to Swan Lake via Lakeside Vista Way, with no additional overland outlet to the downstream streets.
2. **Swan Lake Blvd. OGS Catchment:** Swan Lake Blvd. OGS Catchment is approximately 0.66 ha in size. This catchment contains two buildings with a combined roof area of 0.12 ha, with two out of six visible roof downspouts connected to the storm sewers. The overall imperviousness rate is approximately 75%. Overland flow travels westward and flow to Swan Lake through the double inlet catchbasin at the north end of Swan Lake Blvd.
3. **Swan Club OGS Catchment:** Swan Club OGS collects runoff from the parking lot west of the club building from a catchment area of 0.21 ha. The catch basin east of the building is connected to the 825 mm storm sewer on Lake Side Vista Way rather than the OGS. The building has one visible disconnected roof downspout on the northeast corner, which drains into the catch basin instead of the OGS. Thus, the Swan Club OGS only receives overland runoff from the parking lot through the herringbone opening.

4. **The North Pond catchment:** The North Pond catchment is approximately 12.6 hectares. This catchment consists of single-family residential areas with an imperviousness rate of 74%. A flow splitter weir located at the upstream pond inlet manholes (with a crest elevation of 208.8 m, standing 0.5 m high) diverts low flows to the North Pond and bypasses high flows to Swan Lake.
5. **East Pond Catchment:** The East Pond catchment is approximately 19.3 ha. The East Pond includes two inlets with flow splitter weirs at a crest elevation of 208.7 m (0.4 m high), designed to divert low flows to the North Pond and direct high flows to Swan Lake.

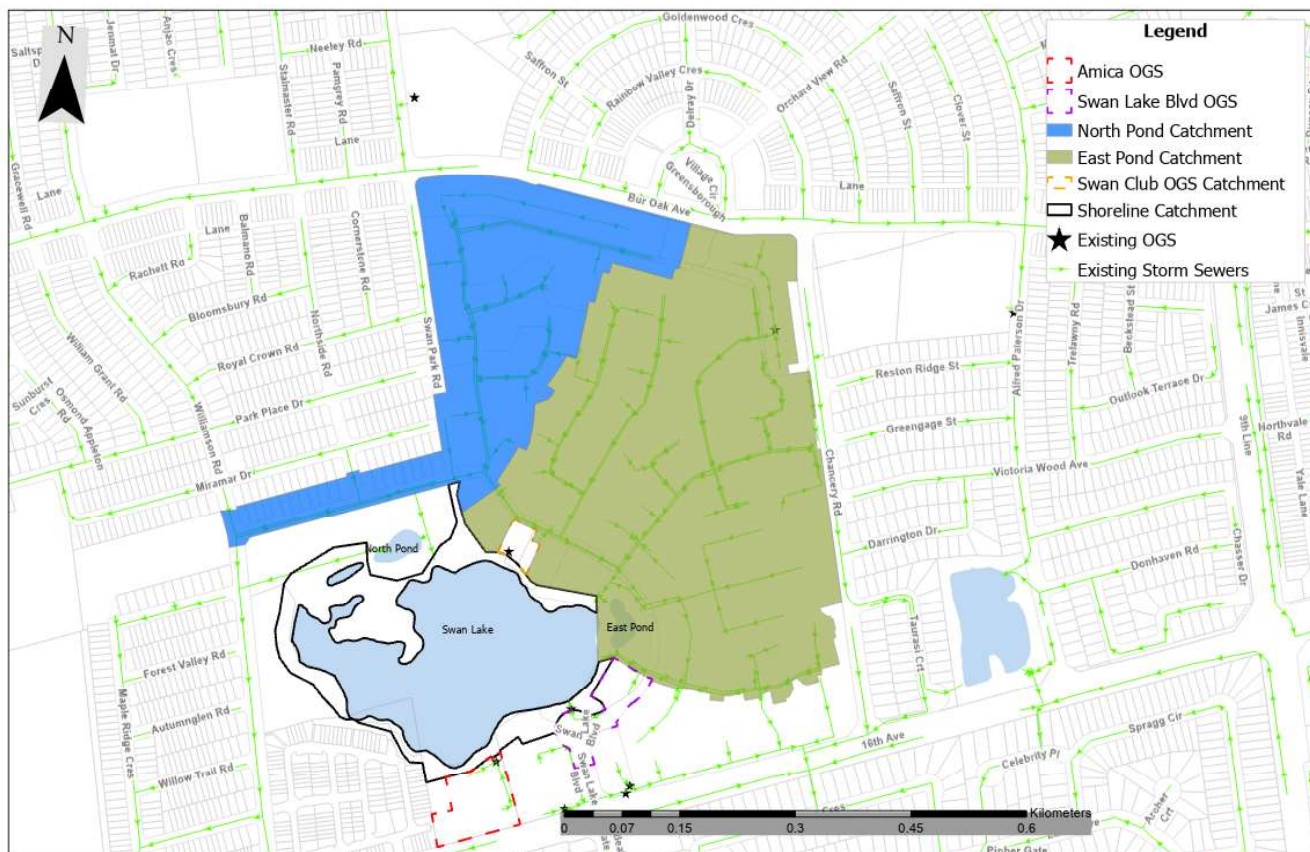


Figure 20: Catchment Area of Diversion Scenarios

4.2 Scenario Analysis

4.2.1 Scenario 1: Redirecting Minor System Flow from AMICA OGS and Swan Lake Blvd. OGS to Sewers on 16th Avenue

Diverting flow from the Amica and Swan Lake Blvd. OGS units requires installing new 450-600 mm pipe, totaling 220 m. In the 'typical' hydrologic year (2013 dataset, as provided by the City), the total inflow to Swan Lake was approximately 18,100 m³ under existing conditions. Note that 18,100 m³ represents the inflow to Swan Lake via the minor system, which does not include the rainfall volume directly received by the Lake and the flow that drains to the North and East pond. The hydraulic model results suggest that this diversion is expected to reduce the typical year flows to Swan Lake by 8,310 m³, with 5,780 m³ coming from the Amica OGS and 2,530 m³ from the Swan Lake Blvd. OGS, based on 2013 rainfall data. The diversion will increase peak flows to the downstream system by 420 L/s during a 100-year design storm event.

New storm sewers will be connected to existing sewer on Kingfisher at the intersection of Swan Lake Blvd. and 16th Ave. Existing sewers on Kingfisher Cove Way / 16th Ave range from 450 mm to 750 mm and provide a capacity of approximately 150 L/s to 920 L/s from upstream to downstream. Under existing conditions, the capacities of these sewers are exceeded during a 100-year design storm event, and the additional flow from the diversion will cause these pipes to surcharge to ground level, as shown in **Figure 21**. To mitigate flood risk, all pipes on Kingfisher Cove Way will need to be upgraded to sizes 750 to 1200 mm for a length of 645 m as shown in **Figure 22**. 100-Year HGL levels in the system with proposed upgrades are shown in **Figure 23**.

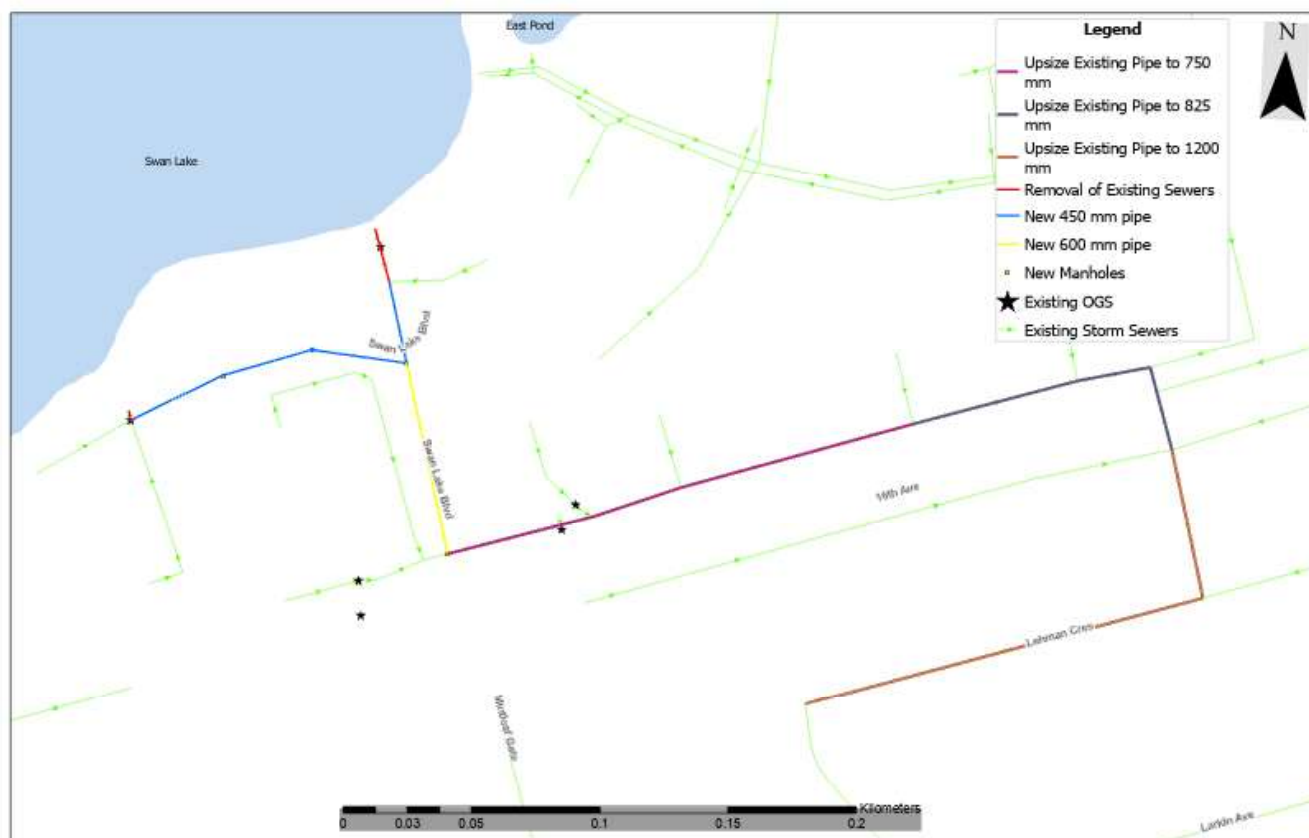


Figure 21: System Upgrade Requirement to Achieve Scenario 1

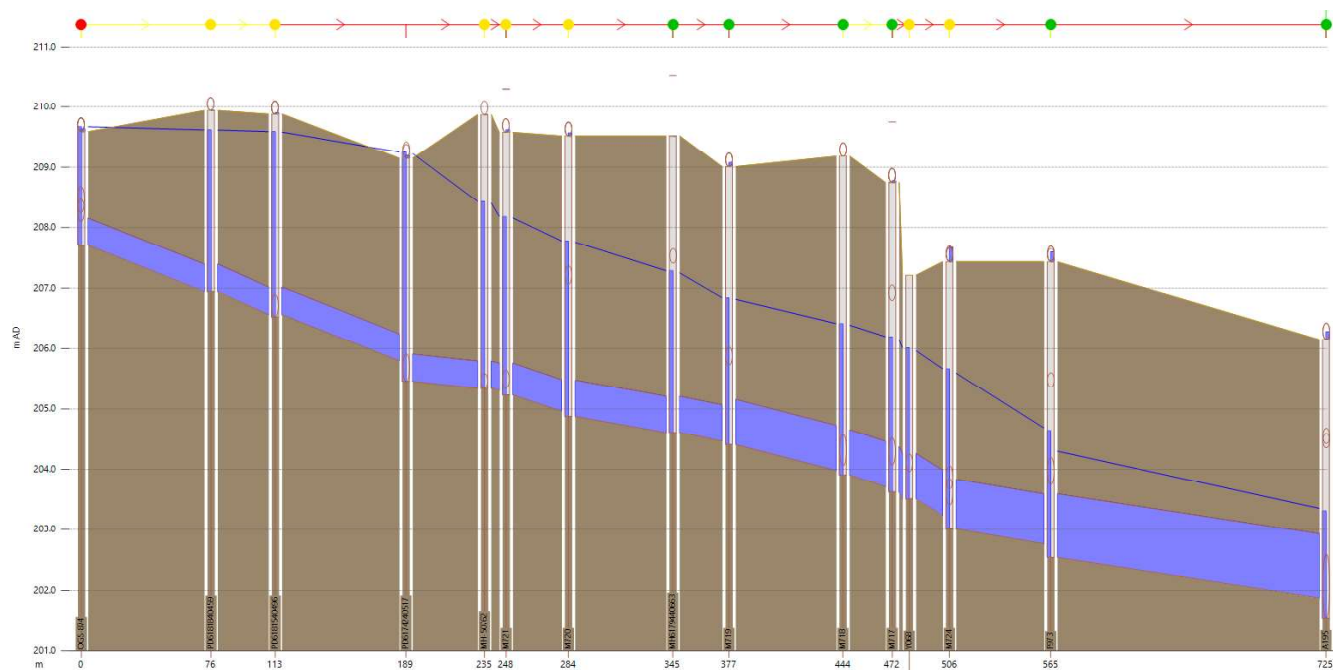


Figure 22: HGL Profile - Scenario 1 Existing Condition – 100-Year Design Storm

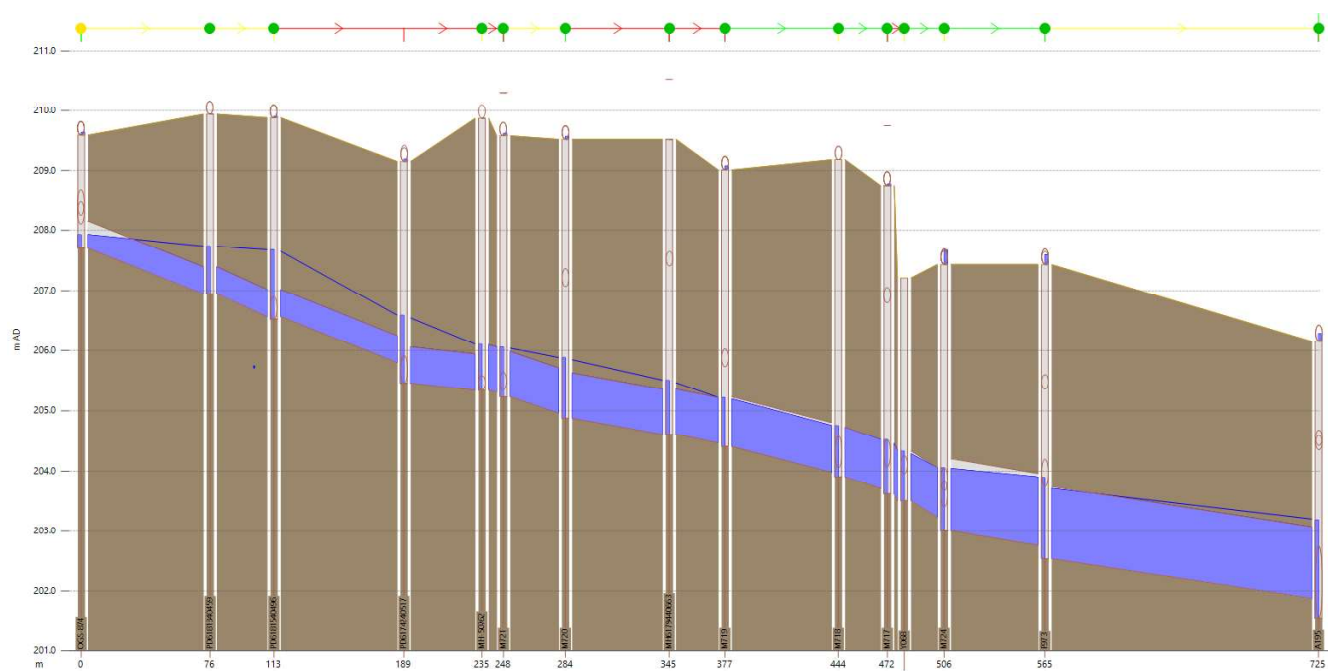


Figure 23: HGL Profile - Scenario 1 with System Upgrades - 100-Year Design Storm

4.2.2 Scenario 2: Redirecting the “First Flush” Portion of Minor System Flow from AMICA OGS and Swan Lake Blvd. OGS to 16th Avenue Sewer (i.e., redirect the most pollutant-laden runoff in a small diversion sewer)

The “first flush” refers to the initial runoff from the first 25 mm of rainfall during a storm event, which often contains higher pollutant concentrations, especially in urban areas with significant impervious surfaces. This runoff can carry significant pollutants into surface waters, including chloride from road salts in winter.

The concept of redirecting the first flush is to divert low flows to downstream sewers while allowing high flows to discharge into Swan Lake. According to the City’s GIS database, the current outfall elevations of the Amica and Swan Club OGS units to Swan Lake are 208.2 m and 208.3 m, respectively. Swan Lake’s normal water level is 208.3 m, and a 100-year design storm raises the Lake level to 208.65 m, indicating that lake water will back up into the upstream sewers during large storm event. To bypass low flows effectively and prevent backup from Swan Lake entering the upstream sewers, the flow split weir must be set at a minimum elevation of 208.6 m. Additionally, flow control should be provided at the inlet to the downstream sewers, allowing high flows to enter Swan Lake by surcharging above 208.6 m. The schematic of this configuration is shown in **Figure 24**.

As shown in the model results, setting the weir crest elevation at 208.7 m and installing 150 mm and 200 mm orifice plates at Amica and Lakeside Vista Blvd OGSs’ downstream pipes will allow all runoff generated by a 25 mm, 4 hour Chicago rainfall event with a peak 5-minute intensity of 62 mm/hr to bypass Swan Lake. Any flow exceeding the peak flow from this event will be diverted to Swan Lake. This setup reduces the typical year flow to Swan Lake by 8305 m³, achieving approximately 99% of the reduction effect of a complete disconnection of the OGS units with Swan Lake (i.e., Scenario 1 provided a reduction of 8,310 m³). This high reduction could be attributed to the absence of extreme rainfall events in the suggested typical year (2013), as the peak intensity of a 25-mm, 4-hour Chicago rainfall event - 62 mm/hr - exceeds most events in this year. Winter storms are generally smaller, and the reduction in winter runoff, which has high chloride content, is identical to that in Scenario 1; overflow through the weir only occurred during intensive summer events, as shown in **Figure 25**.

New storm sewers required for this scenario are required, and pipes are sized to 300 mm, as per the minimum required storm sewer in the City of Markham Engineering Design Criteria. The HGL in the proposed system is shown in **Figure 26**.



Figure 24: Schema - Diverting the First Flush

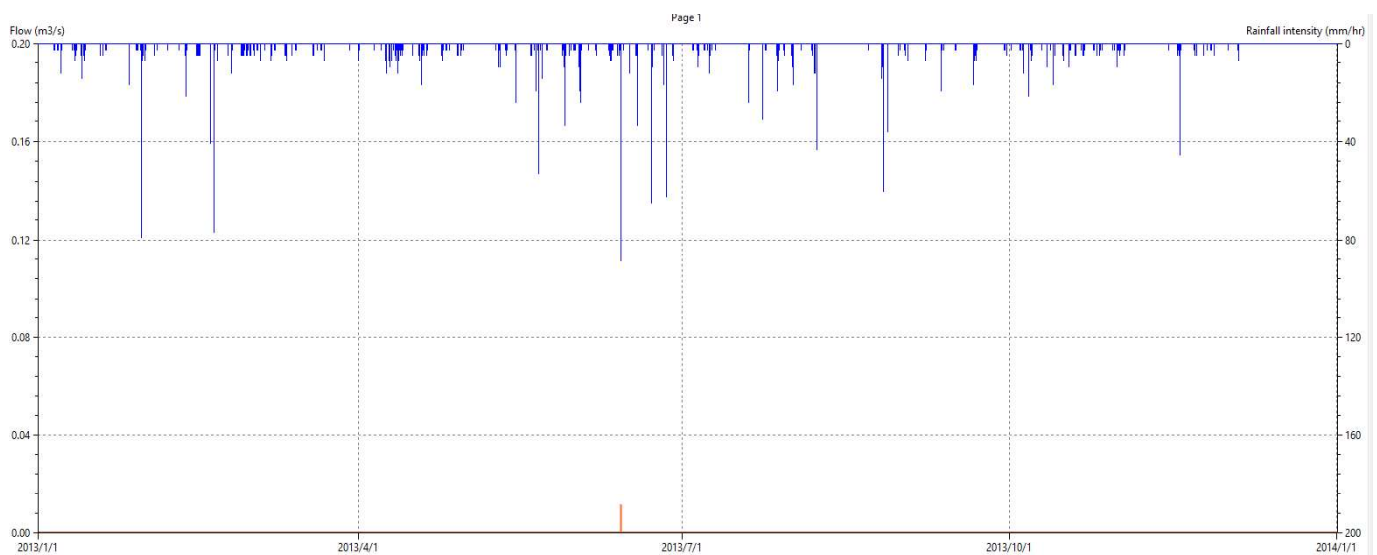


Figure 25: Typical Year Inflow to Swan Lake after Implementing Scenario 2

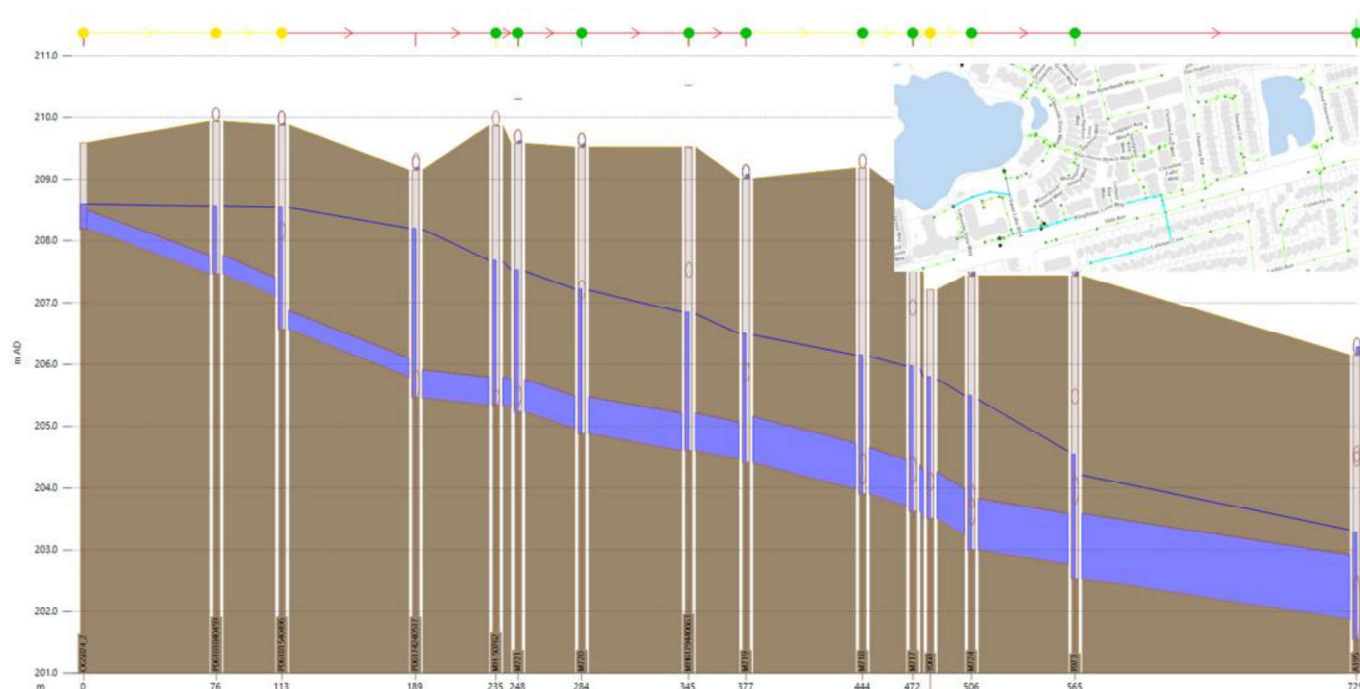


Figure 26: Scenario 2- Downstream HGL Profile

4.2.3 Scenario 3: Redirecting Minor System Flow from AMICA OGS and Swan Lake Blvd. OGS to the Lake Outlet

Identical to Scenario 1, Scenario 3 can achieve a typical year flow reduction of 8310 m³. Scenario 3 involves the installation of 200 metres of 525 mm pipes, and 4 new manholes, as shown in **Figure 27**. New pipes will be connected to the existing 450 mm sewer on Lakeside Vista way at manhole MH-50688, which in turn connect to a 525 mm lake outlet sewer on Blue Heron Beach way. This route offers the advantage of bypassing the undersized sewers on Kingfisher Cove Way. Additionally, the existing sewers on Lakeside Vista Way are buried approximately 4-5 metres below the ground, providing adequate clearance above the assumed basement level of 1.8 metres below ground.

Additional flow from the diversion will increase the HGL level in the downstream sewer however, it would still be below assumed basement level, which is 1.8 m below the ground level, in a 100-year design storm event. Therefore, compared to Scenario 1, Scenario 3 would require fewer downstream improvements, the total required improvements to downstream pipes includes upsizing 95 m of 375 mm pipe to 525 mm, as illustrated in **Figure 28**.



4.2.4 Scenario 4: Redirecting Minor System Flow from Swan Club OGS to the North Pond

The new infrastructure required for this diversion includes one new manhole and 110 m of 300 mm pipe, as shown in **Figure 29**. The active storage in the north pond is approximately 810 m³, at 208.8m, which is the level of the flow splitter weir at the pond inlet. Under existing conditions, 25 mm rainfall will generate approximately 960 m³ of flow, additional flow will cause more spill from the weir. The 450 mm inlet pipe of the North Pond has a capacity of 200 L/s. Any flow exceeding this capacity will surcharge the pipe, diverting high flows above 208.8 m to Swan Lake.

Under existing conditions, a 25 mm rainfall event will cause the surcharge level to reach 208.9 m. Additional flow from the Swan Club OGS will cause the pond capacity to be exceeded during a 25mm event, and increase the weir overflow frequency, resulting in an increase of stormwater flows to Swan Lake from the weir.

This diversion is expected achieve a typical year inflow reduction of 1230 m³, however, additional diverted flow will increase the overflow through the flow control weir at the pond inlet by approximately 240 m³ per year, limiting the net typical year flow volume reduction to 990 m³.

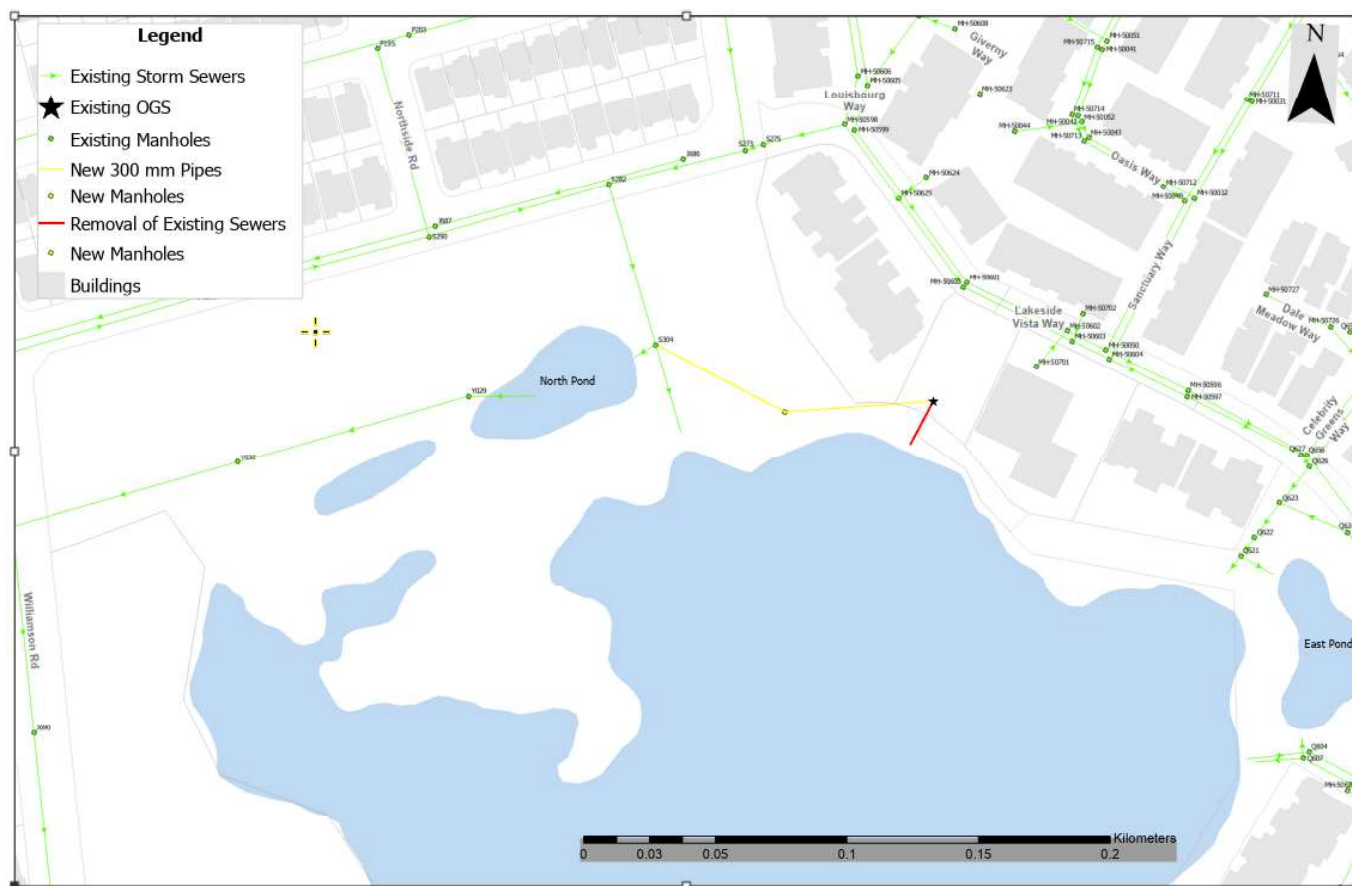


Figure 29: System Upgrade Requirement to Achieve Scenario 4

4.2.5 Scenario 5: Adjusting the Flow Splitter Weir for the East Pond and North Pond to Reduce Flow Bypass to the Lake

The existing configuration of the North and East Ponds are shown in **Figure 30**. Under existing condition, the flow splitter weirs at the inlets direct low flows to the North and East Ponds, while high flows are diverted to Swan Lake. Both weirs back flows up to the 450 mm inlet pipes to each pond, which are both approximately 200 L/s (under free outfall condition before the ponds are filled).

- **East Pond:** The East Pond has an active storage volume of 1,100 m³, at the weir elevation of 208.75 m. When this capacity is reached, any additional inflow overflows the weir at the flow splitter location and discharges directly into Swan Lake. The 1,100 m³ active storage volume represents about 60% of the runoff from a 25 mm storm. The pond initially receives inflow at a peak rate of 200 L/s. Flows exceeding this rate bypass the pond and go directly to Swan Lake due to the 450 mm inlet pipe's capacity. Once the pond fills to its maximum volume of 1,100 m³ (reaching the weir elevation at the flow splitter), nearly all additional flow is directed to Swan Lake.
- **North Pond:** Similarly, the North Pond has an active storage volume of 800 m³ at 208.8 m. When this capacity is reached, excess inflow bypasses the pond, overflowing the weir at the flow split and discharging to Swan Lake. The 800 m³ volume represents approximately 80% of the runoff from a 25 mm storm. Inflows initially reach the pond at a peak rate of 200 L/s, with any flow exceeding this rate directed to Swan Lake due to the inlet pipe's capacity. Once the pond fills to 800 m³, essentially all additional flow is diverted to Swan Lake.

Under existing conditions, the primary function of these ponds is to divert initial stormwater volumes from Swan Lake at the start of each rainfall event. However, the ponds have minimal impact on peak flow control for larger storm events, as they fill quickly and provide no further attenuation once full.

The East Pond has a spill elevation of 209.25 m, at which point it spills to Swan Lake. The weir height at the flow splitter location is 208.7 m, limiting the East Pond level from rising higher than this. There is an opportunity to raise the weir by approximately 0.3 m, thereby increasing the maximum water level in the East Pond to 209.0 m while maintaining 0.25 m of freeboard before spilling into Swan Lake. This would increase the active storage in the pond to approximately 1,200 m³ (existing active storage = 1100 m³).

The north pond has a spill elevation of 209.0 m when it spills to Swan Lake, with the weir height at the flow splitter set at 208.8 m, therefore raising the north pond weir would not significantly increase storage within the North Pond.

By raising the north and east pond weirs to 208.9 m (0.1 m rise) 209 m (0.3 m rise), respectively, the additional storage in the pond would enable a reduction of approximately 5016 m³ of flow to Swan Lake in a typical year.

However, without increasing the 450 mm inlet pipes to the ponds, the flow rate to the ponds would still be limited to about 200 L/s, leading to occasional bypass of flows to Swan Lake during short durations of intense rain.

An additional scenario was analyzed by upsizing the pond inlet pipes to 600 mm, resulting in a reduction of flow to Swan Lake by approximately 5499 m³ in a typical year. The increase in flow to the downstream sewers caused by pond upgrades is negligible; raising the weirs would increase the allowable water depth in the north and east ponds by 0.1 m and 0.3 m, respectively. This would result in approximately 2.5 L/s and 1.9 L/s of additional flow through the 100 mm and 66 mm orifice plates at the north and east pond outlets, respectively. However, it remains uncertain whether the surcharge conditions in the downstream sewer of the north pond, along Williamson Road, would impose any restrictions on pond outflows. Such restrictions could potentially impact the effectiveness of the pond upgrade options. The capacity of the downstream sewer on Williamson Road is not analyzed in this study; further investigation is required to confirm the downstream condition.

Potential upstream catchment impacts as a result of the raised weir height are discussed as a part of the Scenario 6 section.

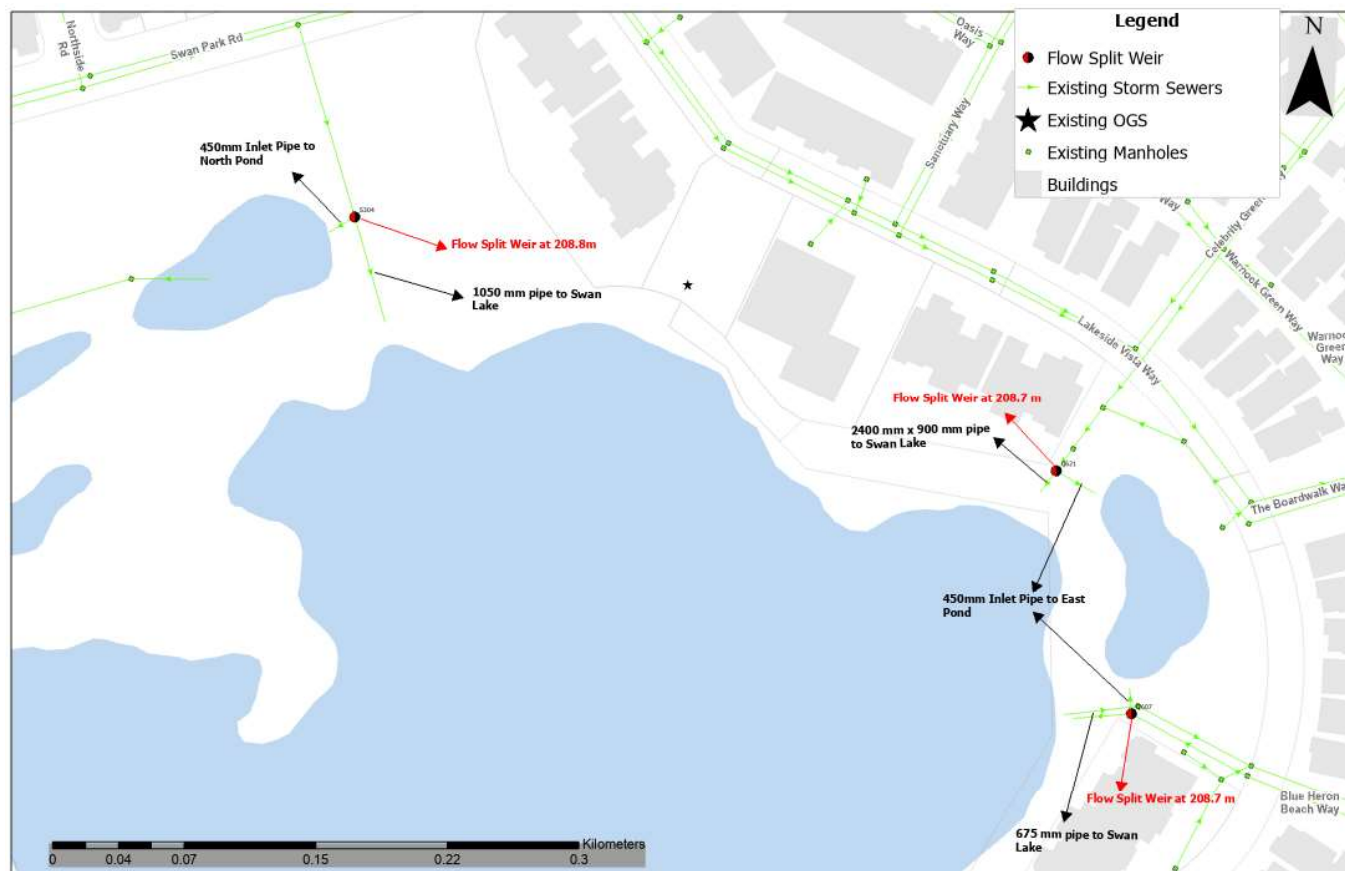


Figure 30: Existing Configurations of North and East Ponds

4.2.6 Scenario 6: Expand Storage Capacity in the East and North Ponds to Reduce Flow Bypass to the Lake (to consider if the redirection scenarios increase flood risk)

In this scenario, in addition to raising the weir and upsizing the inlet pipes, the North and East ponds were proposed to be retrofitted to provide an active storage of 3140 m³ and 3126 m³ respectively. The proposed pond layouts, and stage-storage relation are shown in **Figure 31** and **Table 13**, respectively. Enlargement of the north pond would require elimination of a small section of trail, but residual connections would generally offset any potential negative connectivity consequences. The enlarged facility would still be able to make beneficial use of the local park space for short-term sediment drying during construction and maintenance operations (for example, directly to the northeast of the north pond), although and major maintenance operations would likely necessitate trail restoration.

Enlargement of the east pond in the manner shown would consume the majority of the open space in the SWMF block and would require realignment of the existing trail on the on the south, east and northern sides. While a trail connection still appears to be feasible, the remaining SWM block area is generally understood not to provide sufficient space for sediment drying. This means that bulking would need to occur within the eastern SWMF, or that high-moisture content material may need to be transported offsite as liquid waste (in accordance with O.Reg.

406/19) as part of any future sediment removal operations completed under an expanded pond scenario. This added maintenance complexity can generally be expected to increase the cost of SWMF maintenance.

The proposed additional storage (include raising the weir and upsizing the inlet pipes) could reduce the typical year bypass flow to swan lake by 8,226 m³. The comparison of typical year flow reduction at each inlet between existing condition, raising the weir, raising the weir while upsizing the inlet pipe, and raising the weir while upsizing the inlet pipe and upsizing the pond is summarized in **Table 14**. To analyze the cost efficiency of retrofitting each pond, **Table 15** provides a comparison of the reduction efficiencies between the North Pond and East Pond upgrades.

Table 13: S6 Pond Expansion - Stage-Storage Relation

Stage (m)	Area (m ²)	Average Area (m ²)	Volume (m ³)	Storage (m ³)
Proposed North Pond				
208.3	2916	0	0	0
208.4	3036	2976	298	298
208.6	3280	3158	632	929
208.8	3530	3405	681	1610
209	3787	3659	732	2342
209.2	4051	3919	784	3126
East Pond				
208.2	2569	0	0	0
208.4	2790	2680	536	536
208.6	3018	2904	581	1117
208.8	3252	3135	627	1744
209	3492	3372	674	2418
209.2	3738	3615	723	3141

Table 14: Comparison of Flow Reduction at Inlets

Scenarios	Typical Year Inflow to Swan Lake at Each Inlet (m ³)				Typical Year Reduction Compared to Existing Condition (m ³)	Typical Year Reduction as % of Total Swan Lake Inflow
	North Pond Inlet	East Pond North Inlet	East Pond South Inlet	Total		
Existing	2583	4310	1360	8253	n/a	n/a
Raise Weir (Scenario 5a)	1152	1395	690	3237	5016	27.7%
Raise the weir, upsize the inlet pipe (Scenario 5b)	902	1102	750	2754	5499	30.4%
Raise the weir, upsize the inlet pipe, expand the pond (Scenario 6)	10	17	0	27	8226	45.4%

Table 15: Comparison of Flow Reduction between Upgrading North and East Pond

Scenarios	Reduction at North Pond (m ³)	Reduction at East Pond (m ³)
Existing	0	0
Raise Weir (Scenario 5a)	1431	3585
Raise the weir, upsize the inlet pipe (Scenario 5b)	1681	3818
Raise the weir, upsize the inlet pipe, expand the pond (Scenario 6)	2573	5653

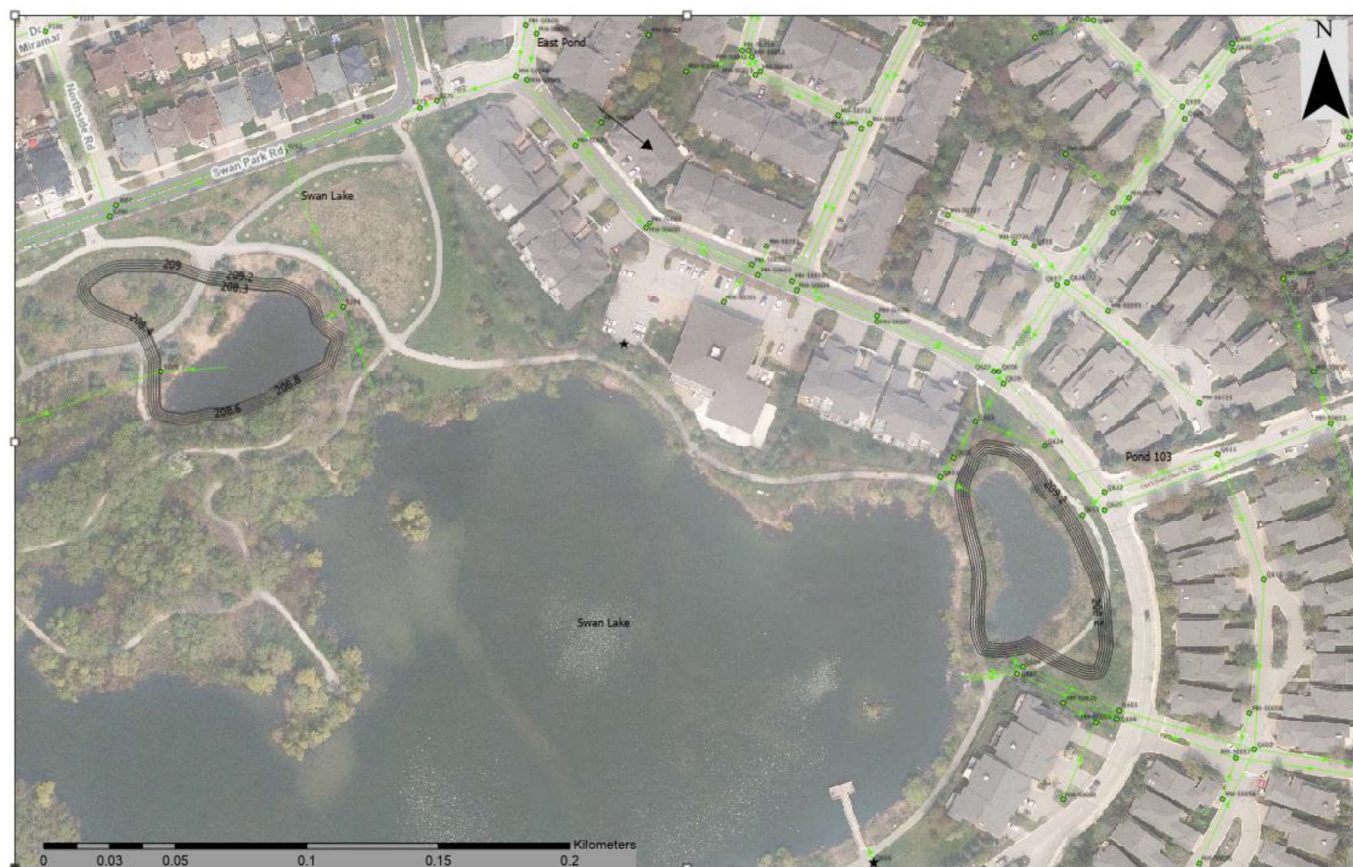


Figure 31: Proposed Pond Contour (Active Storage)

Upstream Catchment Hydraulic Grade line Impacts as a Result of Increased Pond Depths

Among all three pond retrofit scenarios (5a, 5b, and 6), solely raising the weir will result in the highest increase of water level during extreme storm events (e.g., the 100-year design storm), as additional flow are diverted to the ponds without corresponding increase in storage capacity. The increased flow will lead to a rise in the hydraulic grade line (HGL) in upstream sewers. As shown in **Figure 30** to **Figure 35**, the HGL in the sewers upstream of the north pond remained largely unchanged, with only a 0.02 m increase after the weir was raised. This is because the 0.1 m increase in the pond water level does not significantly increase the flow rate through the north pond inlet pipes. Additionally, the north pond has a lower spill level (209 m), and this level is exceeded under existing condition during a 100-year storm, which limits the volume of flow that can be diverted into it.

The HGL in the sewers upstream of the east pond's north and south inlets increased by 0.2 m and 0.15 m, respectively. While this increase will not cause additional manholes to surcharge to the ground surface, it will slightly elevate the basement flooding risk for properties connected to these pipes. This is because the pipes are relatively shallow and do not provide 1.8 m of freeboard to the ground surface. However, since foundation drain collector (FDC) pipes are presented in the catchment area of the east pond, it is possible that basements of these properties are connected to the FDC system rather than directly to the storm sewers. Connectivity tests are recommended to confirm basement connections.

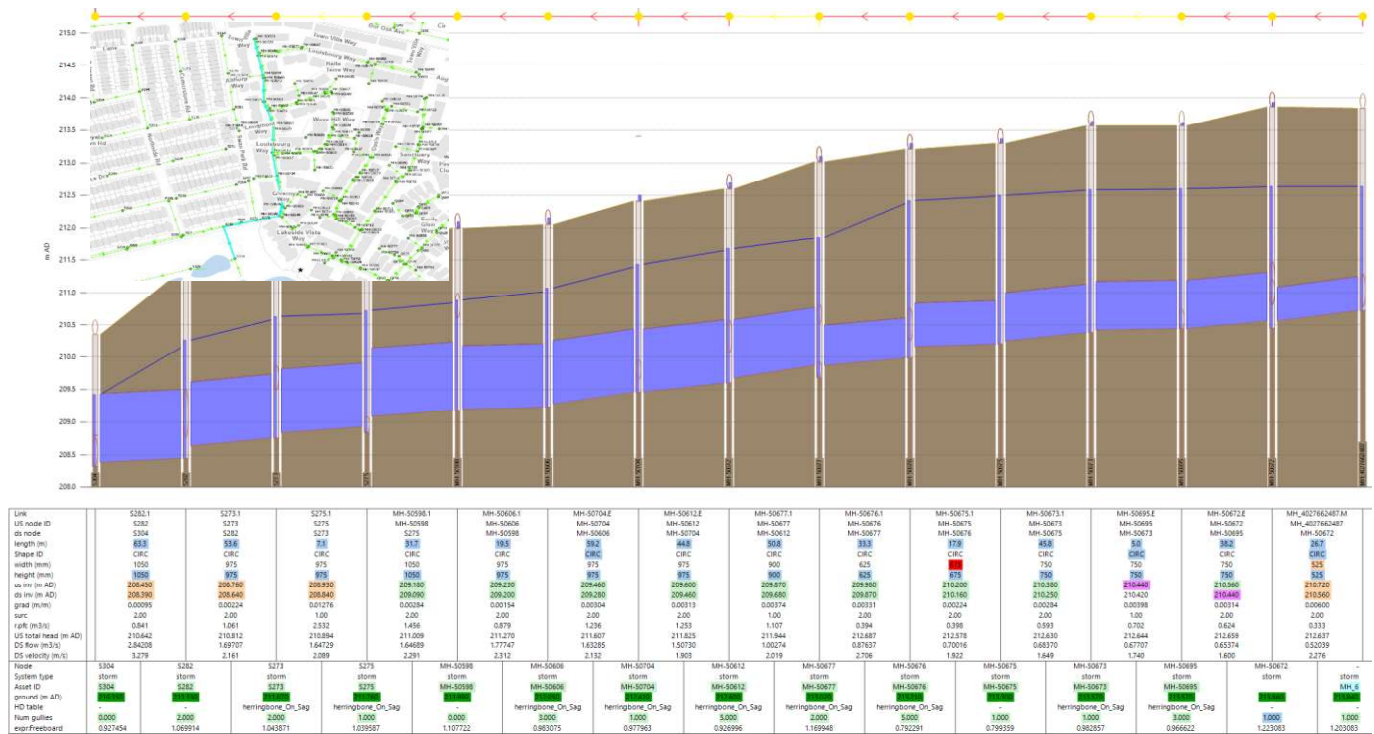


Figure 32: 100 Yr HGL Level – Pipes Upstream to North Pond – Base Scenario

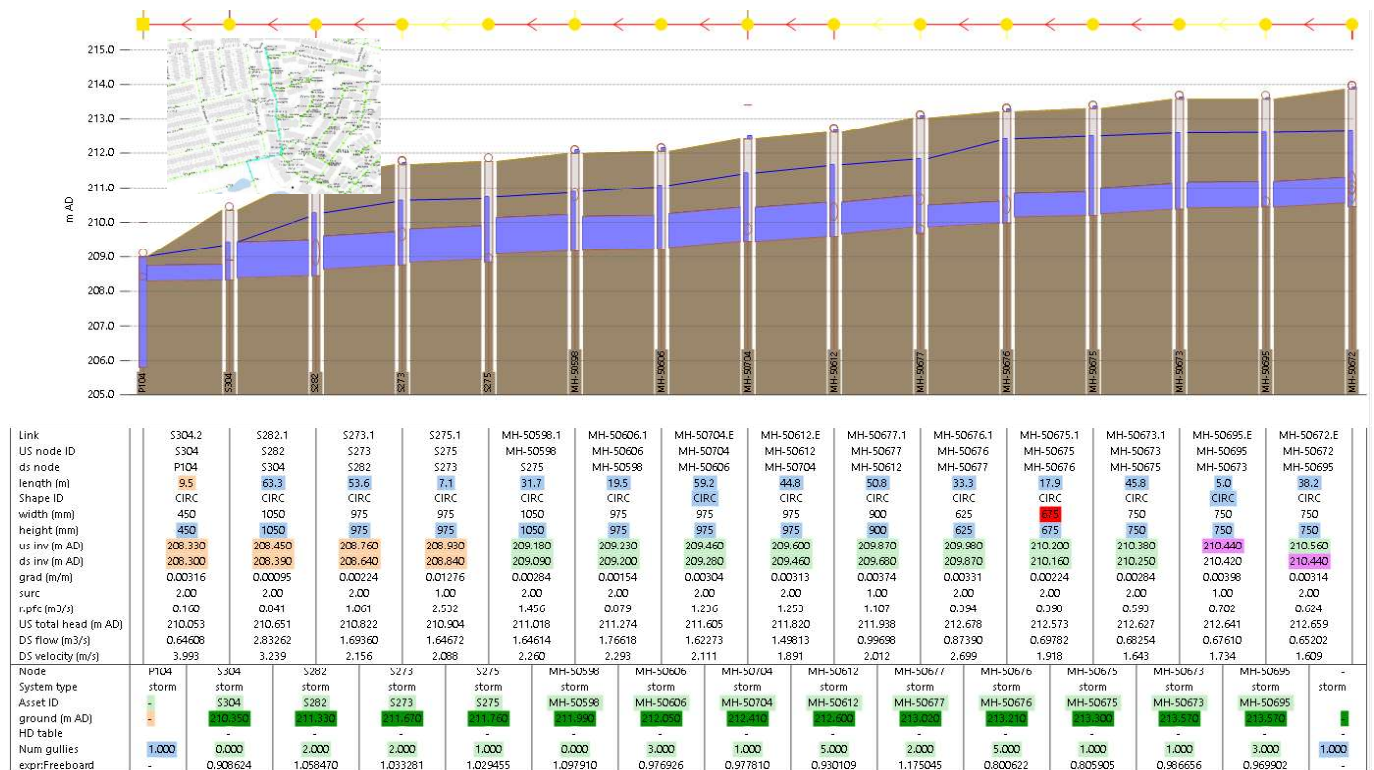


Figure 33: 100 Yr HGL level – Pipes upstream to North Pond – Scenario 5: Raising the Overflow Weir



Figure 34: 100 Yr HGL Level – Pipes Upstream to East Pond North Inlet – Base Scenario

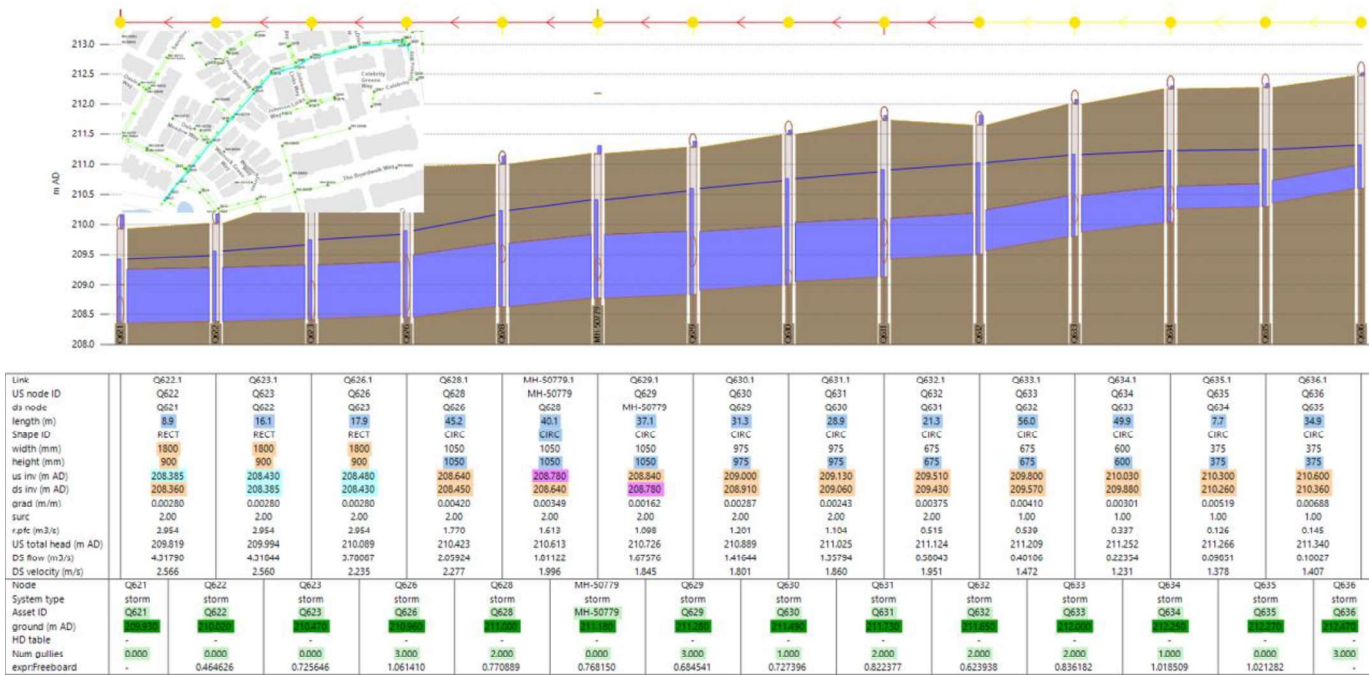


Figure 35: 100 Yr HGL level – Pipes Upstream to East Pond North Inlet – Raising the Overflow Weir



Figure 36: 100 Yr HGL Level – Pipes Upstream to East Pond South Inlet – Base Scenario



Figure 37: 100 Yr HGL level – Pipes Upstream to East Pond South Inlet – Raising the Overflow Weir

4.2.7 Scenario 7: Combine Scenario S4 (Redirecting Minor System Flow from Swan Club OGS to the North Pond) with North Pond Upgrade Options

The previous section (Scenario S4) has illustrated that diverting flow from Swan Club OGS to the north pond would increase the bypass flow to Swan Lake through the overflow weir. This occurs because the additional flow from the Swan Club OGS not only increases the total volume entering the pond but also raises the peak flow rate in the inlet pipe, elevating the HGL level and causing bypass flow to occur more frequently. As a result, Scenario S4 is a less effective option. However, north pond upgrades, including raising the weir elevation, upsizing the pond inlet pipes, and pond expansion, would reduce significantly this overflow. Diverting more flow to the north pond will increase the inflow through 450 mm pipe, potentially increasing the risk of bypass due to insufficient pipe conveyance capacity, making upsizing the pond inlet pipe more necessary.. **Table 16** shows the reduction effect of combining Scenario S4 with different North Pond upgrade options.

Table 16: Comparison of Flow Reduction Between Combined Scenarios

Scenarios	Flow Volume Drains Swan Lake Through Swan Club OGS (m ³)	Flow Volume Drains Swan Lake Through North Pond Overflow Weir (m ³)	Typical Year Flow Reduction (m ³)
Existing	1210	2573	0
S7a- Combining Scenario 4 and the North pond portion of Scenario 5a (Raising the weirs)	0	1335	2468
S7b- Combining Scenario 4 and the North pond portion of Scenario 5b (Raising the weirs and upsizing pond inlet pipes)	0	1083	2720
S7c- Combining Scenario 4 and the North pond portion of Scenario 6b (Pond expansion)	0	13	3790

4.2.8 Scenario 8: Creating Underground Storage Capacity (to attenuate the flows from AMICA OGS and Swan Lake Blvd. OGS before they enter the local sewer system)

Mintleaf Gate was selected as the site for the presumed construction of underground storage pipes to manage the diverted flow from the Amica and Swan Lake Blvd. OGS units before discharging into the downstream pipes. This site was chosen based on the following considerations:

- The existing sewers are undersized, as shown in the model results.
- There will be no constraint in the downstream sewers after the system upgrades on Larkin Street are implemented.

The required size of this storage pipe depends largely on the roof areas directly connected to the storm sewers. The original 2021 model assumes that 44% of the roofs on Mintleaf Gate are directly connected to the storm sewers. However, the actual connection rate may be significantly lower, as observed through Google Earth. Due to this consideration, AECOM conducted a visual inspection from the right-of-way. The results indicate that, of approximately 70 visible downspouts, 7 are connected to the storm or sanitary sewers, while the remainder discharge directly onto the ground, as shown in **Appendix B**. This data has been incorporated into the InfoWorks model to accurately quantify the runoff generated by roof area that drains directly to storm sewers.

Model results indicate that a 208 m long, 2400 x 1200mm box storage pipe with a 550 mm orifice plate for outlet control will accommodate the existing and diverted flow generated by a 100-year design storm event. This setup controls the post-diversion flow to 350 L/s, which is lower than the existing minor system flows on Mintleaf Gate (420 L/s), providing a slight benefit to downstream conveyance capacity.

Through discussions with the City, AECOM was requested to investigate the feasibility of replacing the existing east and north stormwater management ponds with underground storage. Since these ponds were designed not only to provide runoff quantity management but also to deliver water quality benefits (through the inclusion of such design elements as sediment forebays, for example) it is recommended that, rather than constructing an underground box culvert system as proposed for Mintleaf Gateway, subsurface storage chambers - complete with an isolator row - could be constructed to allow sediments to settle.

To provide the same total storage volume (active storage plus permanent storage) as proposed for the north and east pond expansions in Scenario 6, the required chamber sizes are approximately 6,630 m³ for the east pond and 7,370 m³ for the north pond. Applying an estimated unit cost of \$850 per cubic metre, the total cost to replace both ponds would be approximately \$11.9 million, which is significantly higher than the cost of pond expansion. Furthermore, the water surface elevation within Swan Lake and either stormwater management pond indicates that hydrostatic uplift of any subsurface chambers may make such techniques infeasible. While the City may review available groundwater elevation data in this area in order to make an informed decision, the significant cost associated with such works further suggests that this option is infeasible.



Figure 38: System Upgrade Requirement to Achieve Scenario 7 (Box Storage Implementation on Mintleaf Gate)

4.2.9 Scenario 9: Redirecting/Pumping Flows from Some Foundation Drain Collectors (FDCs) toward Swan

Based on the available flow monitoring data in the FDC system at maintenance hole J689, which services approximately 63 properties along Miramar Drive and Swan Park Road, the annual measured flow volume is only 69 m³. This indicates that diverting FDC flow to Swan Lake will not significantly impact the overall chloride levels in the lake. However, this volume may be underestimated due to reduced monitoring accuracy when flow levels are below the sensor detection threshold, especially considering that foundation drainage is typically continuous, uniform, and low in flow rate. Additional flow monitoring is recommended to confirm the impact of this scenario.

5. Scenario Evaluation

Table 17 lists the proposed diversion options and describes the advantages, disadvantages, cost of implementation, flow reduction and cost of downstream improvements for each. This table is used as a screening method to evaluate the overall effectiveness of flow reduction, the impact of each option, and the ease of implementation for each diversion alternative.

Table 17: Scenario Evaluation

Scenario ID	Description	Typical Year Flow Reduction (m³)	Cost of Implementation	Required Access to Private Property	Comment
S1	Redirecting minor system flow from AMICA OGS and Swan Lake Blvd. OGS to sewers on 16th Ave..	8310	\$7,062,688.10	Y	<p>Benefits:</p> <ul style="list-style-type: none"> Completely divert all flows from these 2 OGSs to Swan Lake <p>Disadvantages</p> <ul style="list-style-type: none"> All the downstream sewers on Kingfisher Cove need to be upsized to accommodate the additional flow. Requiring road excavation
S2	Redirecting the "first flush" portion of minor system flow from AMICA OGS and Swan Lake Blvd. OGS to 16th Ave. sewer.	8305	\$1,109,178.53	Y	<p>Benefits:</p> <ul style="list-style-type: none"> Lower peak flow contribution to the downstream sewers by allowing high flows to be diverted to Swan Lake Winter flows with high chloride content can be fully diverted to the downstream; intense summer storms that do not create a chloride concern would still go to Swan Lake. Preventing backwater at AMICA OGS <p>Disadvantages</p> <ul style="list-style-type: none"> Local sewers are lower than normal Lake level, Potential risk of back water from the Lake to enter downstream system during extreme event. Requiring road excavation
S3	Redirecting minor system flow from AMICA OGS and Swan Lake Blvd. OGS to the Lake outlet	8310	\$1,757,024.75	Y	<p>Benefits:</p> <ul style="list-style-type: none"> Able to bypass the undersize sewers on Kingfisher Cove Downstream pipes are buried deep, additional flow to the system is less likely to raise a concern of basement flooding <p>Disadvantages</p> <ul style="list-style-type: none"> Implementation require access to work on gated private properties Requiring road excavation
S4	Redirecting minor system flow from Swan Club OGS to the North Pond	990	\$275,071.88	Y	<p>Benefits:</p> <ul style="list-style-type: none"> This large parking lot may be subject to high winter salt usage. Excavation is on open space. <p>Disadvantages</p> <ul style="list-style-type: none"> Spacing for such construction is very limited. Additional flow to North Pond will increase the spill of water to Swan Lake from the pond. Installing new pipes will require the removal of trees.

Scenario ID	Description	Typical Year Flow Reduction (m³)	Cost of Implementation	Required Access to Private Property	Comment
S5a	Adjusting the flow splitter weir for the East Pond and North Pond to reduce flow bypass to the Lake	5016	\$30,420.00	N	<p>Benefits:</p> <ul style="list-style-type: none"> Minor construction work is required <p>Disadvantages</p> <ul style="list-style-type: none"> Increases basement flooding risks for properties in the east pond catchment area Raising the weir at North Pond won't significantly increase the active storage as the spill level is at 209 m and current weir is at 208.8 m Small inlet pipe size will limit flow to the pond, causing the flow to bypass to the Lake during short durations of intense rain.
S5b	Raising the flow splitter weir at the North and East Ponds and upsizing the inflow pipes	5499	\$124,787.00	N	<p>Benefits:</p> <ul style="list-style-type: none"> Reduces the backwater caused by the limit of the 450mm pipe to enter Swan Lake <p>Disadvantages</p> <ul style="list-style-type: none"> Diverting additional flow will cause the ponds to spill to the Lake more frequently Installing new pipes will require the removal of trees.
S6	Expand storage capacity in the East and North Ponds to reduce flow bypass to the Lake	8266	\$2,963,987.00	N	<p>Benefits:</p> <ul style="list-style-type: none"> Provide more storage <p>Disadvantages</p> <ul style="list-style-type: none"> Requiring additional open spaces to be converted to pond area. Will require the removal of existing lakeside trails and trees East pond lot may not have sufficient space to realign the trail after expansion
S6b	Expand storage capacity in North Pond to reduce flow bypass to the Lake	2573	\$1,662,872.25		<p>Benefits:</p> <ul style="list-style-type: none"> Provide more storage <p>Disadvantages</p> <ul style="list-style-type: none"> Requiring additional open spaces to be converted to pond area. Will require the removal of existing lakeside trails and trees Not cost-effective as the catchment area is relatively small
S6c	Expand storage capacity in East Pond to reduce flow bypass to the Lake	5653	\$1,301,114.75		<p>Benefits:</p> <ul style="list-style-type: none"> Provide more storage Large catchment area <p>Disadvantages</p> <ul style="list-style-type: none"> Spacing in the pond lot is limited for expansion. Requiring additional open spaces to be converted to pond area. Will require the removal of existing lakeside trails and trees

Scenario ID	Description	Typical Year Flow Reduction (m³)	Cost of Implementation	Required Access to Private Property	Comment
S7a	Combining S4 (Divert minor system flow from Swan Club OGS to the North Pond) with S5a (raising the North Pond flow split weir)	2468	\$285,211.88	N	<p>Benefits:</p> <ul style="list-style-type: none"> ■ This large parking lot may be subject to high winter salt usage. ■ Excavation is on open space. ■ Compared to Scenario 4, bypass to the lake is reduced. <p>Disadvantages</p> <ul style="list-style-type: none"> ■ Spacing for such construction is very limited. ■ Installing new pipes will require the removal of trees. ■ Diverting additional flow will cause the pond to spill more frequently, and the additional flow from Swan Club OGS will increase further increase the spill frequency.
S7b	Combining S4 (Divert minor system flow from Swan Club OGS to the North Pond) with S5b (Raising the North Pond flow splitter weir and upsizing the inflow pipes)	2720	\$315,544.13	N	<p>Benefits:</p> <ul style="list-style-type: none"> ■ This large parking lot may be subject to high winter salt usage. ■ Excavation is on open space. ■ Significantly reduces the backwater caused by the flow increased from Swan Club OGS and the limit of the 450mm pipe. <p>Disadvantages</p> <ul style="list-style-type: none"> ■ Spacing for such construction is very limited. ■ Installing new pipes will require the removal of trees. ■ Additional flow from Swan Club OGS will cause the pond to spill more frequently, and the additional flow from Swan Club OGS will increase further increase the spill frequency.
S7c	Combining S4 (Divert minor system flow from Swan Club OGS to the North Pond) with S6 (Expanding North Pond)	3770	\$1,897,359.75	N	<p>Benefits:</p> <ul style="list-style-type: none"> ■ This large parking lot may be subject to high winter salt usage. ■ Bypass flow to Swan Lake caused by the additional flow from Swan Club OGS is reduced. <p>Disadvantages</p> <ul style="list-style-type: none"> ■ Spacing for such construction is very limited. ■ Installing new pipes will require the removal of trees Requiring additional open spaces to be converted to pond area. ■ Not cost-effective as the catchment area is relatively small
S8	Creating underground storage capacity to attenuate the flows from AMICA OGS and Swan Lake Blvd. OGS before they enter the local sewer system (to consider if there is a feasible candidate site and if the redirection scenarios increase flood risk)	8310 (46% of total typical year flow to Swan Lake)	\$6,098,470.63	Y	<p>Benefits:</p> <ul style="list-style-type: none"> ■ No impact to downstream sewers ■ Able to mitigate the basement flood risks on Mint Leaf Gate Way, as these properties were identified as "Med Risk" in the Markham Village and Unionville Study (RVA, 2021) <p>Disadvantages</p> <ul style="list-style-type: none"> ■ New storage sewers need to be buried deep to match the invert level of the downstream sewer, which makes construction more expensive. ■ Excavation will occur on roads.

Scenario ID	Description	Typical Year Flow Reduction (m³)	Cost of Implementation	Required Access to Private Property	Comment
S9	Redirecting/pumping flows from some foundation drain collectors (FDCs) toward Swan Lake	None	Not Analyzed	N	Benefits: <ul style="list-style-type: none">■ Foundation seepage water is considerable cleaner than stormwater runoff. Pumping clean water into the Lake will dilute chloride concentration in the Lake. Disadvantages <ul style="list-style-type: none">■ The flow monitor at J689 shows a flow volume of only 64 m³ per year for FDC. Diverting it into Swan Lake would not provide water quality benefits.

6. Continuous Simulation for Water Budget

A continuous simulation from 2009 to 2024, using rainfall data collected at gauge MA 12, has been conducted to support the City's water and chloride budget analysis. Results from the calibrated model, including the runoff inflows to Swan Lake, the north pond, and the east pond, as well as lake levels and outflows over the simulation period, have been provided to the City. Evapotranspiration and groundwater recharge will be calculated separately by the City, and the results of the water and chloride budget analysis will be provided in a separate memo prepared by the City.

7. Conclusions

The conclusions of the study are:

For the four scenarios that redirect flows from Amica and Swan Lake Blvd. OGS, the conclusions are:

- **S1- Directing flows to 16th Avenue** is the most expensive option, reducing typical year inflow by 8,310 m³ at a cost of \$7.06 million. This high cost is primarily due to the undersized existing pipes on Kingfisher Cove Way, which would be unable to convey the additional flow without causing surface overflow during a 100-year design storm event. To accommodate the increased flow, upgrading approximately 645 metres of storm sewer along this route would be necessary.
- **S8- Directing flows from the Amica and Swan Lake Blvd. OGS units to underground storage pipes** on Mintleaf Gateway is also a costly option with an estimated cost of \$6.10 million to divert 8,310 m³. So, while this option provides a significant advantage by preventing negative downstream impacts, and the storage pipe could mitigate basement flooding risks for properties along Mintleaf Gate, the cost per volume of water redirected from Swan Lake is amongst the highest.
- **S-3 Directing flows to the Lake outlet** avoids the constraints of undersized pipes on Kingfisher Cove Way. This scenario is more cost-effective than diverting flows to 16th Avenue, achieving the same reduction effect at only 25% of the cost (\$1.75 million). While the additional flow will still cause surcharge in downstream pipes, due to the depths of downstream pipes, HGL remains below the assumed basement level (1.8 m underground). As a result, fewer downstream improvements are required compared to diverting flows to 16th Avenue.
- **S2- Directing “first flush” portion of the minor system flow from the Amica and Swan Lake Blvd. OGS units to 16th Avenue** is the most cost-effective option among the four. It requires smaller local pipes and achieves 99% (8,305 m³) of the reduction compared to a complete disconnection of the OGS units, but at a significantly lower cost. The total estimated cost of implementation - \$1.11 million - is only 16% of the cost of diverting to 16th Avenue and 63% of the cost of diverting to the Lake outlet. Furthermore, it avoids the need for downstream upgrades. However, this scenario introduces a risk of Lake water backing up into the upstream and downstream sewers, as the invert levels of these OGS units are at the lake's normal water level. Additional measures to prevent backflow should be considered during implementation.

All four of these options would require additional upgrades to the original proposed solution on Larkin Avenue for the Markham Village area. The costs of these upgrades are not included in the current estimates, potentially making these three scenarios more expensive.

For the three scenarios which involve pond upgrades, the conclusions are:

- **S5a- Raising flow splitter weirs** at pond inlets is the most cost-effective option, reducing approximately 5,016 m³ of stormwater inflow at a cost of approximately \$30,000, however, this option would slightly increase the basement flooding risks for properties in the east pond catchment area. Additionally, due to the limiting size of inlet pipes to the ponds, the flow rate to the ponds would still be limited, leading to occasional bypass of flows to Swan Lake during short durations of intense rain. Diverting additional flow will also cause the ponds to spill to the Lake more frequently.
- **S5b- Raising the flow splitter weirs and upsizing the inflow pipes** will increase the reduction effect to 5,499 m³, however, diverting additional flow will cause the ponds to spill to the Lake more frequently. The cost is approximately \$125 thousand.

- **S6a/6b/6c- Expanding the North and East ponds** increases the typical year flow reduction to 8,226 m³. However, this comes at a significantly higher cost—around \$2.96 million. When comparing the expansion of the two ponds, upgrading the east pond is more cost-effective than upgrading the north pond. This is because, for a similar increase in storage volume, the catchment area of the east pond is twice that of the north pond, allowing it to collect more runoff and achieve a greater flow reduction potential. In addition, the complex terrain at the North Pond would require more volume of soil excavation. Upsizing the east pond will cost \$1.30 million and result in a typical year inflow reduction of 5,653 m³, whereas upsizing the north pond will cost \$1.66 million and achieve a lower typical year inflow reduction of 2,573 m³. Notwithstanding this conclusion, the East Pond is constrained by the adjacent roadway, trail network and limited residual pond block size. Further expansion of this facility may be hampered by constraints in realigning local trails and maintaining setbacks from Lakeside Vista Way. Long-term maintenance of the pond will be impacted by limited space for staging and sediment drying area. Future investigation of these matters is recommended.

For the scenarios that redirect Swan Club OGS to the North Pond, the conclusions are:

- **S4- Diverting the Swan Club OGS** is not considered to be a cost-effective option due to its small catchment area (0.21 ha). This OGS is expected to reduce direct discharge to Swan Lake by 1,210 m³ in a typical year, at a cost of \$275,000, however, without upgrading the North Pond, additional flow directed into the pond will increase bypass flows through the flow splitter weirs, reducing the effectiveness of this solution.
- **S7a- Combining Scenario 4 (diverting Swan Club OGS to the North pond) with the North Pond portion of Scenario 5a (raising the weir)** retains the advantages of both individual scenarios. The benefits include the minimized construction work primarily occurring in open spaces and the cost-effectiveness. Additionally, this combination slightly mitigates the main drawback of Scenario 4, where the additional flow diverted to the north pond increases bypass flow to the Lake. The combined scenario reduces bypass flow to the lake by 2,468 m³. Compared to only raising the weir at the North pond, the combined scenario achieves an additional inflow reduction of 1,037 m³ at a cost of \$285,000.
- **S7b- Combining Scenario 4 with the North Pond portion of Scenarios 5b (raising the weir and upsizing the pond inlet pipe)**, moderately reduces the additional bypass to Swan Lake caused by the extra flow from the Swan Club OGS. This scenario achieves a typical year inflow reduction of 2,720 m³ at a cost of \$315,000. Compared to the Scenario 5b (the North pond portion) this Scenario has increased flow typical year reduction by 1,040 m³.
- **S7c- Combining Scenario 4 (diverting Swan Club OGS) with Scenario 6 (North pond expansion)** provides the highest typical year inflow reduction of 3,770 m³. However, this option comes at a significantly higher cost of approximately \$1.88 million and requires additional long-term maintenance and involves the same constraints as Scenario 6.

One scenario was considered involving redirecting Foundation Drain Collector flows to the Lake:

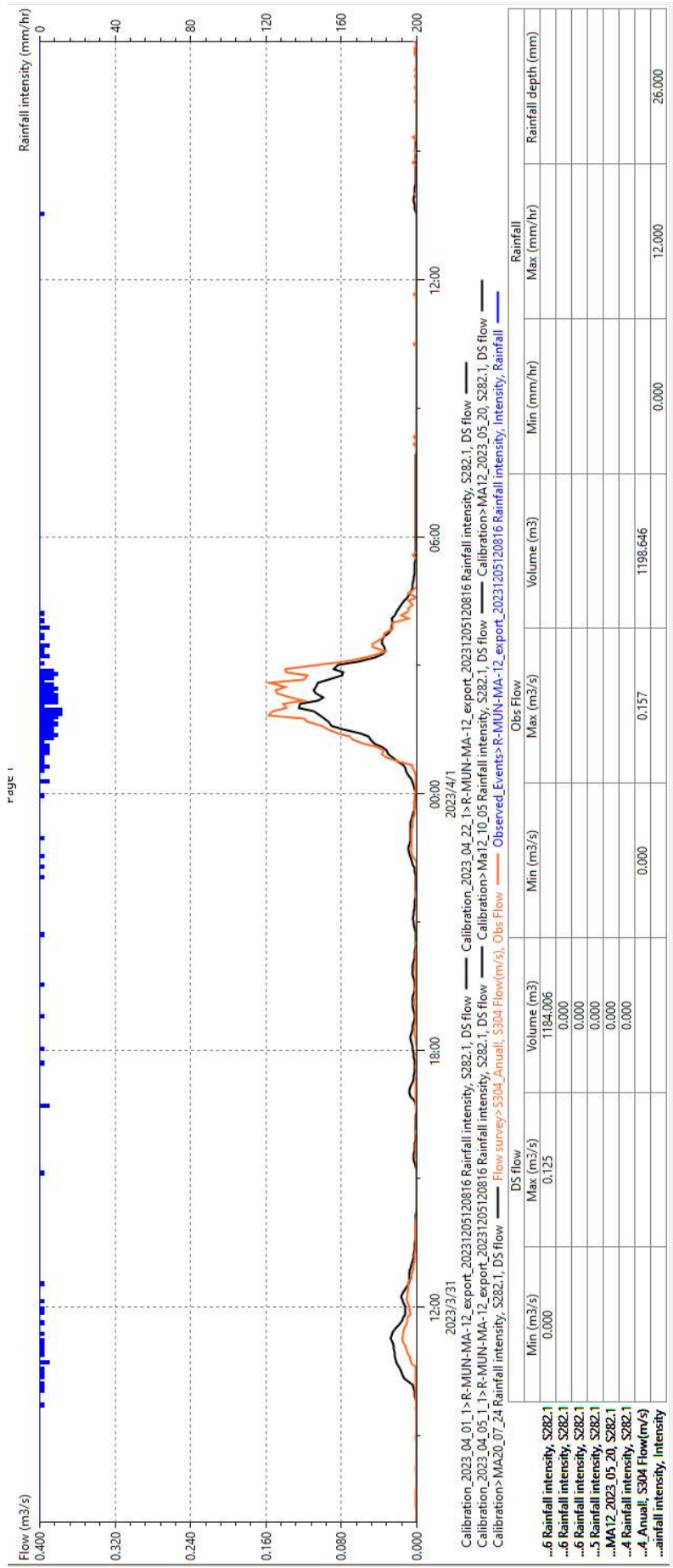
- **S9- Pumping Foundation Drain Collector (FDC) flow to Swan Lake** would not significantly impact chloride levels in the Lake. Flow monitoring data indicates an annual FDC flow of only 69 m³, which may be underestimated due to reduced monitoring accuracy at low water levels. Further FDC flow monitoring at different locations may be required to calibrate the FDC flow parameters.

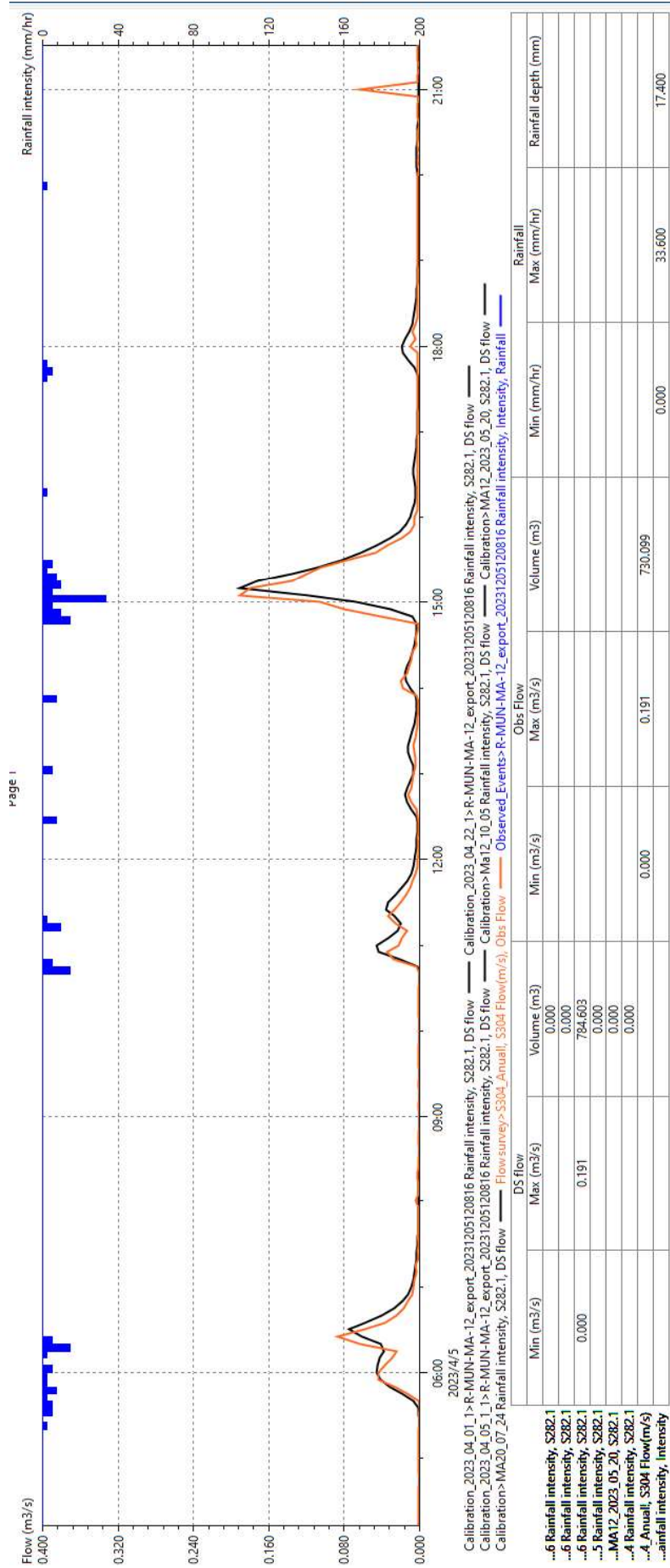
Appendix **A**

Hydrograph Comparison



Gauge S304 Hydrograph comparison for Calibration events







Swan Lake Flow Diversion Assessment



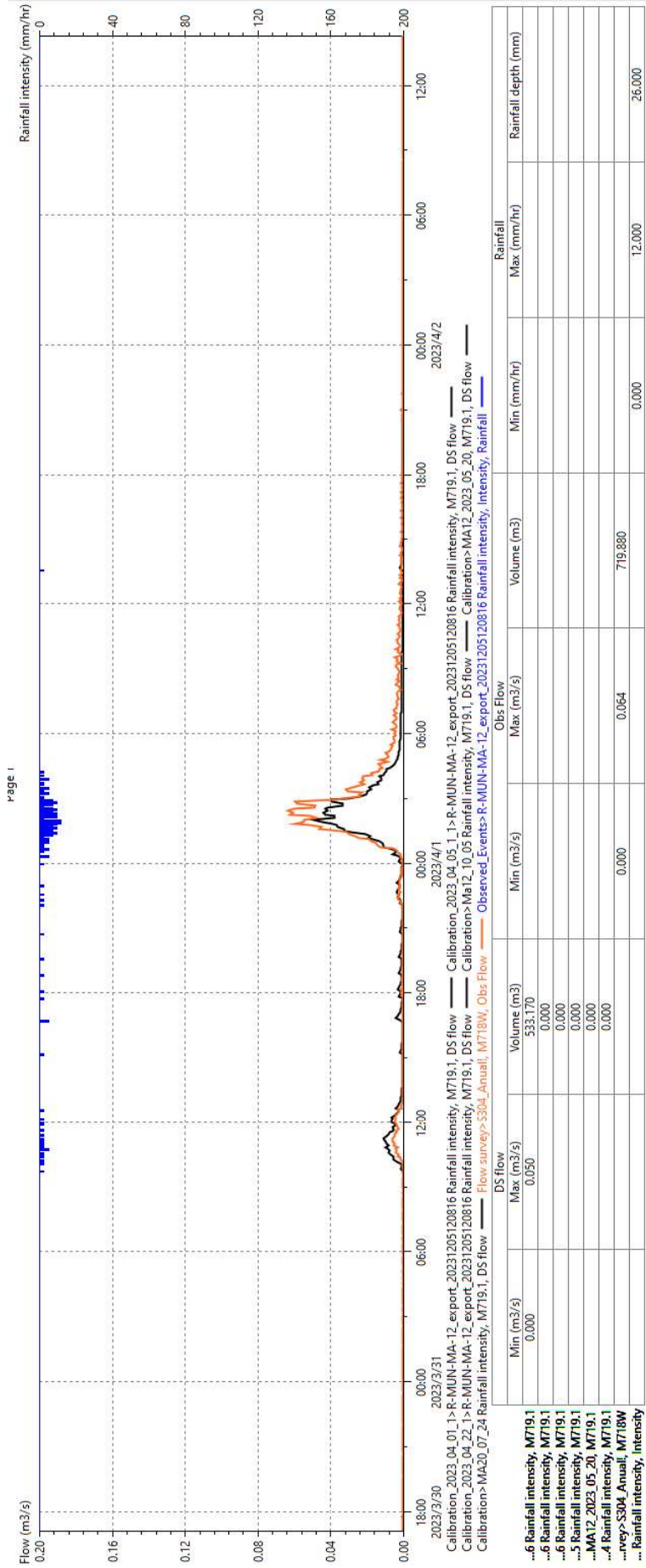
Swan Lake Flow Diversion Assessment

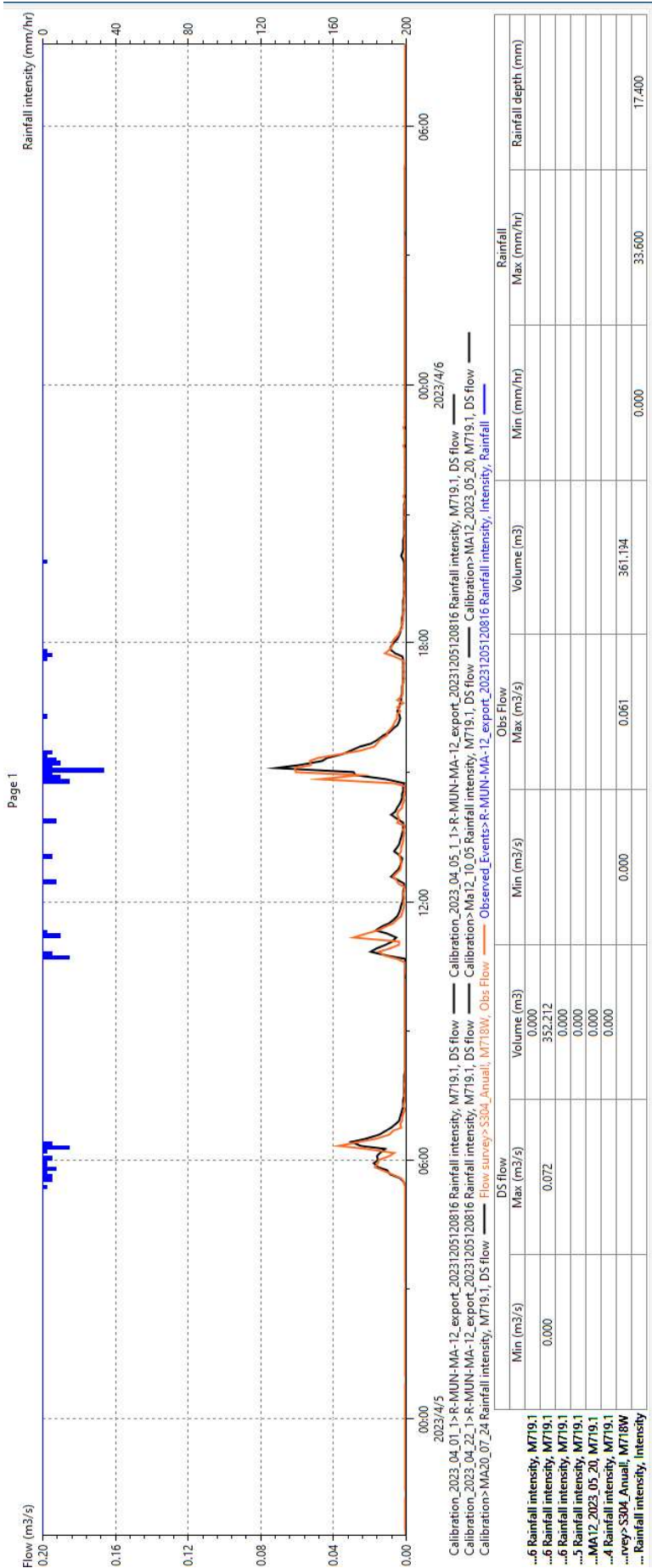


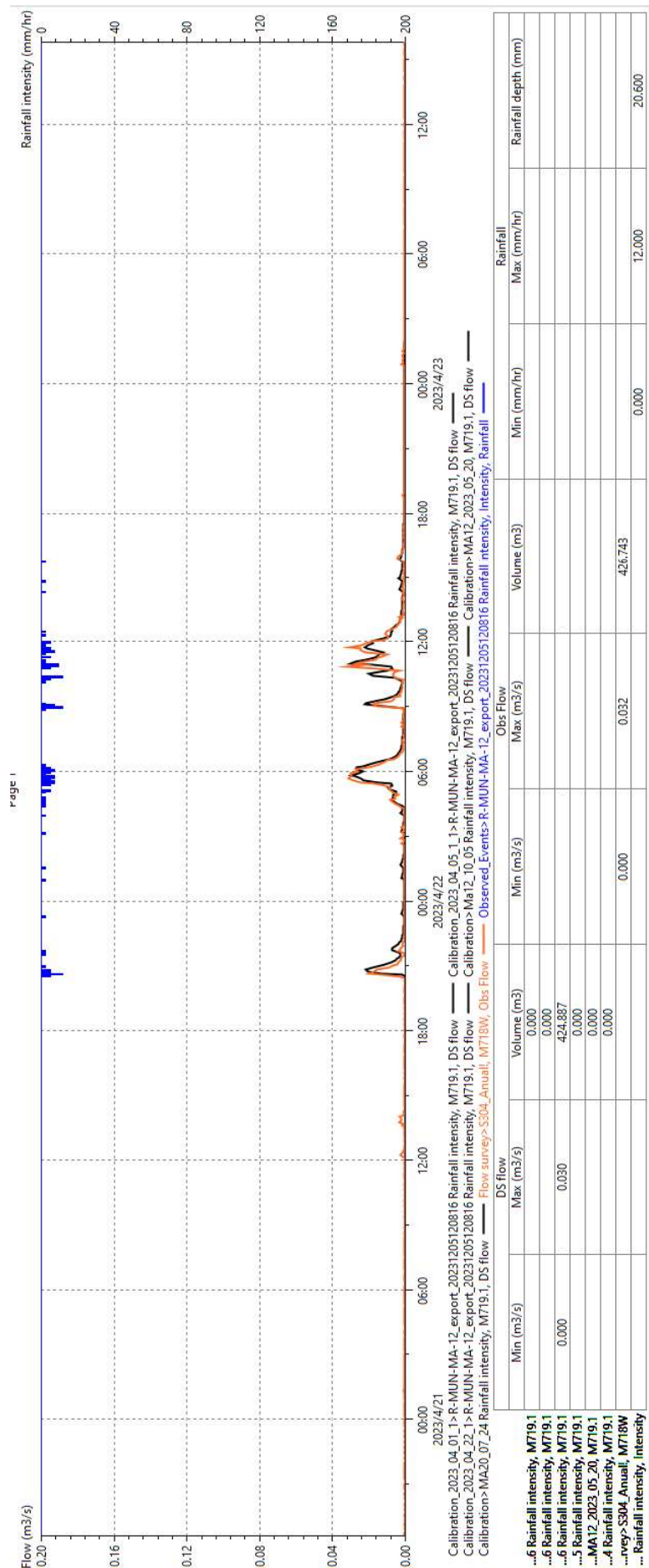
Swan Lake Flow Diversion Assessment

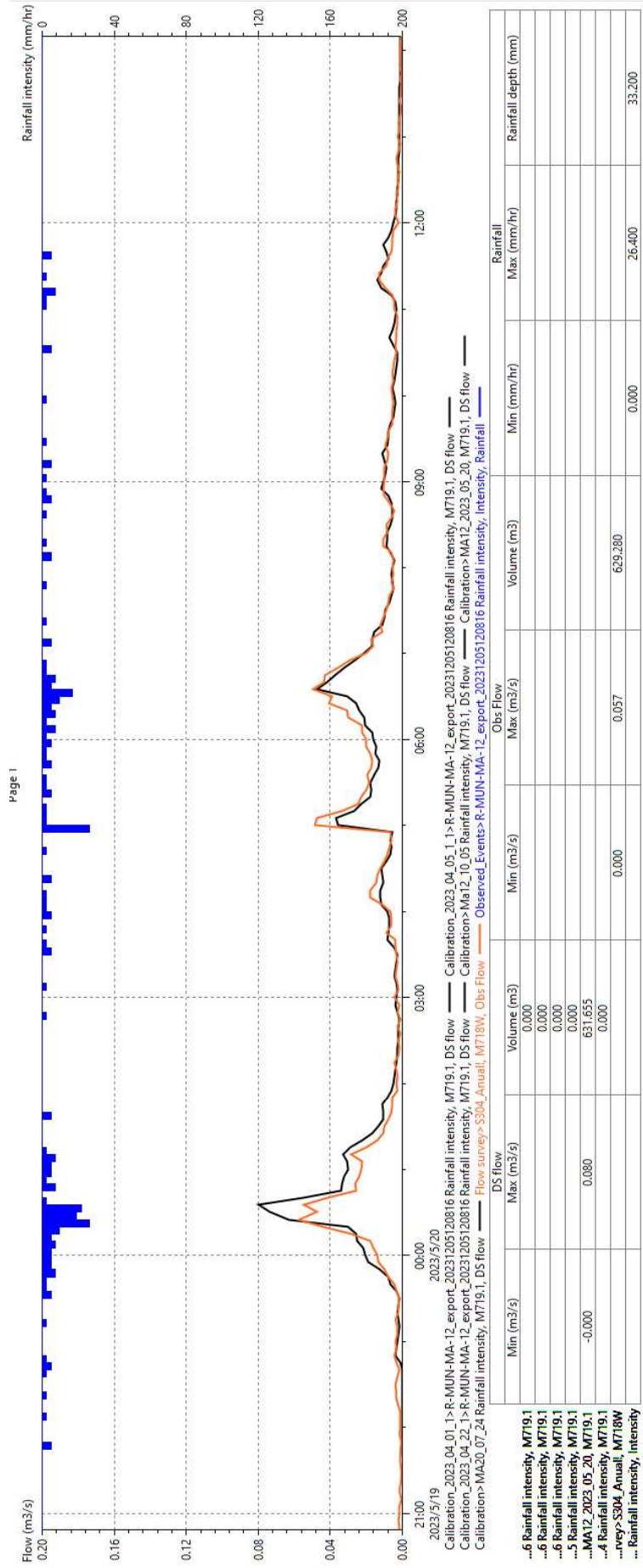


Gauge M719 Hydrograph comparison for Calibration events

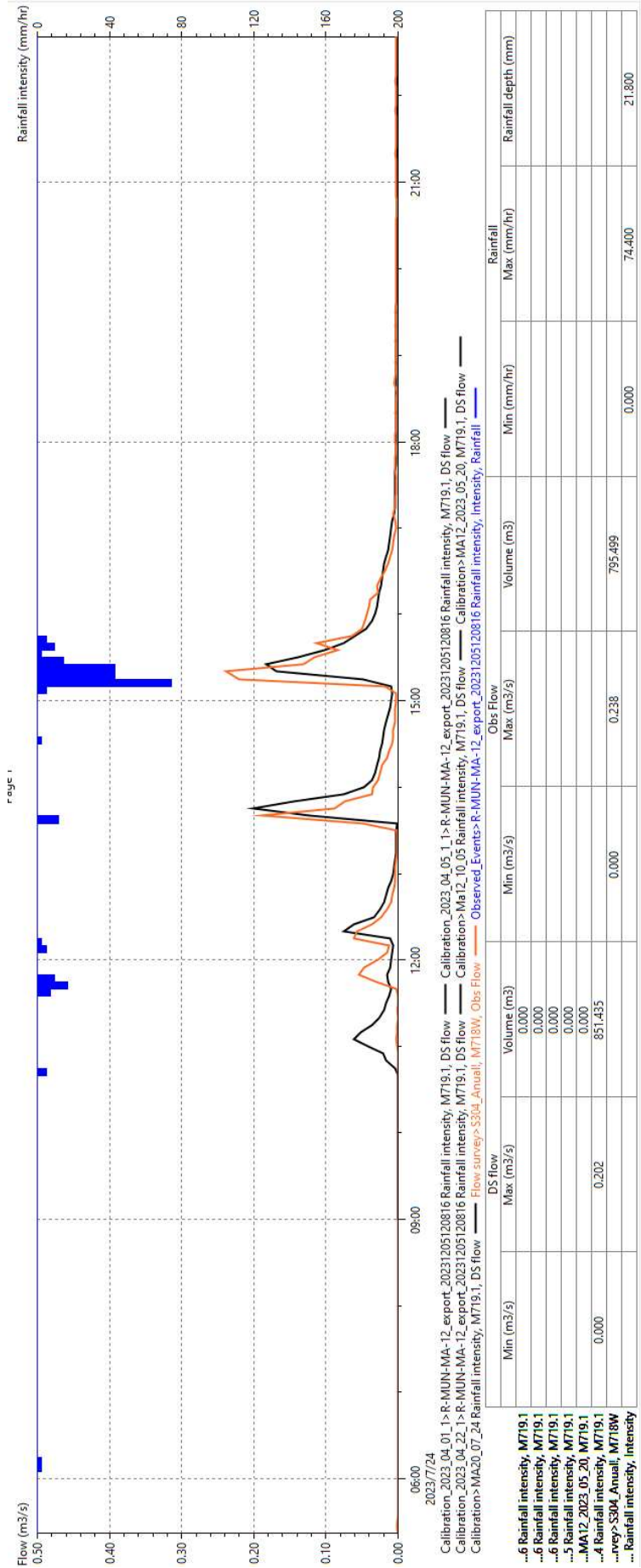


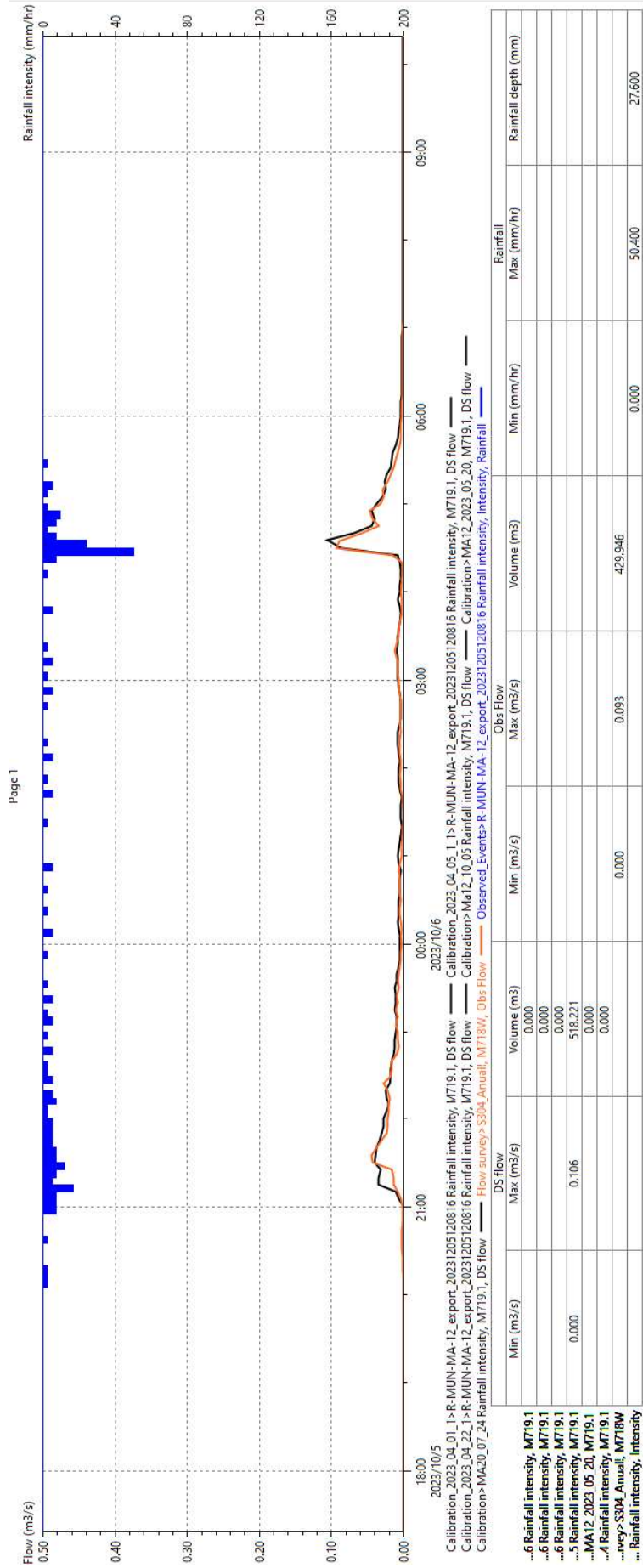






City of Markham
Consolidated Report
Swan Lake Flow Diversion Assessment





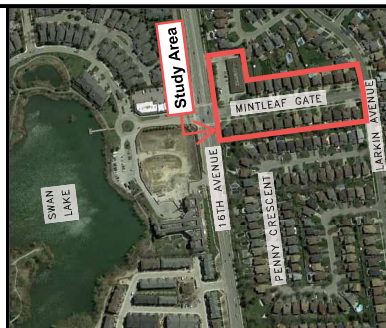


Appendix **B**

Mintleaf Gate Downspout Inspection Results



**SW DIVERSIONS
SWAN LAKE
(City of Markham)**



LEGEND

- Downspout discharging directly to the ground surface

- Downspout likely connected to some piping in the ground

17 House Number



AECOM

Mintleaf Gate,
Markham ON

Downspout Information

FIGURE 1
Sheet 1 of 3



**SW DIVERSIONS
SWAN LAKE
(City of Markham)**



LEGEND

- Downspout discharging directly to the ground surface

- Downspout likely connected to some piping in the ground

17 House Number



AECOM

Mintleaf Gate,
Markham ON

Downspout Information

FIGURE 1
Sheet 2 of 3







Appendix C

Cost Estimate Sheet



Scenario 1: Redirection of minor system flows from AMICA OGS and Swan Lake Blvd. OGS to sewers on 16th Ave.

Item	Part	Size/type	Depth	Quantity	Unit Cost	Cost of implementation	Contingency*	Totaling
Supply and install 375 mm diameter, 0-3 m deep, storm sewers	Part 1 - Storm sewers (unit m)	375 mm (Concrete)	0-3 m	32 m	\$1,230.00	\$39,360.00	\$11,808.00	\$51,168.00
Supply and install 450 mm diameter, 3-4.5 m deep, storm sewers	Part 1 - Storm sewers (unit m)	450 mm	3-4.5 m	113 m	\$1,550.00	\$175,150.00	\$52,545.00	\$227,695.00
Supply and install 600 mm diameter, 3-4.5 m deep, storm sewers	Part 1 - Storm sewers (unit m)	600 mm	3-4.5 m	76 m	\$1,700.00	\$129,200.00	\$38,760.00	\$167,960.00
Supply and install 750 mm diameter, 3-4.5 m deep, storm sewers	Part 1 - Storm sewers (unit m)	750 mm	3-4.5 m	95 m	\$2,000.00	\$190,000.00	\$57,000.00	\$247,000.00
Supply and install 750 mm diameter, 4.5-6 m deep, storm sewers	Part 1 - Storm sewers (unit m)	750 mm	4.5-6 m	94 m	\$2,150.00	\$202,100.00	\$60,630.00	\$262,730.00
Supply and install 825 mm diameter, 3-4.5 m deep, storm sewers	Part 1 - Storm sewers (unit m)	825 mm	3-4.5 m	33 m	\$2,650.00	\$87,450.00	\$26,235.00	\$113,685.00
Supply and install 825 mm diameter, 4.5-6 m deep, storm sewers	Part 1 - Storm sewers (unit m)	825 mm	4.5-6 m	95 m	\$2,700.00	\$256,500.00	\$76,950.00	\$333,450.00
Supply and install 1200 mm diameter, 4.5-6 m deep, storm sewers	Part 1 - Storm sewers (unit m)	1200 mm	4.5-6 m	328 m	\$3,690.00	\$1,210,320.00	\$363,096.00	\$1,573,416.00
Restoration of 0 - 3 m deep urban roads	Part 2 Pavement Restoration (Unit m)	Urban Road	0-3 m	32 m	\$1,690.00	\$54,080.00	\$16,224.00	\$70,304.00
Restoration of 3 - 4.5 m deep urban roads	Part 2 Pavement Restoration (Unit m)	Urban Road	3-4.5 m	317 m	\$1,955.00	\$619,735.00	\$185,920.50	\$805,655.50
Restoration of 4.5 - 6 m deep urban roads	Part 2 Pavement Restoration (Unit m)	Urban Road	4.5-6 m	517 m	\$2,450.00	\$1,266,650.00	\$379,995.00	\$1,646,645.00
Supply and install new maintenance holes and upsize existing maintenance holes to required sizes (minimal 600 mm larger than the largest connected pipe)	Part 3: Maintenance hole Installation and upgrade	30% of total pipe cost	-	-	-	\$687,024.00	\$206,107.20	\$893,131.20
N/A	Part 4: Flow Control Structures	-	-	-	-	\$0.00	\$0.00	\$0.00
Allowance for utility replacement and service reconnection	Part 5: Miscellaneous allowance	15% of total pipe cost	-	-	-	\$343,512.00	\$103,053.60	\$446,565.60
Allowance for catchbasin replacement and lateral reconnection	Part 5: Miscellaneous allowance	7.5% of total pipe cost	-	-	-	\$171,756.00	\$51,526.80	\$223,282.80
							Total	\$7,062,688.10

* 30% contingency accounts for the potential cost increase due to difficult site condition, design changes, market fluctuation, inflation, etc.

[illegible]

Scenario 3: Redirecting minor system flow from AMICA OGS and Swan Lake Blvd. OGS to the Lake outlet

Item	Part	Size/type	Depth	Unit	Unit Cost	Cost of implementation	Contingency	Totalling
Supply and install 525 mm diameter, 0-3 m deep, storm sewers	Part 1 - Storm Sewers (unit m)	525 mm	0-3 m	179	1680	\$300,720.00	\$90,216.00	\$390,936.00
Supply and install 525 mm diameter, 3-4.5 m deep, storm sewers	Part 1 - Storm Sewers (unit m)	525 mm	3-4.5 m	64	\$1,750.00	\$112,000.00	\$33,600.00	\$145,600.00
Supply and install 525 mm diameter, 4.5-6 m deep, storm sewers	Part 1 - Storm Sewers (unit m)	525 mm	4.5-6 m	54	\$1,970.00	\$106,380.00	\$31,914.00	\$138,294.00
Restoration of 0-3 m deep Urban Road	Part 2 Pavement Restoration (Unit m)	Urban Road	0-3 m	179	\$1,690.00	\$302,510.00	\$90,753.00	\$393,263.00
Restoration of 3-4.5 m deep Urban Road	Part 2 Pavement Restoration (Unit m)	Urban Road	3-4.5 m	64	\$1,955.00	\$125,120.00	\$37,536.00	\$162,656.00
Restoration of 4.5-6 m Deep Urban Road	Part 2 Pavement Restoration (Unit m)	Urban Road	4.5-6 m	54	\$2,450.00	\$132,300.00	\$39,690.00	\$171,990.00
Supply and install new maintenance hole s and upsize the existing maintenance hole to the require sizes (minimal 600 mm larger than the largest connected pipe)	Part 3: maintenance hole Installation and upgrade	30% of total pipe cost	-	-	-	\$155,730.00	\$46,719.00	\$202,449.00
-	Part 4: Flow Control Structures	-	-	-	-	\$0.00	\$0.00	\$0.00
							\$0.00	
Allowance for utility replacement and service reconnection	Part 5: Miscellaneous allowance	15% of total pipe cost	-	-	-	\$77,865.00	\$23,359.50	\$101,224.50
Allowance for catchbasin replacement and Lateral reconnection	Part 5: Miscellaneous allowance	7.5% of total pipe cost	-	-	-	\$38,932.50	\$11,679.75	\$50,612.25
								\$1,757,024.75

Scenario 4: Redirecting minor system flow from Swan Club OGS to the North Pond

[illegible]

Scenario 5: Adjusting the flow splitter weir for the East Pond and North Pond to reduce flow bypass to the Lake

Item	Part	Size/type	Depth	Quantity	Unit Cost	Cost of implementation	Contingency	Totaling
N/A	Part 1 - Storm Sewers (unit m)				\$0.00	\$0.00	\$0.00	\$0.00
N/A	Part 2 Pavement Restoration (Unit m)				\$0.00	\$0.00	\$0.00	\$0.00
N/A	Part 3: Maintenance hole Installation and upgrade				\$0.00	\$0.00	\$0.00	\$0.00
Supply and install precast concrete overflow weirs	Part 4: Flow Control Structures	-	-	3	\$7,800.00	\$23,400.00	\$7,020.00	\$30,420.00
N/A	Part 5: Miscellaneous allowance				\$0.00	\$0.00	\$0.00	\$0.00
							Total:	\$30,420.00

Scenario 5b: Adjusting the flow splitter weir and upsizing the pond inlet pipe

[illegible]

S6 Pond Expansion

Item	Part	Size/type	Depth	Quantity	Unit Cost	Cost of implementation	Contingency	Totaling
Supply and install 600 mm diameter, 0-3 m deep, storm sewers	Part 1 - Storm Sewers (unit m)	600 mm	0-3 m	28	\$1,700.00	\$47,600.00	\$14,280.00	\$61,880.00
Restoration of 0-3 m Deep Open Space	Part 2 Pavement Restoration (Unit m)	Open Space	0-3 m	-	-	\$0.00	\$0.00	\$0.00
Supply and install new maintenance hole s and upsize the existing maintenance hole to the require sizes (minimal 600 mm more than the largest connected pipe)	Part 3: maintenance hole Installation and upgrade	30% of total pipe cost	-	-	-	\$14,280.00	\$4,284.00	\$18,564.00
Supply and install pre-casted concrete overflow weirs	Part 4: Flow Control Structures	-	-	3	\$7,800.00	\$23,400.00	\$7,020.00	\$30,420.00
Earth work (Cut and Fill) (Unit m3)	Part 4: Flow Control Structures	Wet Pond		14000	\$120.00	\$1,680,000.00	\$504,000.00	\$2,184,000.00
Allowance for utility replacement and service reconnection	Part 5: Miscellaneous allowance	15% of total pipe cost	-	-	-	\$7,140.00	\$2,142.00	\$9,282.00
Allowance for catchbasin replacement and lateral reconnection	Part 5: Miscellaneous allowance	7.5% of total pipe cost	-	-	-	\$3,570.00	\$1,071.00	\$4,641.00
Allowance for equipment mobilization, demobilization, sediment testing, payment for bonds	Part 5: Miscellaneous allowance	30% of earth work cost				\$504,000.00	\$151,200.00	\$655,200.00
							Total	\$2,963,987.00

S6b Pond Expansion (North Pond Portion)

Item	Part	Size/type	Depth	Quantity	Unit Cost	Cost of implementation	Contingency	Totaling
Supply and Install 600 mm diameter, 0-3 m deep, storm sewers	Part 1 - Storm Sewers (unit m)	600 mm	0-3 m	9	\$1,700.00	\$15,300.00	\$4,590.00	\$19,890.00
Restoration of 0-3 m Deep Open Space	Part 2 Pavement Restoration (Unit m)	Open Space	0-3 m	-	-	\$0.00	\$0.00	\$0.00
Supply and install new maintenance hole s and upsize the existing maintenance hole to the require sizes (minimal 600 mm more than the largest connected pipe)	Part 3: maintenance hole Installation and upgrade	30% of total pipe cost	-	-	-	\$4,590.00	\$1,377.00	\$5,967.00
Supply and install pre-casted concrete overflow weirs	Part 4: Flow Control Structures	-	-	1	\$7,800.00	\$7,800.00	\$2,340.00	\$10,140.00
Earth work (Cut and Fill) (Unit m3)	Part 4: Flow Control Structures	Wet Pond	-	8000	\$120.00	\$960,000.00	\$288,000.00	\$1,248,000.00
Allowance for utility replacement and service reconnection	Part 5: Miscellaneous allowance	15% of total pipe cost	-	-	-	\$2,295.00	\$688.50	\$2,983.50
Allowance for catchbasin replacement and Lateral reconnection	Part 5: Miscellaneous allowance	7.5% of total pipe cost	-	-	-	\$1,147.50	\$344.25	\$1,491.75
Allowance for equipment mobilization, demobilization, sediment testing, payment for bonds	Part 5: Miscellaneous allowance	30% of earth work cost	-	-	-	\$288,000.00	\$86,400.00	\$374,400.00
							Total	\$1,662,872.25

S6c Pond Expansion (East Pond portion)

Item	Part	Size/type	Depth	Quantity	Unit Cost	Cost of implementation	Contingency	Totalling
Supply and install 600 mm diameter, 0-3 m deep, storm sewers	Part 1 - Storm Sewers (unit m)	600 mm	0-3 m	19	\$1,700.00	\$32,300.00	\$9,690.00	\$41,990.00
Restoration of 0-3 m Deep Open Space	Part 2 Pavement Restoration (Unit m)	Open Space	0-3 m	-	-	\$0.00	\$0.00	\$0.00
Supply and install new maintenance hole s and upsiz the existing maintenance hole to the require sizes (minimal 600 mm more than the largest connected pipe)	Part 3: maintenance hole Installation and upgrade	30% of total pipe cost	-	-	-	\$9,690.00	\$2,907.00	\$12,597.00
Supply and install pre-casted concrete overflow weirs	Part 4: Flow Control Structures	-	-	2	\$7,800.00	\$15,600.00	\$4,680.00	\$20,280.00
Earth work (Cut and Fill) (Unit m3)	Part 4: Flow Control Structures	Wet Pond		6000	\$120.00	\$720,000.00	\$216,000.00	\$936,000.00
Allowance for utility replacement and service reconnection	Part 5: Miscellaneous allowance	15% of total pipe cost	-	-	-	\$4,845.00	\$1,453.50	\$6,298.50
Allowance for catchbasin replacement and lateral reconnection	Part 5: Miscellaneous allowance	7.5% of total pipe cost	-	-	-	\$2,422.50	\$726.75	\$3,149.25
Allowance for equipment mobilization, demobilization, sediment testing, payment for bonds	Part 5: Miscellaneous allowance	30% of earth work cost				\$216,000.00	\$64,800.00	\$280,800.00
							Total	\$1,301,114.75

[illegible]

[illegible]

S7c Combine Scenario S4 (Redirecting minor system flow from Swan Club OGS to the North Pond) with S6 (Pond Expansion)

Item	Part	Size/type	Depth	Quantity	Unit Cost	Cost of implementation	Contingency	Totaling
Supply and Install 600 mm diameter, 0-3 m deep, storm sewers	Part 1 - Storm Sewers (unit m)	600 mm	0-3 m	9	\$1,700.00	\$15,300.00	\$4,590.00	\$19,890.00
Supply and Install 300 mm diameter, 0-3 m deep, storm sewers	Part 1 - Storm Sewers (unit m)	300 mm	0-3 m	111	\$1,250.00	\$138,750.00	\$41,625.00	\$180,375.00
Restoration of 0-3 m Deep Open Space	Part 2 Pavement Restoration (Unit m)	Open Space	0-3 m	-	-	\$0.00	\$0.00	\$0.00
Supply and install new maintenance hole s and upsize the existing maintenance hole to the require sizes (minimal 60mm more than the largest connected pipe)	Part 3: maintenance hole Installation and upgrade	30% of total pipe cost	-	-	-	\$46,215.00	\$13,864.50	\$60,079.50
Supply and install pre-casted concrete overflow weirs	Part 4: Flow Control Structures	-	-	1	\$7,800.00	\$7,800.00	\$2,340.00	\$10,140.00
Earth work (Cut and Fill) (Unit m3)	Part 4: Flow Control Structures	-		8000	\$120.00	\$960,000.00	\$288,000.00	\$1,248,000.00
Allowance for utility replacement and service reconnection	Part 5: Miscellaneous allowance	15% of total pipe cost	-	-	-	\$2,295.00	\$688.50	\$2,983.50
Allowance for catchbasin replacement and Lateral reconnection	Part 5: Miscellaneous allowance	7.5% of total pipe cost	-	-	-	\$1,147.50	\$344.25	\$1,491.75
Allowance for equipment mobilization, demobilization, sediment testing, payment for bonds	Part 5: Miscellaneous allowance	30% of total earth work cost				\$288,000.00	\$86,400.00	\$374,400.00
							Total	\$1,897,359.75

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Swan Lake

Annual Meeting with Markham Subcommittee

Environmental Services

Authors: Robert Muir, Manager, Stormwater

Zahra Parhizgari, Environmental Engineer, Stormwater

June 18, 2025

Agenda

- Background
- Completed Work
- Lake Conditions
- Findings of Flow Diversion Study
- 2025 Plan and Recommendations
- Parks Operation Updates

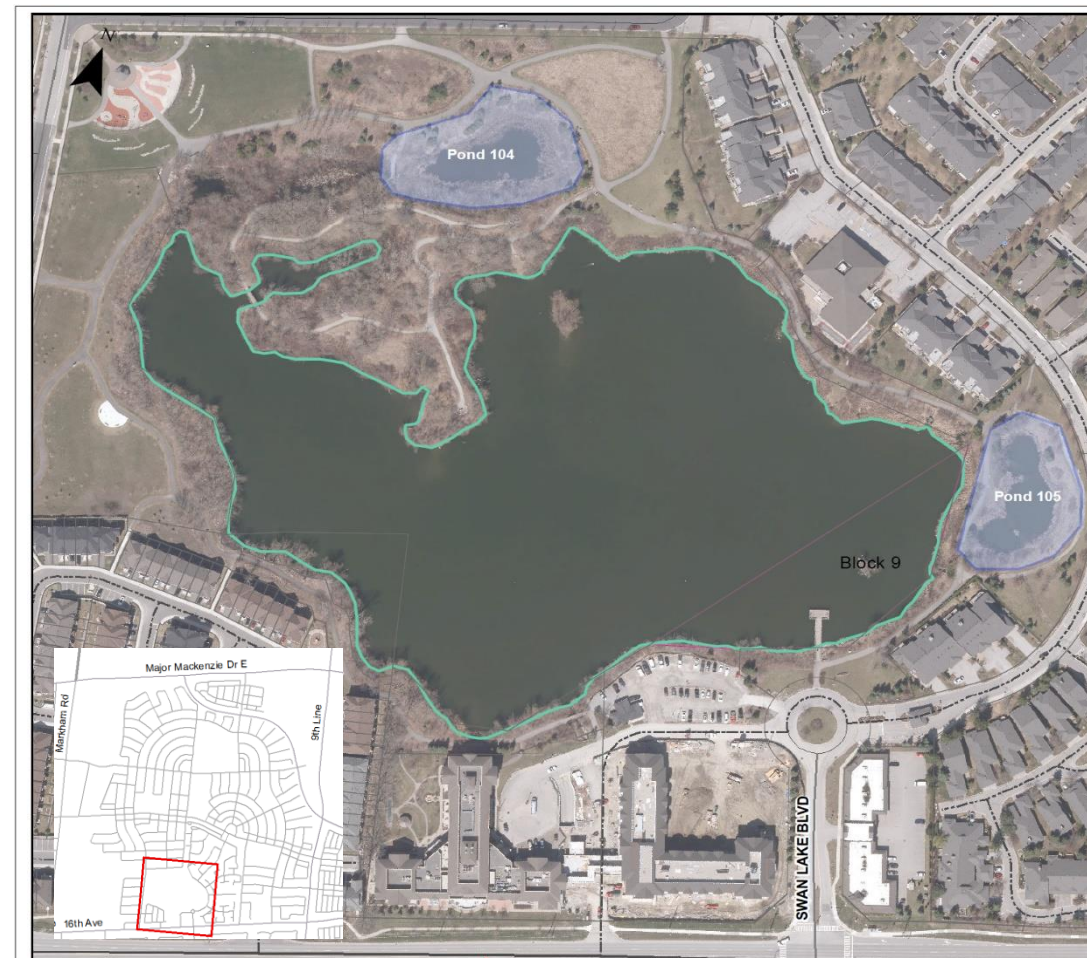




Background

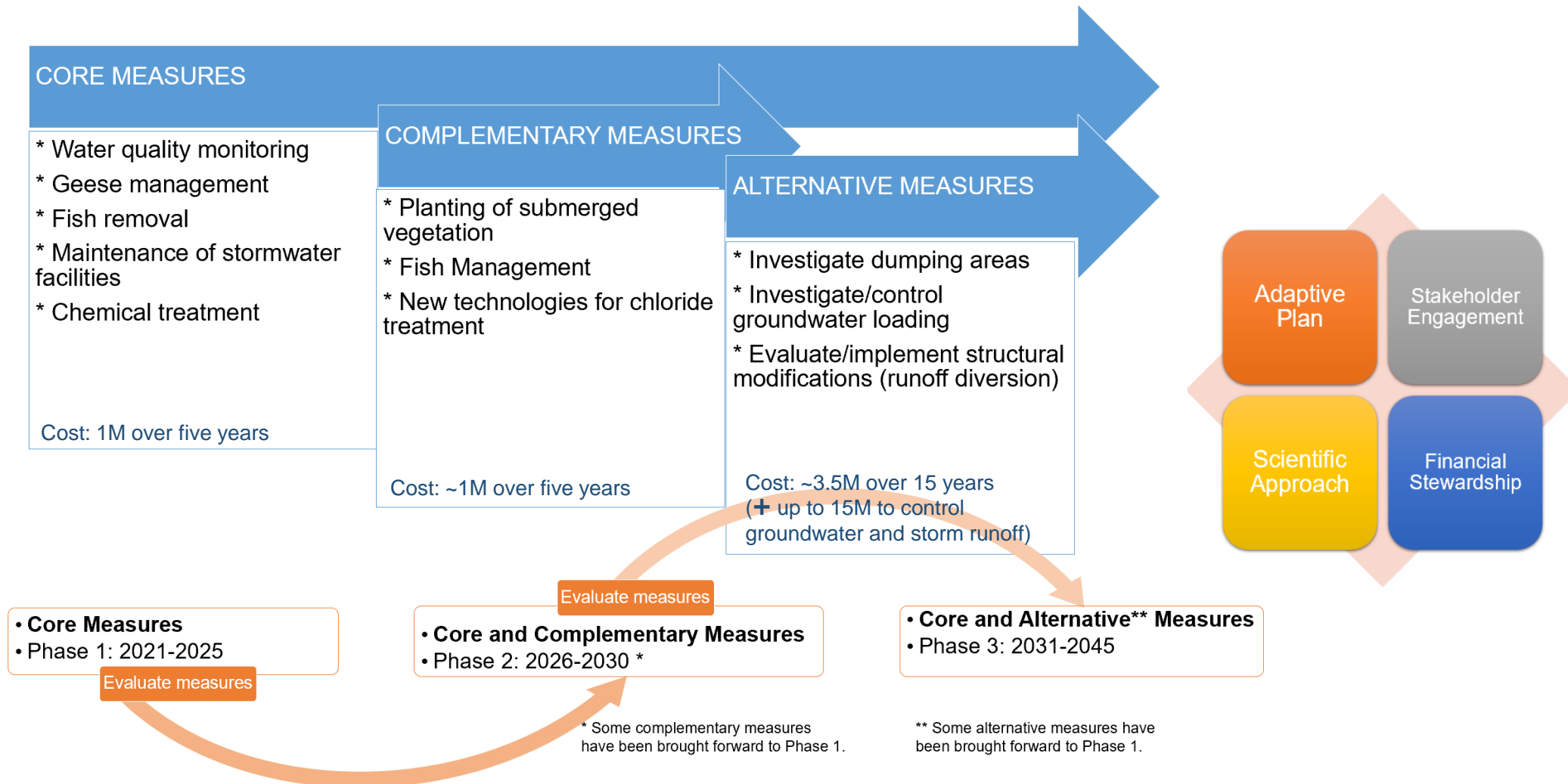
Location and History

- Operated in the 1960s and 1970s as a gravel pit
- Construction waste dump in early 1980s
- Lake created when gravel pit operations stopped
- Drainage area is fully developed with stormwater ponds and oil/grit separators
- Closed system and prone to nutrient build-up and algae growth.
- Winter maintenance introduces chloride to the lake
- Water quality issues noticed since 2010 or earlier
- Active management since 2013, including Phoslock treatment, geese management, and monitoring





Long Term Management Plan (2021)





2024 Council Resolutions

That the following motion passed at the July 29, 2024, Markham Sub-Committee meeting be received for information purposes:

1. That the minutes of the July 29, 2024 Markham Sub-Committee meeting be received for information purposes; and,
2. That the report entitled “Swan Lake- 2024 Water Quality Status and Updates” be received; and,
3. That the deputations from Fred Peters, Friends of Swan Lake Park, Ali Asgary and Satinder Brar, York University, Peter Miller, William Dewberry, and Pamela Nitert, Amica Swan Lake, made to the July 29, 2024 Markham Sub-Committee be received; and,
4. That the funding request by Friends of Swan Lake Park and the CIFAL proposal, as well as the request for shoreline viewing nodes be referred to Staff to report back in the future; and,
5. That Staff continue to implement the Long-term Management Plan for Swan Lake approved by Council in December 2021, including advancement of submerged aquatic vegetation, research into chloride treatment, and flow diversion evaluation (previously in Phases 2 and 3 of the Plan); and,
6. That Staff report back annually on water quality results and evaluation of adapted Core and Complementary measures for consideration in Phase 2 of the Plan through the Markham Sub-Committee with the participation of the Friends of Swan Lake Park; and,
7. That the next review of the Plan will be in 2025 (after the completion of Phase 1 and other measures as listed under item 2) with consideration for a workshop in 2026; and further,
8. That Staff be authorized and directed to do all things necessary to give effect to this resolution.



Completed Work



List of 2024 Activities

Activity	Phase 1 Core Measures (Years 1-5)
Water Quality Monitoring (and annual reporting to Subcommittee)	<input checked="" type="checkbox"/>
Geese Management	<input checked="" type="checkbox"/>
Chemical Treatment	<input checked="" type="checkbox"/>
Fish Management (removal of benthic fish and fish stocking)	<input checked="" type="checkbox"/> (fish stocking in 2025)
Planting of Submerged Plants *	<input checked="" type="checkbox"/>
Maintenance of stormwater management facilities	<input checked="" type="checkbox"/>
Community Engagement	<input checked="" type="checkbox"/>
Flow Diversion Feasibility Study **	<input checked="" type="checkbox"/> (completed recently)
New Technologies for Chloride Treatment *	<input checked="" type="checkbox"/> underway
Phase 1 Review	<input checked="" type="checkbox"/> underway

* Originally planned for Phase 2

** Originally planned for Phase 3

Water Quality Monitoring

Understanding Issues and Planning Solutions

- Chemistry and biology
- Water level
- City staff measurements and Trent University data
- Accredited labs analyze the samples
- Regular checks and observations at the site
- External Specialists review



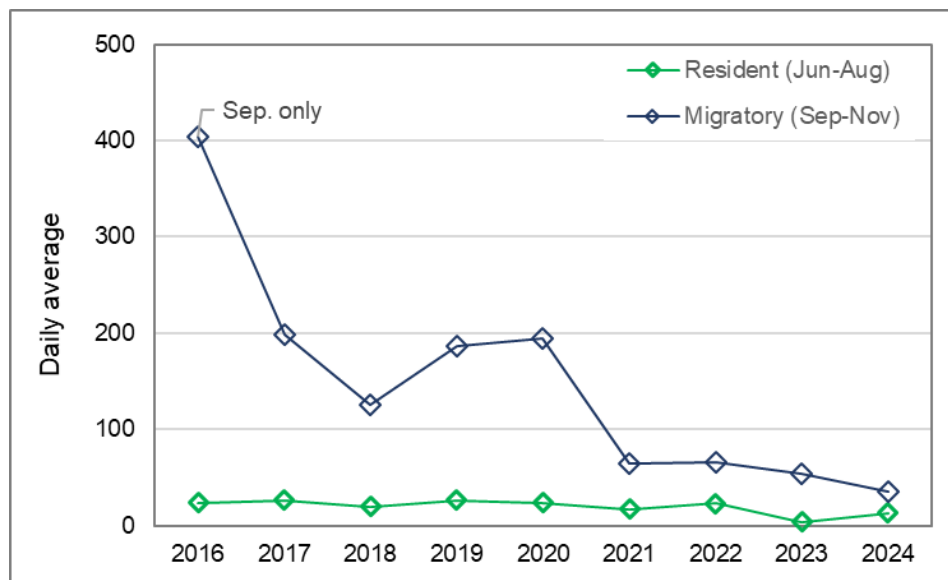


Geese Management

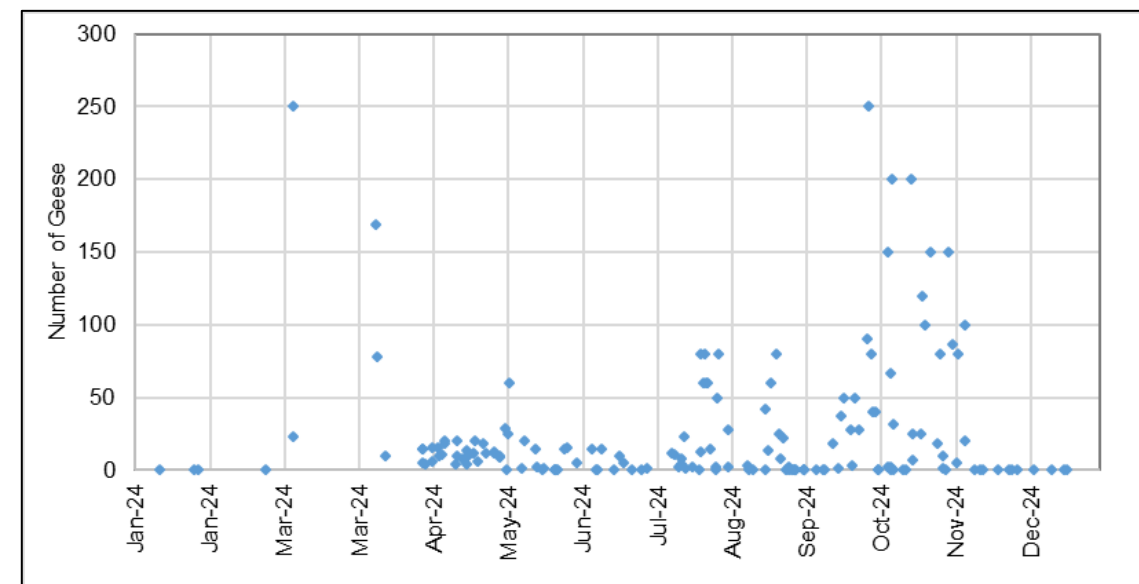
- Nest depredation, laser light, avian distress call and limited strategic zinc crackler pyro
- Geese relocation
- No evidence of strobe lights being effective



Swan Lake Geese Count Survey QR Code



* Some assumptions have been made in calculating the daily average for each year to fill in data gaps.



Chemical Treatment

- Second treatment on June 17-25, 2024
- Nine tonnes of PAC (a substance that reduces phosphorus and algae) was added to the lake.
- Each treatment was followed by 1-2 days of rest for testing and floc formation.
- Water clarity improved to 1 meter and stayed above 0.5 meters until late November.



Fish Management

- Removing bottom-dwelling fish to prevent sediment disturbance.
- Fish in the Lake included Common Carp, Brown Bullhead, and Fathead Minnow.
- Fish stocking completed in May 2025 through consultation with TRCA, MNRF, and by a private contractor.
- 500 juvenile largemouth bass added to the lake.
- Bluegill may be added later, depending on availability at MNRF facility.



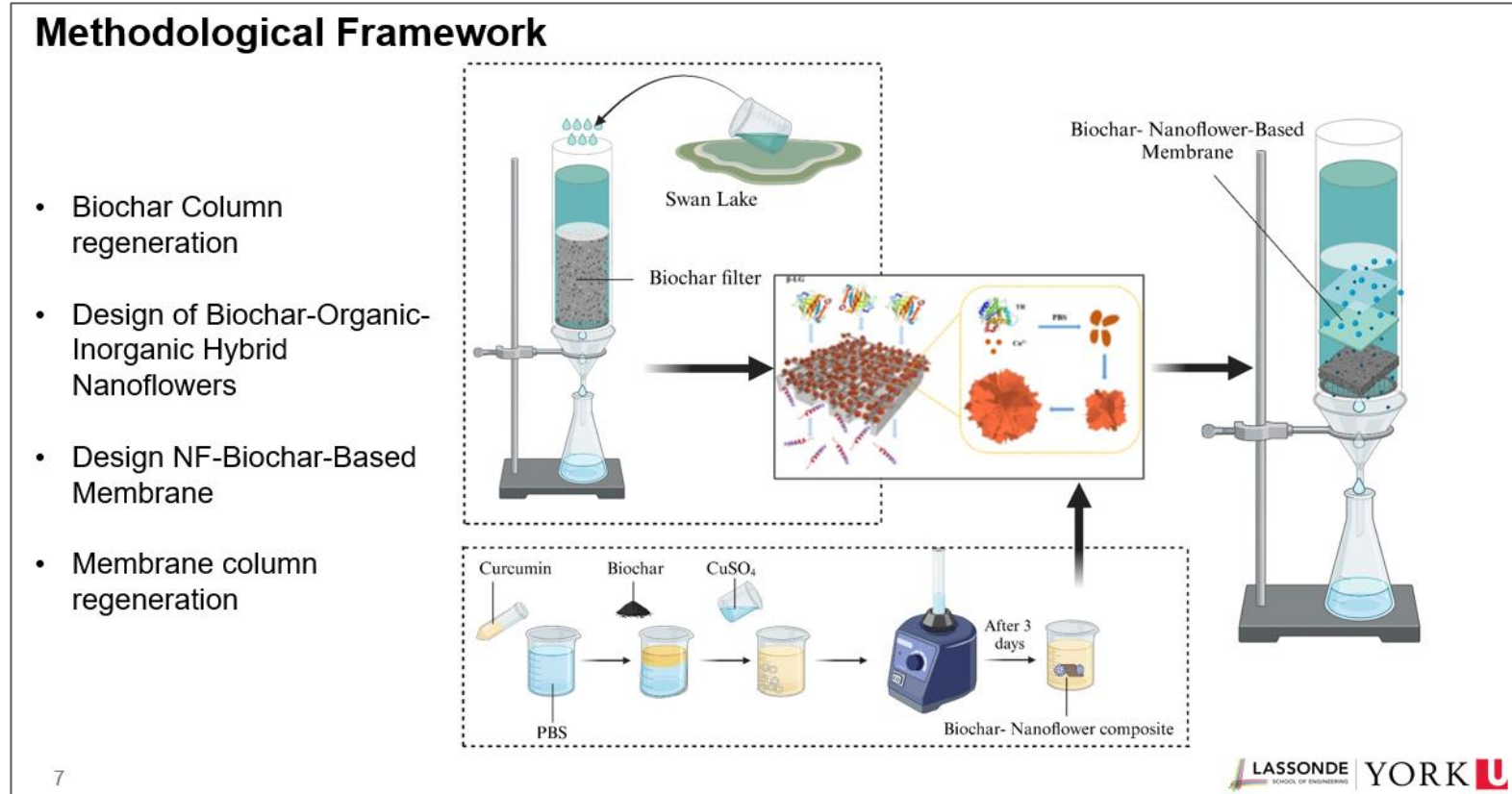
Planting of Submerged Aquatic Vegetation

- Submerged aquatic vegetation (macrophytes) can compete with and help mitigate algae (phytoplankton) growth
- Macrophytes will increase water clarity, which in turn, enhances their own growing conditions.
- TRCA planted 3000 stems of wild celery in fenced locations on the north site in 2023 and 2024
- Naturally growing aquatic plants were also abundant in 2024
- In 2025, existing plants will be monitored for survival and natural propagation
- Further SAV planting will be assessed through the five-year review process



Assessment of New Technologies for Chloride Treatment

- Initially planned for 2027; advanced to 2024 at the request of FOSLP
- Lab-scale testing of a 'Biochar-nanoflower-based Column Filtration System' for removing chloride from Lake water- proposed and implemented by York University
- Work underway, expected to be completed by end of 2025





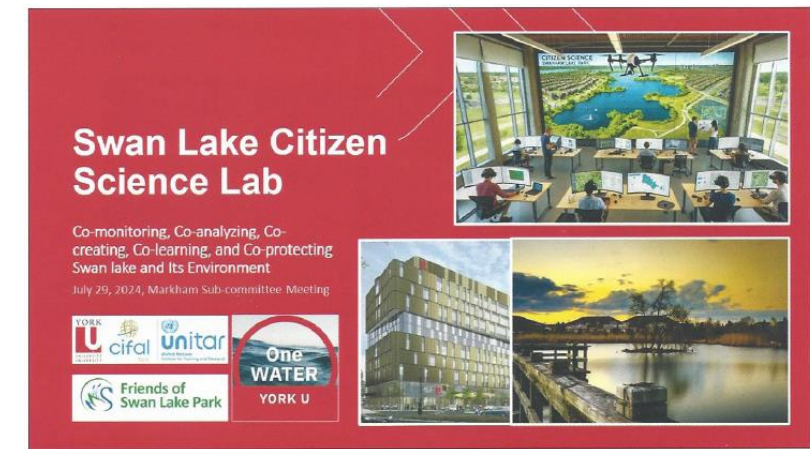
Maintenance of Stormwater Management Facilities

- Stormwater pond assumption process underway

Community Engagement – Water Quality Monitoring

Council directed staff to review CIFAL York proposal on Swan Lake Citizen Science Lab:

- Proposal reviewed by staff and Dr. Karst-Riddoch, Principal Aquatic Scientist at AECOM
- Drone image would only provide qualitative imagery, while City (and Trent University) collect quantitative measurements
- Swan Lake is small enough for visual monitoring; drones are more useful in large systems
- There is no need for 'co-monitoring' of water quality
- The City does not regulate drone use and image disclosure- York and FOSLP to follow applicable regulations and avoid disturbing park use by residents and wildlife
- *Other objectives of CIFAL proposal outside the scope of Environmental Services*





Community Engagement- Funding Request

Council directed staff to review FOSLP's funding request:

- FOSLP requested funding support to hire environmental consultants to 'advise on the rehabilitation of both Swan Lake and Swan Lake Park'
- The City has allocated funding in 2025 for an external consultant to review the water quality program
- Additional fundraising is not needed to move the program review forward
- The City is investing significant funds, and adapting measures based on the approved plan
- Research and technical studies originally planned for later phases are already underway, as requested by FOSLP.

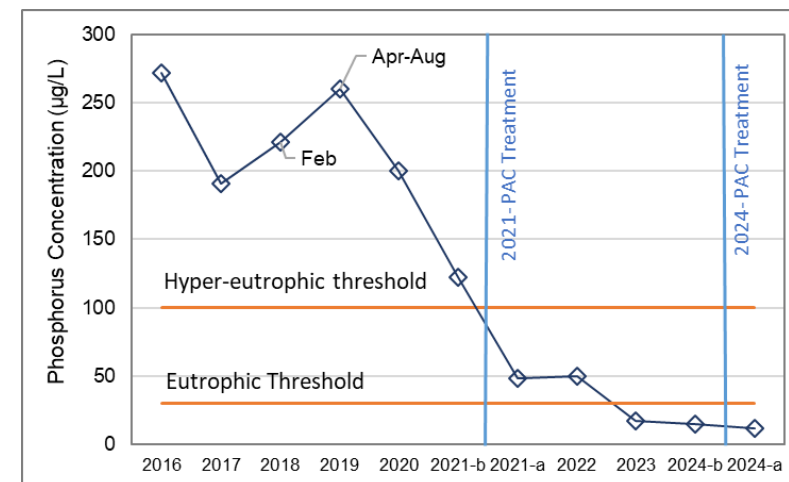


Lake Conditions

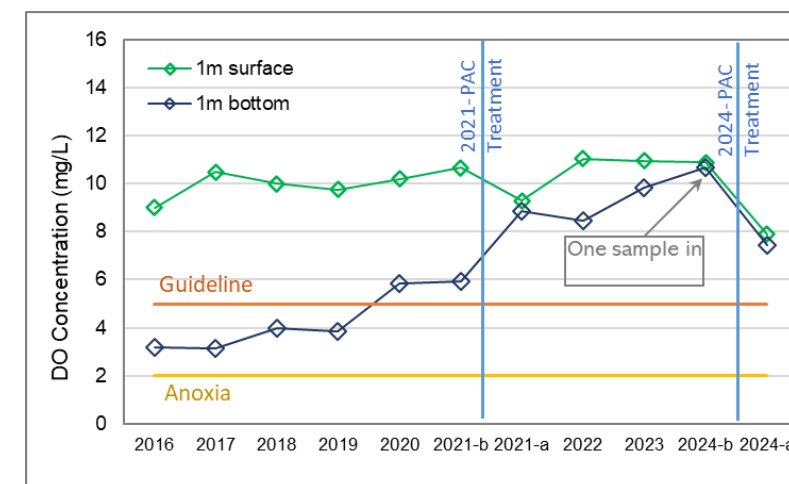


Water Quality- Nutrients and Oxygen

- Total Phosphorus:
 - Average under 30 µg/L during growing season (below the threshold for eutrophic condition)
 - Decreased significantly after each treatment
- Total Nitrogen:
 - Average below 0.65 mg/L during growing season (below the threshold for eutrophic condition)
 - Dominant forms not bioavailable
- Dissolved Oxygen:
 - Surface concentration > 6.4 mg/L all year
 - Bottom concentration showed a decline during dry months of August and September about 50% time
 - Data do not indicate anoxia-driven phosphorus enrichment



Phosphorus



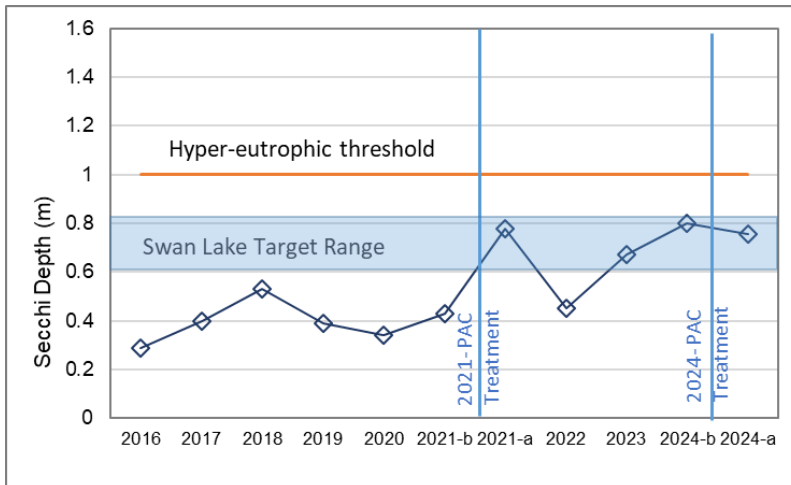
Dissolved Oxygen

Water Quality- Algae and Clarity

- Algal growth:
 - Surface bloom not occurring since treatment
 - 2024 Cyanobacteria cell numbers 40% lower than 2023
 - Chlorophyll-a within eutrophic state
- Clarity:
 - Above >0.6 m (target), except occasional decline to 0.5 m in fall
 - Growing-season average within target



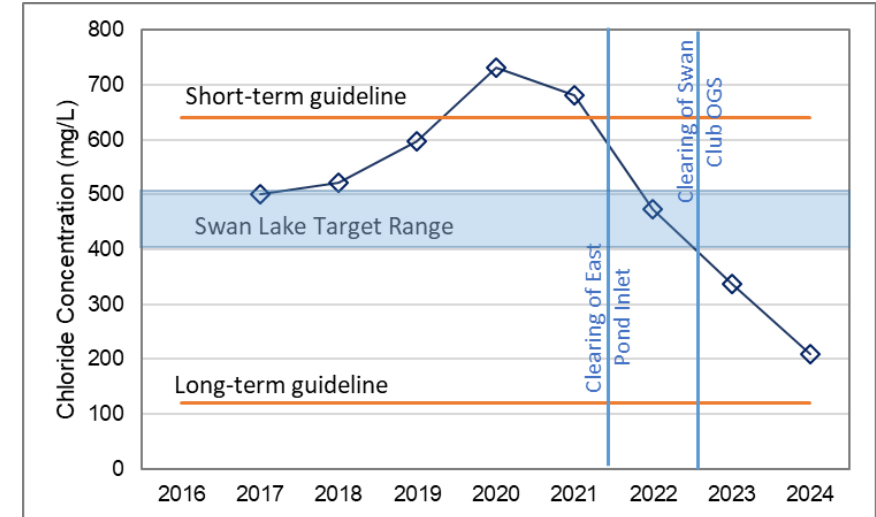
Algal bloom before treatment
 (photo from July 2020)



No surface bloom since treatment (photo from July 2024)

Water Quality- Chloride

- Chloride enters the Lake through winter maintenance:
 - Swan Lake Village Corporation
 - City roads
 - Residents north of the Lake
 - AMICA Corporation to the south
- Previously on upward trend likely due to blockages, which resulted in untreated flows to the Lake
- Since 2021 decreased due to clearing the blockage at the East Pond inlet and Swan Club OGS
- The presence of minnows in large numbers indicate chloride concentrations are not negatively impacting aquatic life

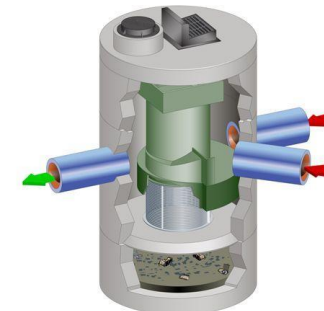
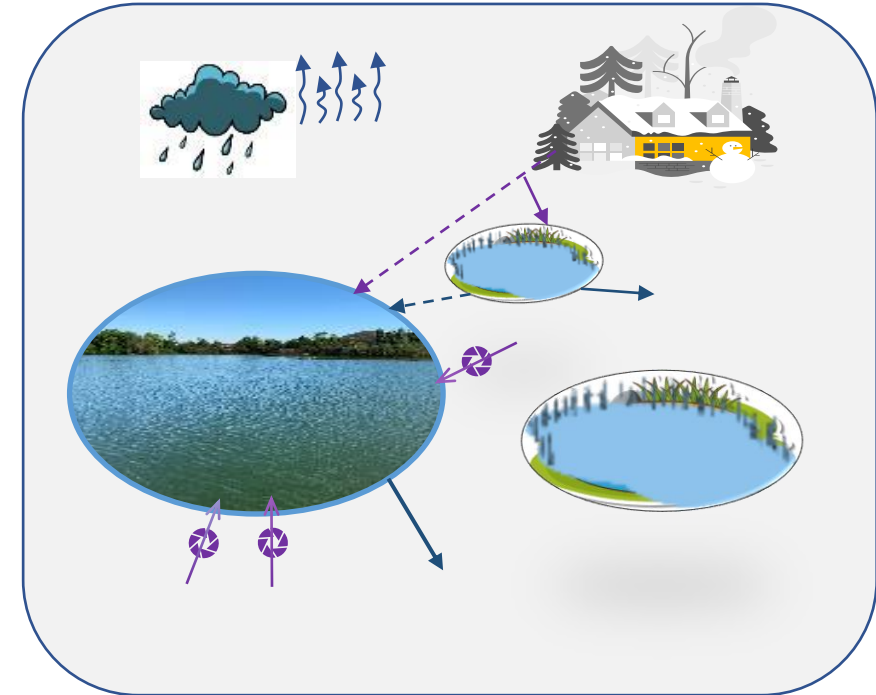




Findings of Flow Diversion Study

Flow Diversion Feasibility Study

- Alternative measure in Phase 3; brought forward at the request of FOSLP
- Stormwater runoff from the catchment area contains nutrients, metals, chloride and other constituents
- Stormwater management is accomplished using:
 - Ponds: store water and release gradually, treat water by sedimentation, infiltration, plant uptake (mostly solids- limited efficiency for chloride)
 - Oil and Grit Separators (OGS): no storage, treat water by sedimentation (lower efficiency than ponds; mostly solids- limited efficiency for chloride)
- Flow diversion involves redirecting all or part of runoff from the lake (after or before treatment in ponds and/or OGSs)

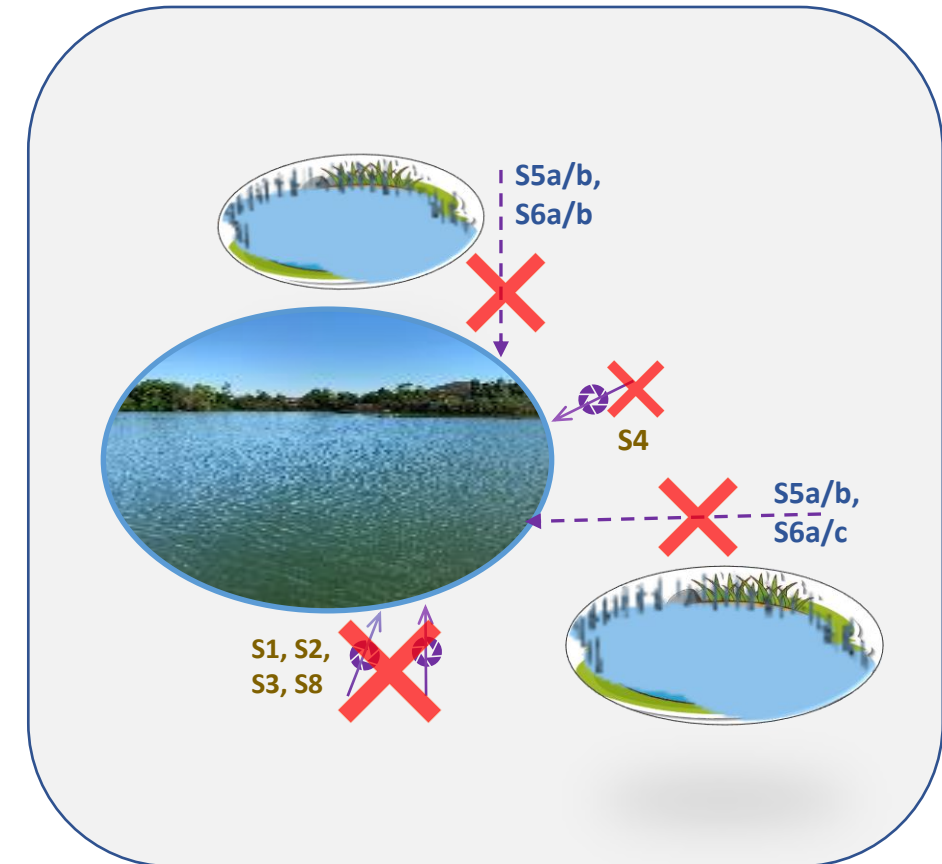




Flow Diversion Scenarios

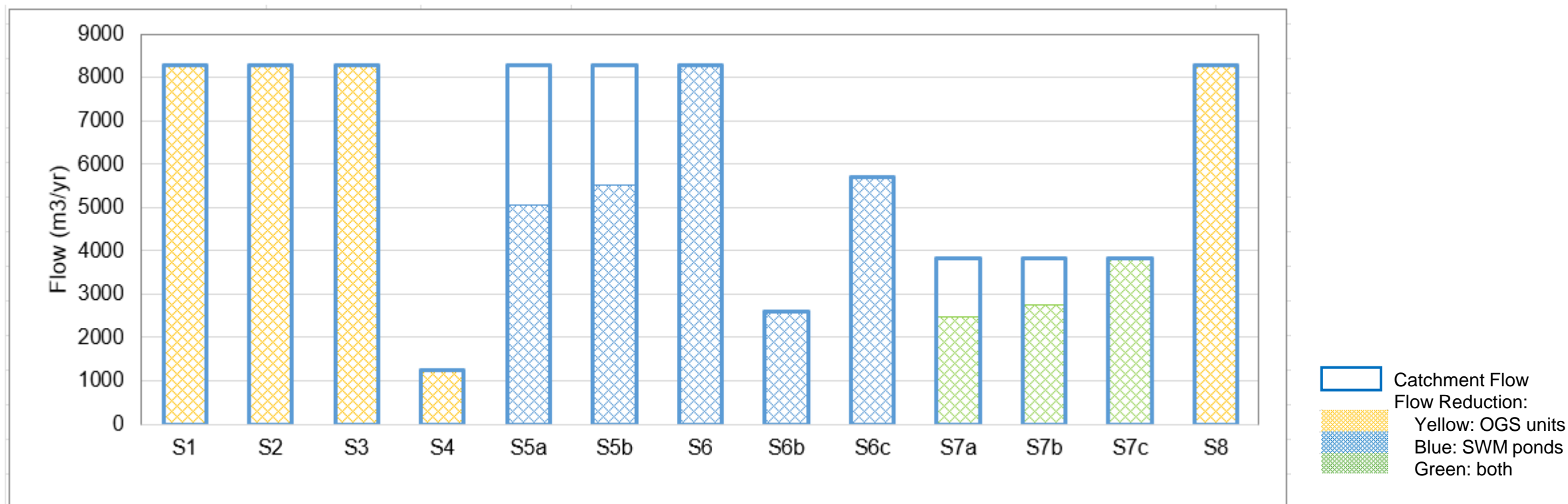
S1	Divert minor system flow Amica and Swan Lake Blvd. OGSs to 16t Ave.
S2	Divert only "First Flush" from Amica and Swan Lake Blvd. OGSs to 16th Ave.
S3	Divert minor system flow Amica and Swan Lake Blvd. OGSs to Lake outlet
S4	Divert minor system flow from Swan Club OGS to the North Pond
S5a	Raise the flow split weir at the North and East Ponds
S5b	Raise the flow split weir at the North and East Ponds while upsizing the inflow pipes
S6a	Expand storage capacity in north and east pond
S6b	North pond portion of S6
S6c	East pond portion of S6
S7a	Combine S4 with raising north pond weir (5a)
S7b	Combine S4 with raising north pond weir and upsize pond (5b)
S7c	Combine S4 with S6b
S8	Divert minor system flow Amica and Swan Lake Blvd. OGSs to underground storage

Yellow: OGS units
Blue: SWM ponds
Green: both





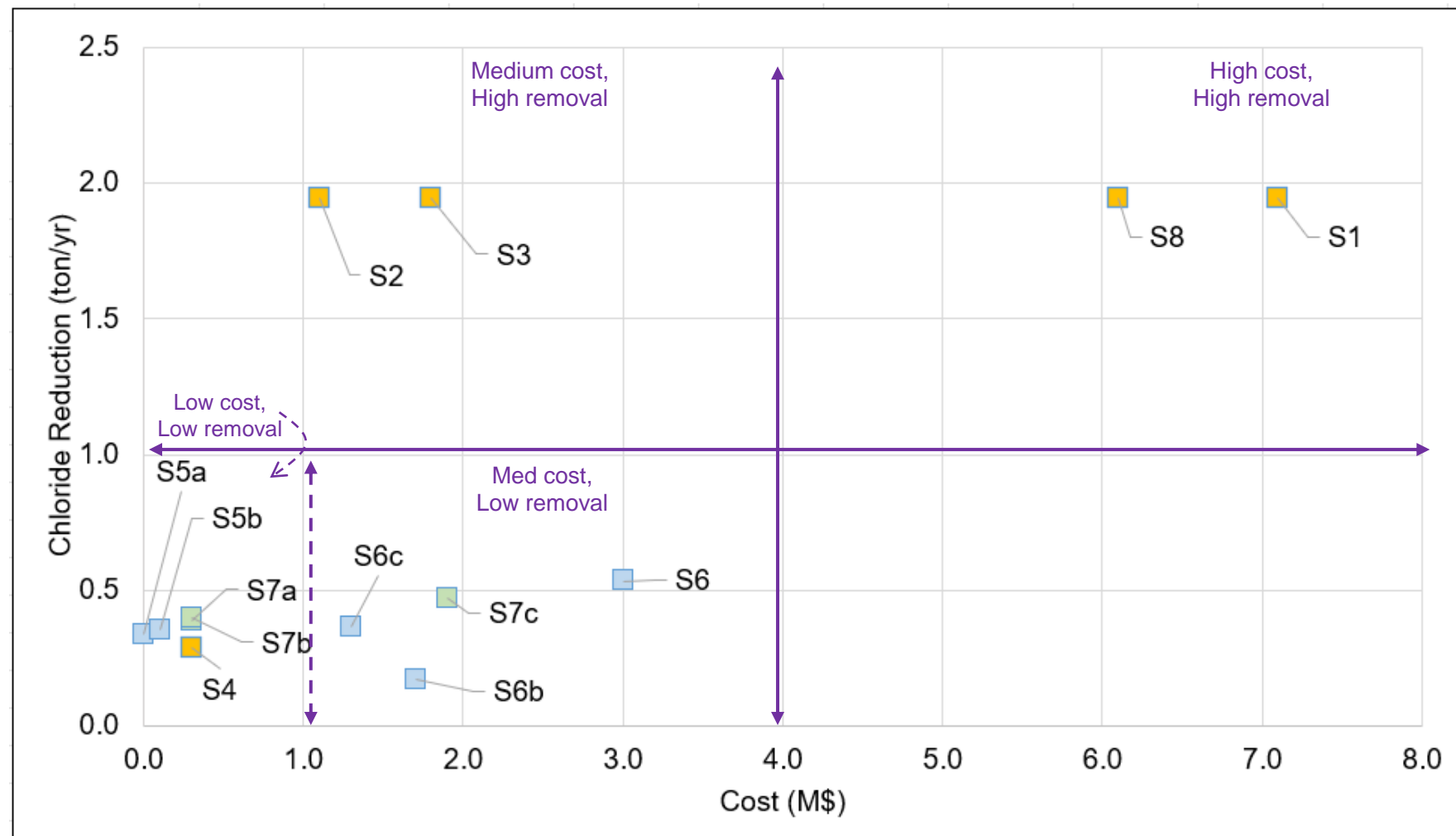
Flow Reduction Impact





Chloride Reduction Impact and Costs

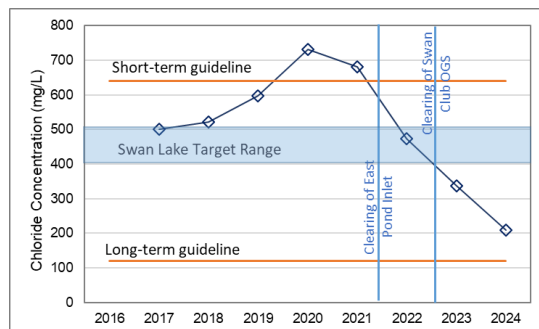
- Redirecting OGS flows achieves the highest chloride reduction at high cost
- Pond expansion is less costly but achieves lower chloride reduction and has negative impacts (see next slide)
- Low-cost scenarios with low chloride reduction outcome have negative impacts (see next slide)





Overall Evaluation

- Given that chloride concentration are favorable, even the highest-ranking scenario are not recommended at this point.
- Could be considered in Phase 3, if needed.



Scenario	Cost (M\$)	Cl-Removal	Other Impacts	Overall Rank
S1	7.1	High	Requires downstream pipe upgrade	Low
S2	1.1	High	Risk of backflow	Medium
S3	1.8	High	-	High
S4	0.3	Low	Low effectiveness due to constraint at North Pond	Low
S5a	0	Low	Increases risk of basement flooding, potential for bypass, ownership	Low
S5b	0.1	Low	Increases flow bypass, ownership	Low
S6	3	Low	Disturbance and loss of park and maintenance area, ownership	Low
S6b	1.7	Low	Significant soil removal, disturbance and loss of park area, ownership	Low
S6c	1.3	Low	Constraint by roadway and trails, limited space for pond cleaning, ownership	Low
S7a	0.3	Low	Ownership	Low
S7b	0.3	Low	Ownership	Low
S7c	1.9	Low	Significant soil removal, disturbance and loss of park area	Low
S8	6.1	High	Requires underground storage construction	Low



2025 Plan and Recommendations



2025/2026 Planned Activities

CORE MEASURES:

- Water quality monitoring and annual reporting to Subcommittee
- Geese and fish management
- Community engagement
- Shoreline restoration (Operations)
- Continue pond assumption process
- 5-year review in 2025

COMPLEMENTARY AND ALTERNATIVE MEASURES:

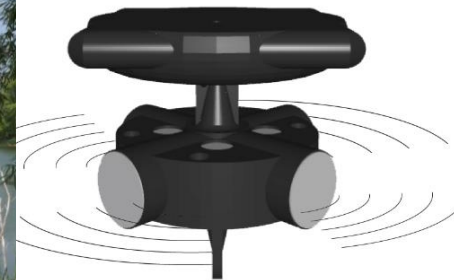
- Monitoring of planted Submerged Aquatic Vegetation
- Continue the Chloride Treatment Pilot Study

NEW COMPLEMENTARY MEASURES:

- ☐ Ultrasound Pilot Study
- ☐ Research by Trent University on Rare Earth Elements

Ultrasound Pilot Project

- City implemented an ultrasound pilot project to control algal growth in a stormwater pond in 2023 with promising results
- A low-cost and durable measure used to control algae growth in drinking water reservoirs
- It induces vibration and ruptures gas vacuoles (i.e., which control algae buoyancy), sinking algae to lower light levels of the pond deactivating algae growth
- Device installed in May 2025



Before Ultrasonic Treatment – September 2018

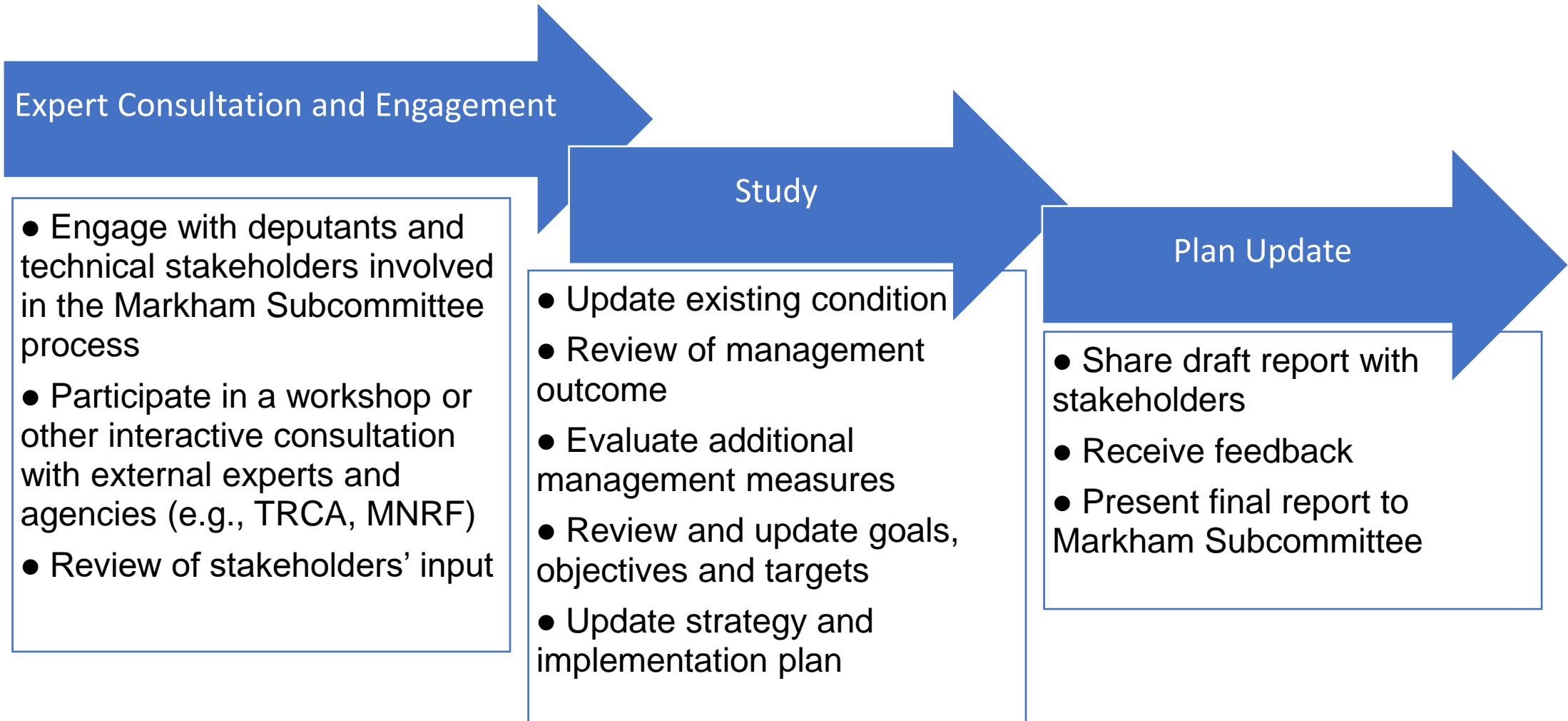


After Ultrasonic Treatment – September 2023





Plan Review Process



Program Outcomes

- Significant improvements in water quality and habitat have been realized, including aquatic vegetation and fish community
- Water quality now consistently meets expectations for shallow urban water bodies
- Innovative technologies and academic research are actively being evaluated
- Structural modification for chloride reduction is not required at this stage
- Phase 1 of the Long-Term Management Plan has successfully met all established goals and targets.





Recommendations

1. THAT the report entitled “Swan Lake- 2024 Water Quality Status and Updates” be received;
2. AND THAT Staff continue to implement the Long-term Management Plan for Swan Lake approved by Council in December 2021, including advancements previously made from Phases 2 and 3 of the Plan;
3. AND THAT Staff report back annually on water quality results and evaluation of adapted Core and Complementary measures for consideration in Phase 2 of the Plan through the Markham Sub-Committee with the participation of the Friends of Swan Lake Park;
4. AND THAT Staff consider findings and evaluations of chloride diversion options in Phase 3 of the Plan if required given future chloride levels in the Lake;
5. AND THAT the Plan review be initiated in 2025 with consideration for a workshop to review external feedback;
6. AND THAT Staff be authorized and directed to do all things necessary to give effect to this resolution.



Parks Operations

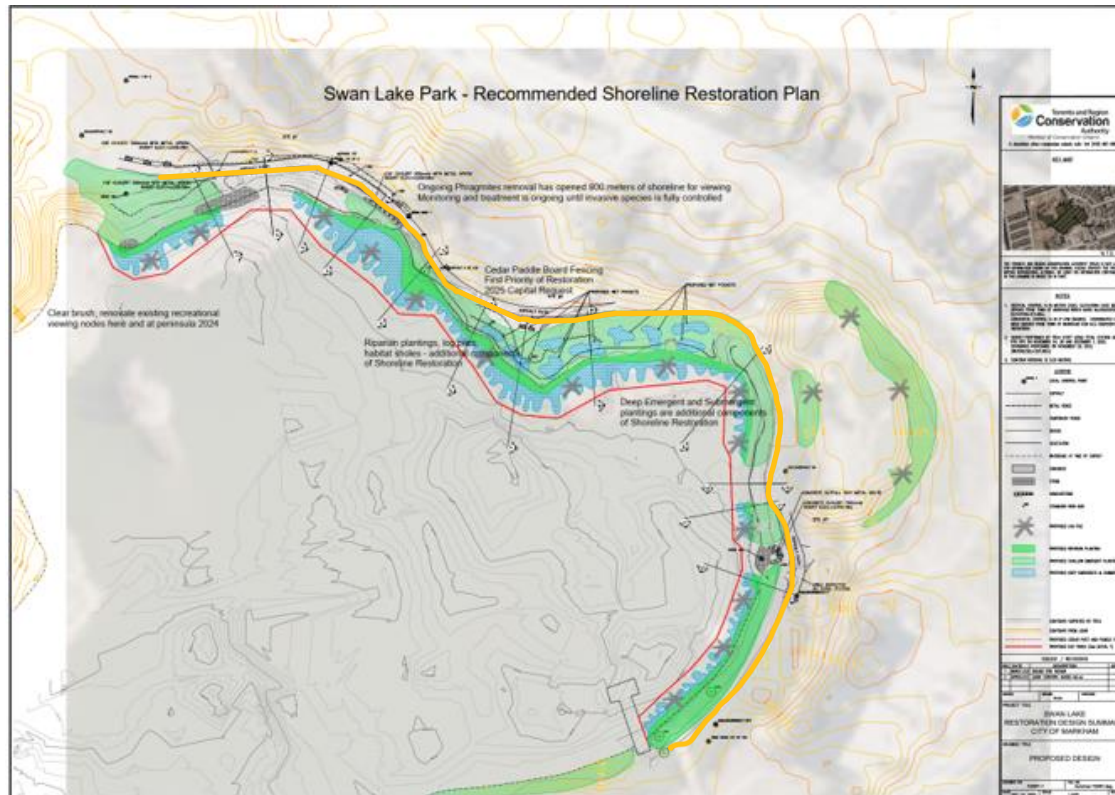


Community Engagement – Parks

- Following review of design concepts, staff did not recommend adopting recreational viewing nodes due to cost and with consideration for City-wide equity
- As noted by Regional Councillor Ho at the 2024 annual meeting, funding is limited for this type of construction
- Staff presented the approved shoreline restoration plan at the 2024 Public Meeting which did not include recreational viewing nodes
- Staff spoke to the approved plan and outlined the proposed scope of work for 2025 and beyond.
- There has been no change in the conditions affecting the recommended shoreline plan.

Parks Operations Next Steps

- The City is advancing the approved shoreline restoration plan which includes permanent barrier fence for waterfowl in the area where Phragmites was removed as shown by the thick yellow line of the restoration plan and image of typical fencing which will be closer to the waters edge when constructed.





Questions?



Swan Lake

Annual Meeting with Markham Subcommittee

June 18, 2025

